

1 **Effects of historic and projected climate change on the range**
2 **and impacts of an emerging wildlife disease**

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14

15 **Abstract**

16 The global trend of increasing environmental temperatures is often predicted to result in more
17 severe disease epidemics. However, unambiguous evidence that temperature is a driver of
18 epidemics is largely lacking, because it is demanding to demonstrate its role among the complex
19 interactions between hosts, pathogens and their shared environment. Here we apply a three-
20 pronged approach to understand the effects of temperature on ranavirus epidemics in UK
21 common frogs, combining *in vitro*, *in vivo* and field studies. Each approach suggests that higher
22 temperatures drive increasing severity of epidemics. In wild populations, ranavirosis incidents
23 were more frequent and more severe at higher temperatures, and their frequency increased
24 through a period of historic warming in the 1990s. Laboratory experiments using cell culture and
25 whole animal models showed that higher temperature increased ranavirus propagation, disease
26 incidence, and mortality rate. These results, combined with climate projections, predict severe
27 ranavirosis outbreaks will occur over wider areas and an extended season, possibly affecting
28 larval recruitment. Since ranaviruses affect a variety of ectothermic hosts (amphibians, reptiles
29 and fish), wider ecological damage could occur. Our three complementary lines of evidence
30 present a clear case for direct environmental modulation of these epidemics and suggest
31 management options to protect species from disease.

32 **Keywords:** ranavirus, temperature, virulence, climate change, emerging infectious disease, host-
33 pathogen interactions, common frog, *Rana temporaria*, amphibian population decline

34 **1 Introduction**

35 Interactions between hosts, pathogens and their shared environment shape infectious disease
36 outcomes, the timing of outbreaks, and the invasiveness of pathogens (Engering, Hogerwerf, &
37 Slingenbergh, 2013). Climatic conditions at local and landscape scales represent a critical
38 dimension of the host environment - affecting behavior (e.g. aggregation), stress and immunity -
39 but can also directly affect pathogen (and vector) growth and survival (Altizer, Ostfeld, Johnson,
40 Kutz, & Harvell, 2013; Engering et al., 2013; Epstein, 2001; Grassly & Fraser, 2006; Rohr et al.,
41 2011). As such, climatic conditions modulate host-pathogen interactions and operate on multiple
42 timescales: acting annually in driving seasonality (Altizer et al., 2006; Grassly & Fraser, 2006)
43 as well as over longer time periods in determining responses to climate change (Dobson et al.,
44 2015) and influencing the rate and pattern of invasions by emerging pathogens (e.g. Seimon et
45 al., 2007).

46 Despite this, invasion through pathogen range expansion and environmental change have
47 frequently been viewed as mutually exclusive factors in explaining disease emergence, as
48 indicated clearly in the framing and use of the “novel pathogen” (spread of an exotic pathogen
49 through naïve populations) and “endemic pathogen” (emergence due to perturbations of
50 interaction between hosts and native pathogens) hypotheses (Rachowicz et al., 2005). The
51 “global panzootic” in amphibians caused by the fungal pathogen, *Batrachochytrium*
52 *dendrobatidis*, serves as a prominent example of how climate change and pathogen range
53 expansion have been pitted against one another as alternative explanations of declines (Berger et
54 al., 1998; Lips, Diffendorfer, Mendelson III, & Sears, 2008; Pounds et al., 2006; Rohr, Raffel,
55 Romansic, McCallum, & Hudson, 2008). There is strong evidence supporting the rapid
56 international spread of a global panzootic lineage of *B. dendrobatidis* during the 20th century

57 (Farrer et al., 2011; O’Hanlon et al., 2018), but the proposed conflict between the two hypotheses
58 now appears to have been reconciled in a framework that incorporates both as key drivers of
59 emergence and outcomes, explaining observations of decline in regions where the impacts have
60 been greatest (Cohen, Civitello, Venesky, McMahon, & Rohr, 2018; Cohen et al., 2017; Raffel et
61 al., 2013; Rohr et al., 2008). Thus, it seems likely that the previous mindset of treating
62 environmental change and pathogen range expansion as conflicting has hampered understanding
63 of the patterns of emergence and the focusing of mitigation efforts.

64 Establishing a role for climate in disease emergence can be very challenging. Increasing
65 environmental temperature is a key component of climate change, which is cited as a driver of
66 infectious disease emergence and severity, but evidence for this is scarce and it is often difficult
67 to discriminate between the effect of temperature and other aspects of climate (Harvell et al.,
68 2002). The direct and indirect influences of temperature on host-pathogen interactions (Altizer et
69 al., 2006; Clare et al., 2016; Garner, Rowcliffe, & Fisher, 2011) and its nonlinear effects on
70 incidence and severity (Bosch, Carrascal, Durán, Walker, & Fisher, 2007; Raffel et al., 2013;
71 Walker et al., 2010) represent considerable challenges to a better understanding of disease
72 emergence (Rohr et al., 2011). Most research effort in this area has focused on human diseases
73 (Aguirre & Tabor, 2008), and vector-borne diseases (e.g. malaria, dengue, chikungunya) in
74 particular (Harvell et al., 2002; McMichael, Woodruff, & Hales, 2006).

75 In the current study, we investigate the effect of temperature on the interaction between
76 ranaviruses and their amphibian hosts, a host-pathogen system that offers the possibility of direct
77 experimental manipulation, and a well characterized recent history of pathogen invasion into the
78 UK (Price, Garner, Cunningham, Langton, & Nichols, 2016). Ranaviruses are large double-
79 stranded DNA viruses (family *Iridoviridae*) that can be highly pathogenic to ectothermic
80 vertebrates (Gray, Miller, & Hoverman, 2009; Price et al., 2014; Rosa et al., 2017). Ranavirus

81 infections of amphibians are notifiable to the World Organization for Animal Health due to their
82 potential to cause severe disease outbreaks as well as the risks of international spread through
83 trade (Schloegel, Daszak, Cunningham, Speare, & Hill, 2010; Schloegel et al., 2009). Ranavirus
84 growth and virulence can be affected by temperature (Ariel et al., 2009; Bayley, Hill, & Feist,
85 2013; Brand et al., 2016; Rojas, Richards, Jancovich, & Davidson, 2005) and environmental
86 temperature is considered to be one possible explanation for observations of seasonality in
87 outbreaks (Brunner, Storfer, Gray, & Hoverman, 2015). Indeed, incidents of ranavirosis in frogs
88 in the USA were recently shown to be uncoupled from a pulse in transmission or the density of
89 susceptible hosts, and instead were coincident with temperature increases and developmental
90 changes in frog larvae (Hall, Goldberg, Brunner, & Crespi, 2018).

91 Ranaviruses are distributed globally but outbreaks of disease are extremely patchy - a pattern
92 which is not yet understood. Some disease outbreaks have been shown to result from human
93 translocations of ranavirus (Jancovich et al., 2005; Picco & Collins, 2008; Price et al., 2016),
94 while other studies have found infections to be widespread at national scales without evidence
95 for disease, which may reflect an historic association (Warne, LaBumbard, LaGrange,
96 Vredenburg, & Catenazzi, 2016; Whitfield et al., 2013). The seasonal patterns and the
97 observations of a temperature effect in laboratory studies raise the possibility that environmental
98 conditions could drive invasion success and routes in cases where ranaviruses are undergoing
99 range expansion as well as climate change being a driver of disease emergence in regions where
100 the associations between viruses and hosts are historic and widespread.

101 In the UK, recurrent amphibian mass-mortality incidents caused by ranavirus have resulted in
102 severe population declines of the common frog (*Rana temporaria*) (Teacher, Cunningham, &
103 Garner, 2010). Genetic evidence supports multiple pathogen introductions into the UK whilst
104 spatiotemporal models suggest that ranavirus spread rapidly, facilitated by translocations of

105 unspecified infectious materials by people (Hyatt et al., 2000; Price et al., 2016). Disease
106 outbreaks are strongly seasonal, peaking in the summer months and appearing to mostly affect
107 adult animals, contrasting with other regions where larvae or metamorphic animals are the worst-
108 affected age-classes (Brunner et al., 2015). However, the detectability of the main UK host, the
109 common frog, is also strongly seasonal and there has been no previous attempts to explicitly
110 control for host population density, host activity or observer effort in examining the periodicity
111 of outbreaks (Cunningham, 2001; Cunningham et al., 1996; Teacher et al., 2010).

112 In this study, we investigated the role of temperature as a driver of disease outbreaks in UK
113 common frogs infected with ranaviruses in the frog virus 3 (FV3) lineage through a combination
114 of epidemiological modelling of a long-term study of disease in wild populations of UK common
115 frogs (Cunningham, 2001; Price et al., 2016), *in vitro* experiments involving manipulation of the
116 host-pathogen environment and similar *in vivo* experiments using natural hosts. Our aims were to
117 examine the role of temperature in shaping host-pathogen interactions to address whether it 1)
118 has been a factor explaining the pattern of invasion which can be used to predict future changes
119 in epidemiology under projections of climate change, and 2) can explain the observed seasonal
120 patterns in disease occurrence and the contrast in affected host age-class compared to other
121 temperate regions experiencing amphibian mortality incidents due to the same type of ranavirus.

122 **2 Materials and Methods**

123 **2.1 Temperature as a predictor of frog mortality and incident** 124 **severity**

125 **Temperature-dependence of ranavirus incidence:** We used data from the Frog Mortality
126 Project (FMP), a flagship citizen science project which collected information on amphibian
127 mortality incidents for over twenty-five years in order to study disease occurrence in wild
128 populations (Price et al., 2016). The project provided a structured reporting system and guidance
129 to volunteer reporters on how to report amphibian mortality incidents, including a range of
130 variables describing the environment and its management. The FMP dataset has been reliably
131 filtered for incidents of ranavirosis previously (North, Hodgson, Price, & Griffiths, 2015; Price et
132 al., 2016; Teacher et al., 2010) based on detailed post-mortem examinations of dead and sick
133 amphibians reported from multiple sites over multiple years (Cunningham, 2001; Cunningham
134 et al., 1996; Price et al., 2017). In this study, the same criteria as Cunningham (2001) and Price et
135 al. (2016) (the presence of indicative signs of disease [‘ulceration’, ‘red spots on the body’, and
136 ‘limb necrosis/loss of digits’; see Supporting information Appendix S1 and Figure S1] and a
137 minimum of five dead animals) were used to create a binary variable describing the ranavirus
138 status of each incident. To date, no diagnosis other than ranavirosis has been made from
139 pathological examinations of common frogs from such mortality incidents in the UK
140 (Cunningham, 2001; Cunningham et al., 1996; Price et al., 2017). The detectability of
141 amphibians varies seasonally, whilst reporting to the FMP was subject to possible temporal and
142 spatial variation in effort (Price et al., 2016). We controlled for these potential biases indirectly
143 through the inclusion of mortality incidents caused by factors other than ranavirosis as
144 previously (Price et al., 2016). Data on the timing of the onset of mortality (available at the

145 resolution of month only) and the incident location were used to download the monthly average
146 of the daily maximum temperature (T_{MAX30}) for the month of onset of disease for each incident
147 from the Met Office (Met Office, 2017b) (further details in Supporting information Appendix
148 S2).

149 Factors affecting ranavirus incidence were investigated using a standard logistic regression
150 model incorporating variables describing the environment (temperature), pond (volume, shading,
151 and presence of marginal and floating vegetation [e.g. see Raffel, Michel, Sites, & Rohr, 2010]),
152 other aquatic vertebrate species present in addition to common frogs (toads, newts, fish) and
153 geographic region (government office region) as predictors of ranavirus status (Model 1) fitted
154 with the R function *glm2* (Marschner, 2011; R Core Team, 2017). To explore the relationship
155 between temperature and the probability of a ranavirus-positive observation further, it was
156 modelled as a sigmoid (logistic) transition between an upper and lower mean frequency (Model
157 2). The four parameters of the curve (upper and lower limits, the location and the slope of the
158 transition) were fitted using the *mle2* function in the R package, *bbmle* (Bolker & R Core Team,
159 2017; R Core Team, 2017). Starting values for the *mle2* algorithm were obtained for slope and
160 location (intercept), using the *glm2* function with a bespoke link function for binomial data (Data
161 S2). A confidence interval around the fitted line was generated using the delta method to
162 compute variance of a function with the *emdbook* package (Bolker, 2008, 2016). Model 2 was
163 compared using Akaike's Information Criterion (AIC) to a standard logistic curve (a simplified
164 version of Model 1 with only temperature terms retained as predictors).

165 **Effect of increased temperature on the severity of disease incidents:** In order to analyze the
166 effect of temperature on the severity of mortality incidents, the number of dead frogs and the
167 number of surviving frogs (estimated by reporters) at each reported incident were combined as
168 estimates of the proportion of the population that died (each pond/frog population was included

169 in the analysis only once) and used as the response term in a generalized linear model using the
170 quasibinomial family to account for overdispersion in the data (Crawley, 2013). Temperature
171 (T_{MAX30}), the predicted ranavirus status, their interaction, and additional variables which
172 described the pond environment (presence of other species and physical characteristics of ponds)
173 were used as predictors of severity. Since the total number of dead animals at each incident was
174 used to generate the proportion of the population that died when filtering the dataset for
175 ranavirus-consistent incidents, the criterion requiring mortality incidents to include five or
176 more dead frogs was dropped but the signs of disease data were used more stringently instead
177 (observations of two of the three indicative signs of disease used above were now required).

178 **Temperature preceding ranavirus outbreaks with precise timestamps:** We previously
179 screened a UK amphibian and reptile tissue archive for ranavirus and returned a database of
180 mortality incidents for which the presence of ranavirus was established using molecular
181 diagnostic methods (Price et al., 2017). After filtering for incidents with a precise georeference
182 (postcodes or grid references) and timestamp (date found, submitted, or examined if a post-
183 mortem examination was conducted on receipt, which were each assumed to approximate closely
184 to the day of death due to the rapid decomposition of amphibian carcasses), this dataset
185 contained a total of 197 incidents, for which ranavirus had been detected from 31. All incidents
186 were overlaid on the UK grid of 5 x 5 km squares and the maximum daily temperatures for the
187 date matching the timestamp and the 50 days prior to the death(s) were extracted from plain text
188 data files in the Met Office UK Climate Projections 2009 dataset (Met Office, 2017a), which
189 were downloaded using the R package *Rcurl* (Temple Lang and CRAN team 2018; Price, 2018).
190 Ranavirus status was used as the response variable in a series of logistic regression analyses
191 (generalized linear models using the binomial family and the logit link function) with the average
192 maximum daily temperature in the week preceding the mortality incident (T_{MAX7}) or the number

193 of consecutive days in the previous seven where temperature exceeded 16°C ($\text{DAYS}_{T>16}$) as
194 predictors. These models were also run with region (Government Office Region) or latitude as an
195 additional predictor to further control for any effect of spatial variation in temperature. As above,
196 seasonal variation in the detectability of amphibians was controlled for by inclusion of mortality
197 incidents caused by factors other than ranavirosis.

198 **Effects of historic warming and seasonality:** To check whether prior warming (over the time
199 course of the dataset; 1991-2010) had altered the rate of ranavirus incidents and to assess
200 seasonality in the data, we first decomposed the annual signal and trend across years in the time
201 series of temperatures and rates of ranavirosis incidents. The mean of the average daily
202 maximum temperature during the month of onset of mortality incidents from all reports in the
203 FMP dataset (1991-2010) was calculated for each month with reports, as well as the numbers of
204 reports that were consistent or otherwise with ranavirosis. Generalized Additive Mixed Models
205 [GAMM; R function *gamm*; package *mgcv* (Wood, 2003, 2004, 2017)] were used to fit smooth
206 splines to both the within-year (seasonal; cyclic cubic regression spline) and across-year (cubic
207 regression spline) patterns and autocorrelation structures (of order one) were used to model
208 residual correlation within years. The number of ranavirosis incidents as a proportion of total
209 reports was then modeled as a function of the seasonal trend (smoothed with a cubic regression
210 spline as above), the trend in temperature across years (predicted using the GAMM above and
211 smoothed with a cubic regression spline) and time using a Generalized Additive Model (GAM)
212 and the binomial family with logit link function (function *gam*; package *mgcv*).

213 Model assumptions were verified by plotting residuals against fitted values and against each
214 covariate in the model (function *gam.check*; package *mgcv*). Autocorrelation among residuals
215 was assessed using the *acf* function in the *stats* R package (Venables & Ripley, 2002). Model
216 predictions and residuals were extracted and visualized using functions in the R packages, *visreg*

217 and *ggplot2* (Breheny & Burchett, 2013; Wickham, 2016). The predictive power of the model
218 incorporating the across-year temperature trend was then assessed by comparison to a model
219 containing the temporal trend in the rate of ranavirus incidents by dividing the dataset into two
220 training and test sets (taking half and three-quarters of the data for training respectively). GAMs
221 incorporating the smoothed seasonal trend and either a smoothed (across-year) time trend *or* the
222 smoothed temperature trend were fitted using the training datasets and compared in terms of their
223 ability to predict patterns in the test datasets.

224 **2.2 Virus growth *in vitro***

225 To investigate the effect of temperature on viral growth, two UK isolates of FV3 [RUK11,
226 isolated from the kidney of a diseased common frog that died with systemic hemorrhages in
227 Middlesex in 1992, and RUK13, isolated from the skin of a diseased common frog that died with
228 skin ulceration in Suffolk in 1995 (Cunningham, 2001)] were grown at a range of temperatures
229 in two cell lines (epithelioma papulosum cyprini [EPC, derived from the fathead minnow fish,
230 *Pimephales promelas* (Winton et al., 2010); ECACC 93120820] and the iguana heart reptile line
231 [IgH2; ECACC 90030804]). Cells were grown on 96 well plates until more than 90% confluent
232 and then inoculated with virus in a ten-fold dilution series ranging from an estimated multiplicity
233 of infection of approximately 2×10^{-6} to 2×10^3 (five wells per dilution with an additional one
234 well per dilution receiving a sham exposure of cell culture media only as a negative control).
235 Titers of viral isolate stocks were equalized by reference to qPCR scores [following Leung et al.
236 (2017)]. Plates were incubated at six temperatures (10, 14, 18, 22, 26, and 30°C) and monitored
237 daily for cytopathic effect (plaques in the cell layer). After six days, plates were scored for viral
238 growth by counting the number of replicates at each dilution where cytopathic effect was evident
239 and calculating the Tissue Culture 50% Infective Dose (TCID₅₀) using the method of Reed &

240 Muench (1938). The effects of temperature and cell line on viral growth (the mean titers of the
241 two isolates of each type) were analyzed using a linear model in R.

242 **2.3 *In vivo* assessment of effect of temperature on virulence**

243 To investigate whether an effect of temperature on *in vitro* viral growth was reflected in altered
244 virulence *in vivo*, 60 overwintered common frog metamorphs (*R. temporaria*) were randomly
245 allocated to one of six treatments (10 animals per treatment): three exposure treatments (sham,
246 RUK11, RUK13) crossed with two temperatures (20°C [“low”] and 27°C [“high”]). Temperature
247 was maintained by placing five individually-housed frogs selected at random from each of the
248 three exposure treatments into one of four climate-controlled chambers constructed from
249 polystyrene boxes - two held at 20°C and two held at 27°C (Figure S2; see Supporting
250 information Appendix S3 for a comprehensive description of the set up). RUK13 was used at a
251 titer of 1.58×10^5 TCID₅₀ mL⁻¹ and RUK11 at a titer of 1.58×10^7 TCID₅₀ mL⁻¹ (see
252 Supporting information Appendix S3 for details of preparation of inocula and exposure
253 methods).

254 Individuals reaching endpoints (either gross signs of hemorrhaging or ulceration; Figure S1) and
255 all surviving individuals at the end of the experiment (day eight post-exposure) were euthanized
256 following a schedule 1 method for amphibians. The number of hours post-exposure that animals
257 were found dead or at endpoints was recorded for survival analysis. Survival data were analyzed
258 by fitting a mixed-effects Cox proportional hazards regression model in R using the *coxme*
259 package (Therneau, 2018) with exposure and temperature as fixed effects and the climate-
260 controlled chamber as a random effect to account for pseudo-replication due to placement of
261 individual experimental units inside the four climate-controlled chambers.

262 A second *in vivo* experiment was performed in the same host species, using one ranavirus isolate
263 (RUK13) and a single temperature (20°C), but two exposure doses (“low” and “high”; detailed
264 methods in Supporting Information Appendix S1). The effects of temperature and dose on
265 disease progression and survival were compared.

266 **2.4 Effect of projected climate change on timing of ranavirus**

267 **outbreaks**

268 Baseline temperature data were generated by calculating the mean daily maximum temperature
269 for each calendar month for the period 1991-2010 (the same period covered by our dataset of
270 reported frog mortality incidents) for each 5 x 5 km grid square used by the Met Office UK
271 Climate Projections 2009 (UKCP09) project (Met Office, 2017b). UKCP09 integrated a range of
272 climate models to predict future climate under the Intergovernmental Panel on Climate Change
273 (IPCC) B1 (“low”), A1B (“medium”) and A1FI (“high”) future emissions scenarios (IPCC,
274 2000; Jenkins et al., 2009). Model outputs include probabilities that climatic variables will
275 exceed user-selected values at several future timepoints. We used the threshold temperature from
276 ranavirus occurrence “Model 2” (see section “Temperature-dependence of ranavirus incidence”),
277 at which the risk of ranavirus outbreaks rose sharply, in order to generate probabilities of the
278 mean daily maximum temperature exceeding 16°C for calendar months during the period 2070-
279 2099 for each UK 25 x 25km grid square and under each future emissions scenario. These
280 probabilities were visualized by coloring grid squares on maps based on a probability threshold
281 of 0.5.

282 **3 Results**

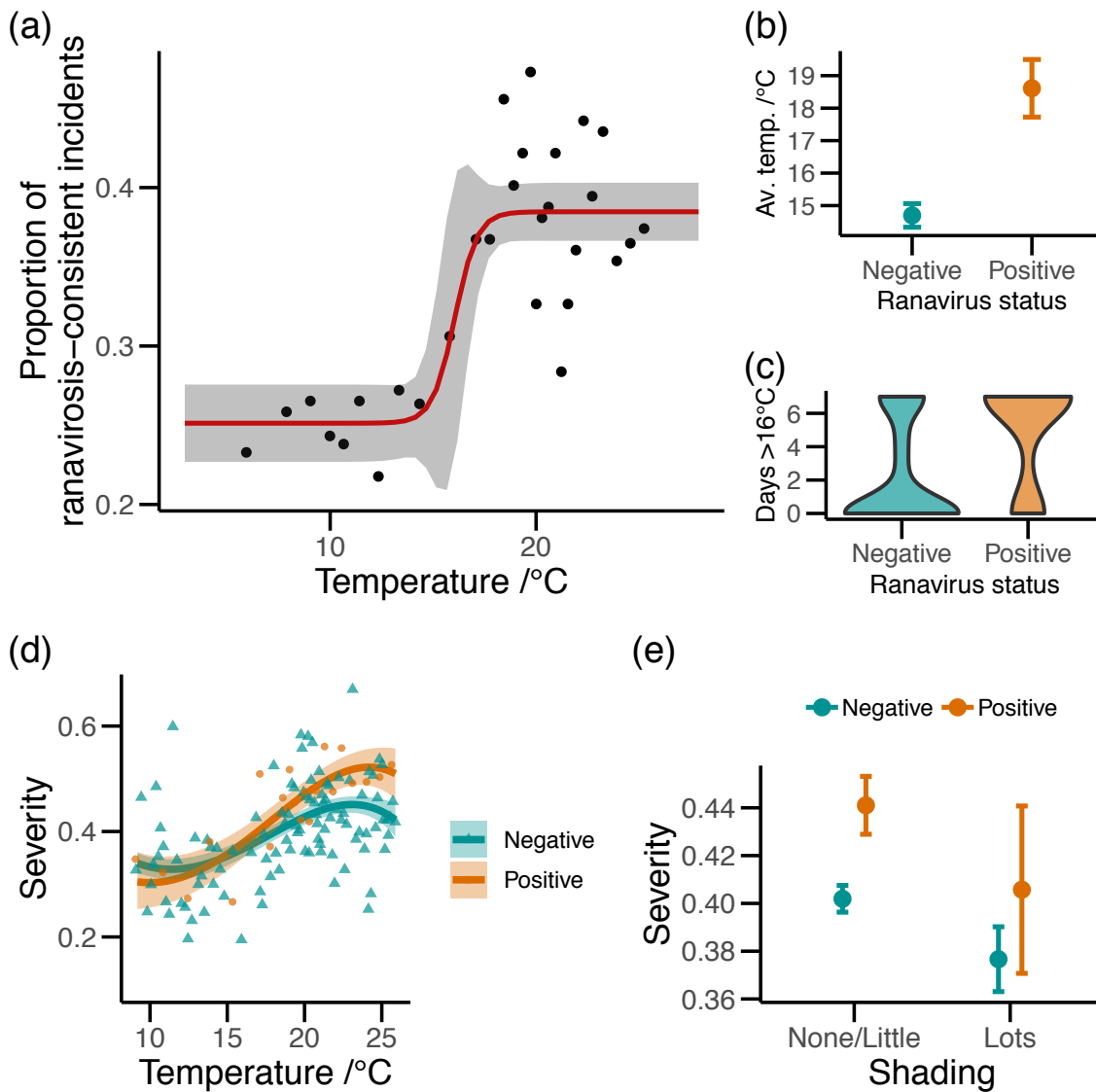
283 **3.1 Effect of temperature on disease occurrence and severity in the** 284 **wild**

285 The finalized FMP dataset used in this study contained 4385 unique records, of which 1497
286 were classed as ranavirosis-consistent. The logistic model (Model 1) revealed a highly
287 significant, non-linear effect of temperature on the proportion of ranavirosis-consistent incidents
288 observed (Table S1). The minimal adequate model also retained newts, fish and shading: the
289 presence of either type of animal in ponds increased the proportion of ranavirosis-consistent
290 incidents observed whilst shading reduced this proportion (Table S1). To explore the relationship
291 with temperature further, a model with a transition between an upper and lower frequency was
292 fitted (Model 2), which improved the fit to the data compared to a simplified version of Model 1
293 comprising only the terms describing the non-linear relationship with temperature (AIC scores
294 were 5565 and 5576 respectively). Model 2 shows a step-change: below approximately 16°C,
295 25.1% of incidents were ranavirosis-consistent, rising to 38.5% after the temperature threshold
296 was crossed (Figure 1a). The difference between incidents that were ranavirosis-consistent and
297 the remainder ('non-ranavirus') is also apparent in the distribution of temperature records: the
298 non-ranavirus category being strongly bimodal with peaks at both low and high temperature,
299 whereas most of the ranavirosis-consistent incidents were reported at higher temperatures with
300 few outliers at much lower temperatures (Supporting information Figure S3).

301 Temperature was again a highly significant predictor of ranavirus status when the records with
302 precise timestamps and confirmed ranavirus-positive status were analyzed. This more precise
303 information about timing enabled a fine-scale examination of the effect of temperature in the

304 days preceding incidents. The average temperature in the seven days preceding incidents (T_{MAX7})
305 was a significant predictor of ranavirus status (odds ratio = 1.20, 95% confidence interval = 1.09
306 – 1.31; Figure 1b). The temperature threshold where the proportion of ranavirus incidents
307 increased sharply in the analysis of the full FMP dataset (Model 2) was approximately 16°C. A
308 second model - with the number of consecutive days where the daily maximum temperature in
309 the week preceding incidents exceeded 16°C ($DAYS_{T>16}$) as a predictor - also indicated that
310 warmer temperatures were a good predictor of ranavirus status with each additional warm day
311 raising the odds that incidents were caused by ranavirus (odds ratio = 1.33, 95% confidence
312 interval = 1.16 – 1.51; Figure 1c). The 16°C threshold model had a slightly lower AIC score than
313 the model using the average temperature as a predictor (155 compared to 158).

314



315

316 **Figure 1. Warm temperatures increased the frequency and severity of incidents of**
 317 **ranavirosis involving wild common frog populations of the United Kingdom.** (a) The effect
 318 of temperature on ranavirosis incident rate (the proportion of citizen science reports of frog
 319 mortality that were classified as ranavirosis-consistent). The line represents the fitted maximum
 320 likelihood model of a logistic transition between a lower and upper frequency. The shaded area
 321 around the line represents the 95% confidence interval, calculated using the delta method. Points

322 represent the observed data, grouped in 30 windows each containing 146-148 individual records.
323 (b-c) Temperature in the week preceding frog mortality incidents confirmed by molecular
324 methods predicted ranavirus status (“Positive” or “Negative”). (b) T_{MAX7} by ranavirus status. (c)
325 $DAYS_{T>16}$ by ranavirus status. (d) The severity of frog mortality incidents (estimated proportion
326 of population that died) was consistently greater at higher temperatures, particularly in the case
327 of ranavirosis-consistent incidents (orange). The plot shows fitted lines (and 95% confidence
328 intervals) from a generalized linear model (quasibinomial regression) of severity as a function of
329 ranavirus status and temperature (T_{MAX30} for the month of onset of mortality incidents, including
330 quadratic terms). Points are a summary of raw data after binning individual data points (mode of
331 twenty data points per bin) and calculating the mean temperature and overall proportion of dead
332 frogs within each bin. (e) Large amounts of shading around ponds reduced the severity of
333 ranavirosis-consistent mortality incidents (orange) compared to other incidents (green). In panels
334 (b) and (e), points represent means and error bars represent ± 1 standard error of mean.

335

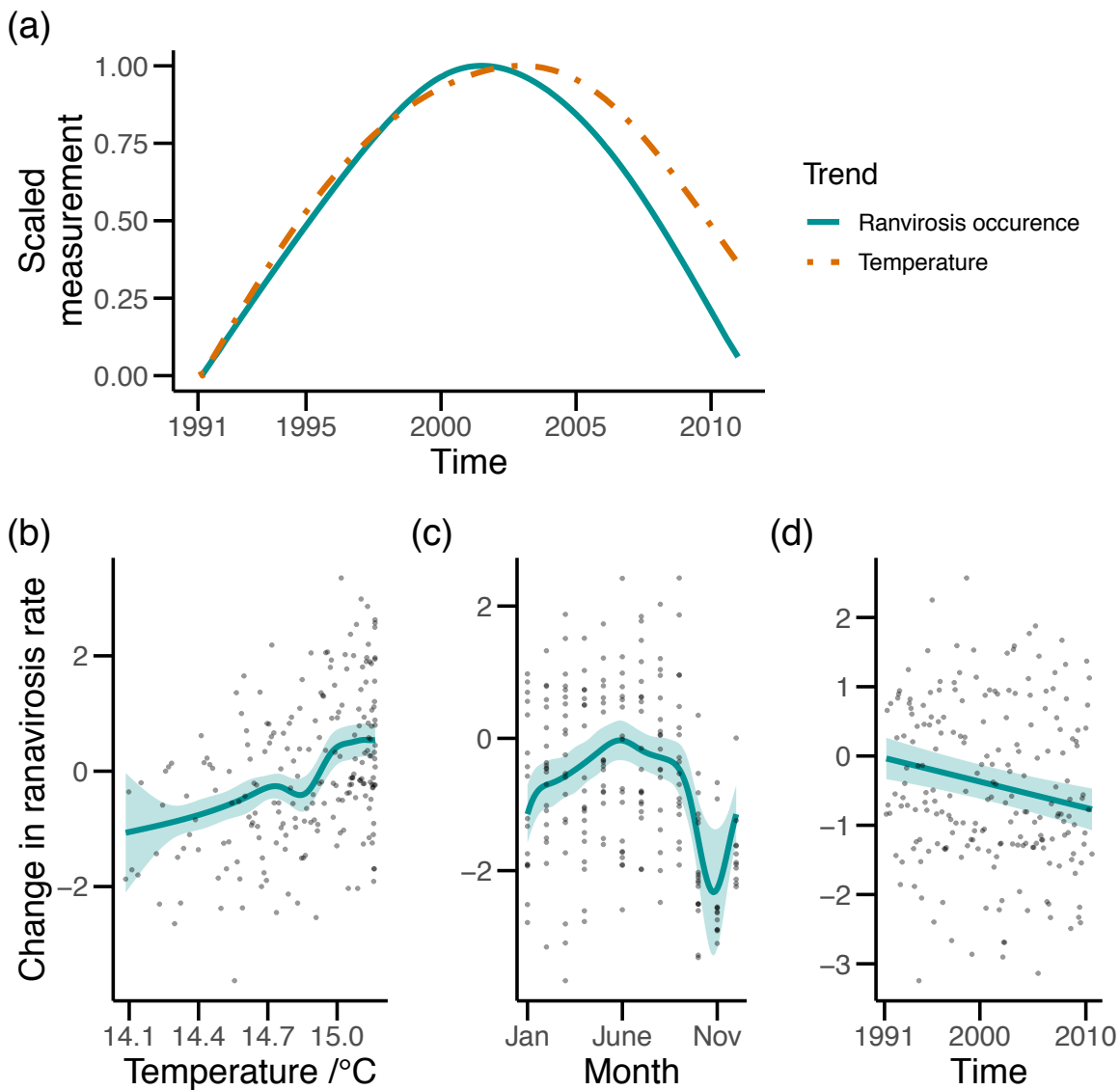
336 The FMP database contains data on the severity of outbreaks (the estimated proportion of the
337 frog population that died) for the years 1991-2000. After removing records with missing values,
338 we produced a dataset for investigating severity that contained 2667 records, of which 427
339 incidents were classified as ranavirosis-consistent. The effects of temperature and covariates
340 previously identified as having an influence on the occurrence or severity of ranavirosis in UK
341 common frogs (North et al., 2015) were explored using a logistic model. The ranavirus status and
342 temperature (T_{MAX30} at month of mortality onset) plus their interaction, as well as the log-
343 transformed pond volume, shading around ponds, the amount of both the marginal and floating
344 vegetation (“none/little” or “lots”), the presence of toads, the presence of newts, the presence of
345 fish, and the region were all included as predictors in the model. After model simplification, the

346 minimal adequate model retained temperature, ranavirus status and their interaction but there
347 were also significant effects of the presence of toads, the presence of fish, shading, pond volume
348 (non-linear), marginal vegetation and region (Table S2). Temperature had a significant, positive
349 effect on incident severity, each 1°C increase in temperature leading to a 3.1% increase in the
350 proportion of the population that died ($p = 7.56 \times 10^{-12}$). There was a significant interaction with
351 ranavirus status ($p = 7 \times 10^{-4}$): at low temperatures, the severity of ranavirosis-consistent
352 incidents was lower than other types of incident but at higher temperatures ranavirosis-consistent
353 incidents were more severe (Figure 1d; Supporting information Figure S4). Toads reduced the
354 severity of mortality incidents whilst the presence of fish increased severity (Supporting
355 information Figure S5) as found previously (North et al., 2015). Shading and marginal vegetation
356 decreased the severity of incidents. Notwithstanding, and irrespective of which covariate was
357 considered, the effect of increasing temperature increased the severity of disease. This is perhaps
358 best illustrated by the effects of pond shading, where increasing the amount of shading (and,
359 presumably, decreasing the maximum temperatures that frogs would have been exposed to) was
360 associated with reduced severity of ranavirosis and a reduction in the disparity in the severity of
361 incidents between ranavirosis-consistent and non-ranavirus incidents (Figure 1e).

362 **3.2 Historic climate change and seasonality**

363 Temperatures in our study region varied across years, showing a warming trend that peaked in
364 2002 before cooling up to 2010, as well as showing marked seasonality (GAM component of
365 GAMM: $R^2 = 0.93$; Table S3). The proportion of ranavirosis incidents followed a remarkably
366 similar pattern, in terms of the strong seasonality and the trend across years (GAM component of
367 GAMM: $R^2 = 0.36$; Table S4; Figure 2a). The across-year temperature trend was a highly
368 significant predictor of the rate of ranavirus incidents (GAM: effective degrees of freedom (edf)

369 of smoothed term = 6.65, $p = 4.61 \times 10^{-26}$), explaining the pattern of increasing rates which
370 peaked in 2001, with the full model explaining approximately half the deviance ($R^2 = 0.48$;
371 Figure 2; Table S5). Model validation indicated conformity to assumptions. The temperature
372 model, when trained on subsets of the data, predicted trends in test datasets effectively in
373 contrast to models incorporating only the across-year trend in ranavirosis rates, which showed no
374 predictive power (Supporting information Figure S6). Temperature showed a strong seasonal
375 pattern which was reflected in the pattern of ranavirosis incidents where seasonality was also
376 marked (GAM: edf = 7.65, $p = 2.45 \times 10^{-13}$; Figure 2c). Overall, after accounting for the effect
377 of temperature, there was a small but significant decrease in the proportion of ranavirosis
378 incidents between 1991 and 2010 (GAM: coefficient = -0.00315 [unit of time was months], $p =$
379 2.31×10^{-5} ; Figure 2d).



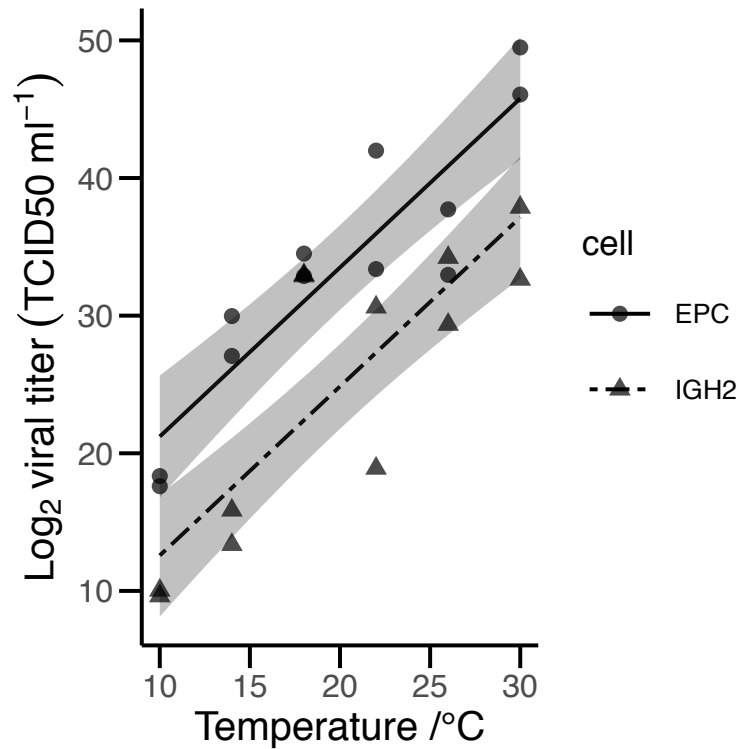
380

381 **Figure 2. Effect of historic climate and seasonality on rate of ranavirosis incidents.** (a)
 382 comparison of smoothed trends in temperature and rates of ranavirosis incidents (on standardized
 383 scale [0-1]) over period of dataset (1991-2010). (b-d) Effect of predictors of the rate of
 384 ranavirosis incidents (from generalized additive model) against residuals: Smoothed change in
 385 rate of ranavirus incidents with temperature (b), smoothed seasonal change in rate of ranavirosis
 386 incidents (c), and change in rate of ranavirosis incidents over time (d). Shaded areas represent
 387 95% confidence intervals.

388 3.3 *In vitro* assessment of viral growth rates and *in vivo* tests of 389 virulence

390 We examined *in vitro* viral growth using two UK isolates of FV3 (RUK11 and RUK13; see
391 methods for detailed descriptions of isolates) and two cell lines. Each isolate was incubated at a
392 range of temperatures up to 30°C with each cell line, but regardless of cell line or isolate,
393 increasing temperature resulted in exponentially increased rates of plaque formation (Figure 3).
394 A linear model of log viral titers against temperature, cell line and their interaction revealed
395 significant effects of temperature (coefficient = 1.23, $p < 1e^{-30}$) and host cell line (IgH2
396 compared to EPC, coef = -8.64, $p < 1e^{-30}$) but no interaction (analysis of variance, comparing
397 model with interaction term to a model with main effects only: $F_{df=1} = 0.0047$, $p = 0.95$).

398



399

400 **Figure 3. Effect of environmental temperature on growth of UK *Frog virus 3* (FV3) *in vitro*.**

401 Observed data (points) and predictions from linear model (lines with 95% confidence interval
 402 shaded) of FV3 growth at a range of environmental temperatures in fish (EPC; solid line) and
 403 reptile (IgH2; dashed line) cells. Growth was measured using the TCID50 method and is shown
 404 on a log scale. An increase in temperature of 1°C results in more than a doubling of viral growth
 405 (2.34 times).

406

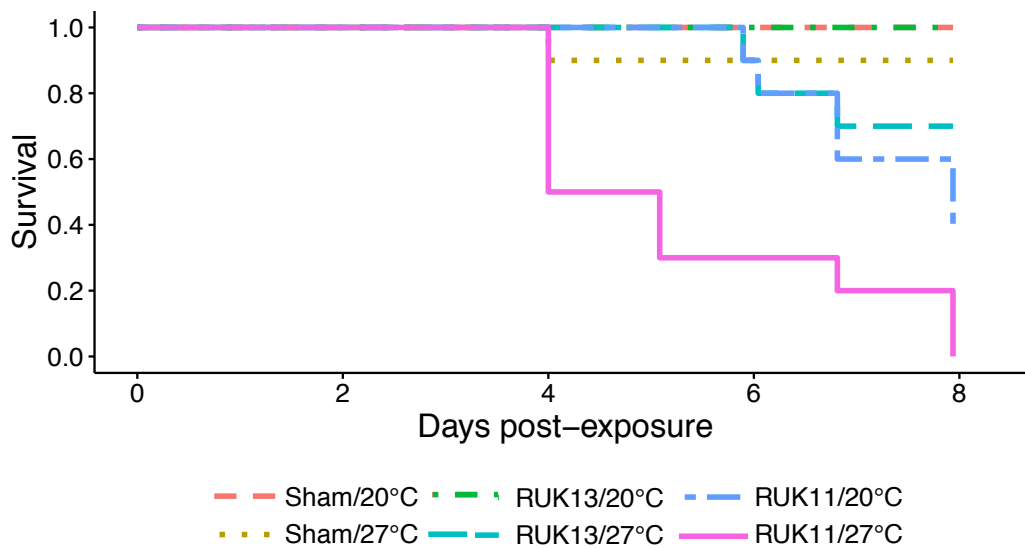
407 The effect of temperature on the response of common frogs to viral exposure was assessed in
 408 order to validate results from cell culture models *in vivo*. Temperature was a highly significant
 409 predictor of survival: 20 of 60 animals died or were euthanized on reaching humane endpoints,
 410 of which 14 were from high temperature treatments and six were from low temperature

411 treatments (Figure 4). Overall there was a 5.33 times higher risk of death in the high temperature
412 treatments ($p = 0.005$; Table 1). Titers of viral inoculates were not equalized between isolates.
413 All individuals exposed to RUK11 (at a high dose) and maintained at high temperature died or
414 reached endpoint by the eighth day post-exposure compared to six of ten individuals maintained
415 at low temperature (Figure 4). Of the animals exposed to RUK13 (at a relatively low dose
416 compared to RUK11), three individuals died or reached endpoint in the high temperature
417 treatment compared with none at the low temperature. There was also a significant effect of
418 exposure treatment: the expected hazard of animals exposed to RUK11 was 41.6 times higher
419 than animals receiving a sham exposure ($p = 0.0004$). These results are largely in line with a
420 study examining survival of common frog tadpoles exposed to a North American isolate of FV3,
421 which showed that mortality was increased at 20°C compared to 15°C (Bayley et al., 2013).

422 The second *in vivo* experiment examining the effect of dose on disease outcome and progression
423 in juvenile common frogs (Supporting Information Appendix S1) complements the findings of
424 the *in vivo* temperature experiment. The dose experiment suggests that a viral load threshold
425 exists which must be crossed before gross signs of disease develop. We found the outcome and
426 presentation of disease as well as the viral quantity in tissues at death to be largely independent
427 of dose: all animals exposed to either low or high viral doses died, presented with the same set of
428 signs (Figure S1 & S7), and had similar quantities of virus in their tissues at death (Figure S8a).
429 However, the onset and progression of disease was delayed at low dose (the development of
430 disease and death both occurred significantly later; Figure S7) reflecting the lower viral loads of
431 individuals (measured by swabbing animals after infection; Figure S8b) in this treatment. Also,
432 viral loads of dead individuals were greater than those of live individuals (both when repeated-
433 measures from the same individuals were compared [Figure S8c] and when individuals that were
434 euthanized part-way through the experiment were compared to those that died [Figure S8d]).

435 These results all suggest that elevated viral loads lead to the onset of disease and that the viral
 436 capacity to cross a threshold concentration is a more important determinant of whether disease
 437 develops than the initial dose. It seems likely that higher temperature and higher initial dose each
 438 serve as ways to reach this putative threshold for disease sooner - either through more rapid viral
 439 growth or a greater initial intensity of infection respectively - and explain the delays and/or
 440 reductions in observations of severe outcomes in the other respective treatments (low
 441 temperature or low dose).

442



443

444 **Figure 4. Effect of temperature and ranavirus exposure on survival of common frogs.**

445 Kaplan-Meier survival plot showing the proportions of surviving animals through time plotted
 446 for each of six treatments (n=10 frogs per treatment); three exposure treatments (Sham, RUK11,
 447 RUK13) at each of two temperatures (20°C [“low”] or 27°C [“high”]).

448

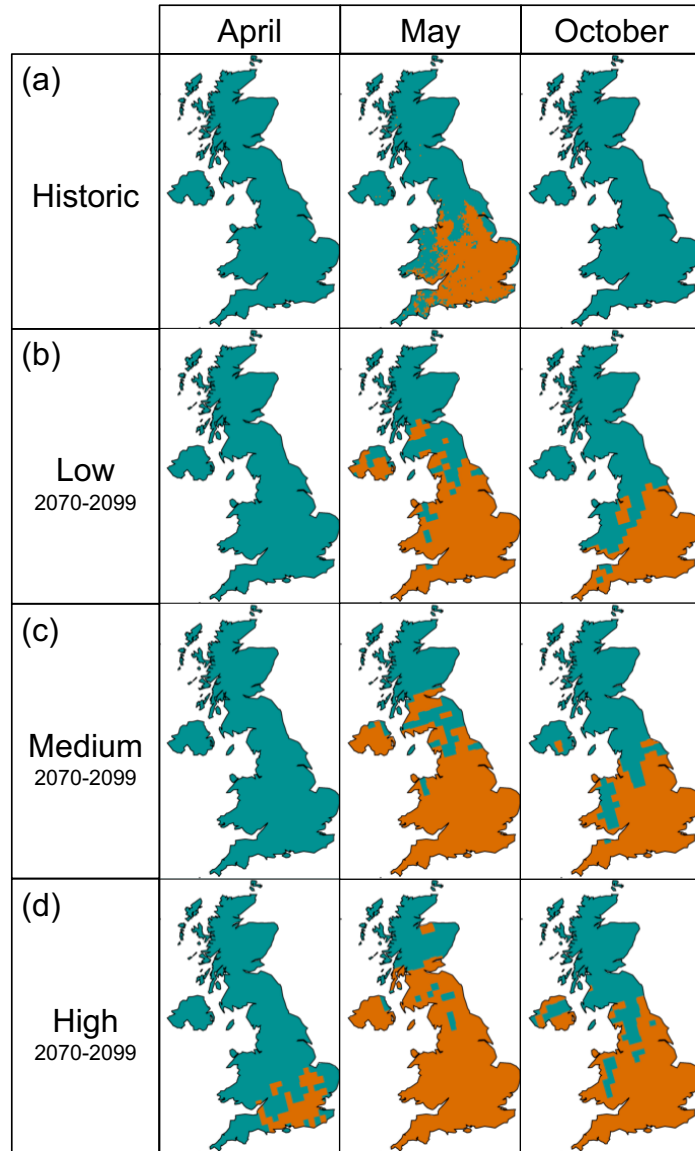
449 **Table 1. Effect of temperature and ranavirus exposure treatments on the survival of**
 450 **common frogs.** Coefficients and hazard ratios from a mixed effects Cox Proportional Hazards
 451 model of survival in response to exposure (“Sham” as the reference level) and temperature
 452 (“Low” as the reference level) treatments.

	Coefficient	Hazard ratio (e^{coef})	Standard Error	z-value	p-value
RUK13 exposure	1.17	3.23	1.16	1.01	0.310
RUK11 exposure	3.73	41.6	1.05	3.56	<0.001
“High” temperature (27°C)	1.67	5.33	0.59	2.84	0.005

453

454 3.4 Impact of future climate on timing of outbreaks

455 The UK climate is expected to warm considerably over the remainder of the century (Chen &
 456 Tung, 2018; Jenkins et al., 2009). A warmer climate would expand the geographic area where
 457 environmental conditions are likely to be suitable for severe incidents of ranavirosis (T_{MAX30}
 458 exceeding 16°C; Model 2). For example, using 16°C as the point value where the risk of
 459 ranavirosis outbreaks increases, the suitable geographic area for ranavirosis occurrence in May is
 460 projected to increase by 134% by 2070 under a high emissions scenario compared to the historic
 461 baseline and by 84% under a low emissions scenario (Figure 5). The projected changes in UK
 462 temperatures are also likely to extend the duration of the “disease season”, creating favorable
 463 conditions for disease during the spring and autumn as well as in the summer. Temperatures are
 464 likely to become more suitable for severe outbreaks in a large part of England in April under a
 465 high emissions scenario and in October under any of the range of IPCC scenarios investigated.
 466 These are months that we expect to have experienced limited incidence and severity of
 467 ranavirosis previously, but temperatures are projected to change to such a degree that this
 468 limitation will be removed for large areas of the UK (Figure 5).



469

470 **Figure 5. Projected shifts in geographic extent and the temporal window of ambient**
 471 **temperatures high enough to increase the risk and severity of ranavirosis incidents in the**
 472 **UK under different emissions scenarios.** (a) Past climate - orange regions are those where the
 473 average monthly temperatures (daily maximum temperature) in the UK (5 x 5 km grid squares)
 474 exceeded 16°C for the period 1991-2010. (b-d) Future climate - orange regions are those where
 475 the projected average monthly daily maximum temperatures in the UK (25 x 25 km grid squares)

476 have a greater than 50% probability of exceeding 16°C under a range of future emissions
477 scenarios for the period 2070-2099: (b) Intergovernmental Panel on Climate Change (IPCC)
478 scenario B1 (low), (c) scenario A1B (medium) and (d) scenario A1FI (high).

479

480 **4 Discussion**

481 By combining the results of laboratory experiments with the analysis of two types of
482 epidemiological dataset relating to disease outbreaks in wild amphibian populations, we have
483 revealed a pattern of remarkably consistent evidence supporting a strong, positive effect of
484 temperature on the occurrence and severity of ranavirosis in UK common frogs. Temperature
485 predicts both the incidence and severity of disease outbreaks through higher temperatures
486 increasing viral growth which, in turn, manifests as an increased rate of disease occurrence in
487 both experimental and wild frog populations. Temperature is, of course, correlated with a
488 multitude of other factors that may be considered alternative or additional drivers of disease
489 outbreaks. That seasonal patterns were found in ranavirosis occurrence as well as temperatures
490 raises the possibility that another seasonal factor - correlated with temperature - could have
491 driven ranavirosis outbreaks. However, temperature was associated with disease incidents at two
492 different timescales (within and across years). The association between temperature and the rate
493 of ranavirus incidents across years suggests that temperature is driving the rate of disease
494 incidents in the long term and may also drive the seasonality, since both a correlated seasonal
495 factor and a correlated non-seasonal factor acting across years would otherwise be required to
496 explain the observed patterns. Furthermore, the use of *in vitro* and *in vivo* studies in combination
497 with modeling of field datasets serves as a ‘triangulation’ process (Plowright, Sokolow, Gorman,

498 Daszak, & Foley, 2008) and strongly suggests a causal link between temperature and disease
499 occurrence in this system.

500 Our results are consistent with an historic effect of climate on the rate and timing of ranavirus
501 incidents and suggest that the invasiveness of this introduced pathogen may have been
502 influenced by the suitability of local climate. We have previously shown that ranavirus was
503 introduced to the UK and spread rapidly in England (Price et al. 2016). The small overall
504 decrease in the rate of disease incidents observed over the twenty years of data analyzed, which
505 was observed after controlling for the effect of temperature (Figure 2d), may be due to fewer
506 opportunities for spread arising from a more complete colonization of the suitable range or might
507 be explained by efforts to limit the risk of translocations through advice disseminated in the
508 media (e.g. BBC, 2008; Price et al., 2016). Alternatively, the severe impacts of recurrent
509 ranavirus outbreaks - previously shown to have caused declines of common frogs in South-East
510 England (Teacher et al., 2010) - might have led to the extirpation of populations or reduced their
511 size/density to the point where transmission no longer occurred, or disease outbreaks were no
512 longer detectable.

513 Climate projections were focused on the temporal and spatial changes in the suitable region for
514 outbreaks of ranavirosis but use a single point value from a regression model and therefore do
515 not capture uncertainty around the temperature threshold for increased risk. However, the
516 projections illustrate that any degree of climate change will likely play a role in shaping future
517 ranavirus disease dynamics in UK common frogs, altering both the geographic extent and the
518 length of the temporal window of heightened disease risk and severity. The potential impacts of
519 climate warming on disease ecology, therefore, could have critical ramifications for the
520 continued survival of amphibian populations across the UK.

521 Although it is challenging to detect disease and mortality in larval amphibians, all evidence
522 points to adult common frogs as the major life history stage and species affected by ranavirosis
523 in the wild in the UK (Cunningham, 2001; Duffus, 2009; Price et al., 2017). This observation is
524 intriguing since larval forms are usually more affected by FV3 elsewhere in the world and
525 common frog larvae have been shown to be highly susceptible to wild-type ranaviruses in the
526 laboratory (Duffus, Nichols, & Garner, 2013; Duffus, Nichols, & Garner, 2014; Gray et al.,
527 2009). Our findings suggest that temperature could be an important determinant of the
528 partitioning of disease among life-history stages, with common frog larvae metamorphosing
529 before pond water reaches a temperature high enough to trigger outbreaks of ranavirosis. This
530 situation could be altered as the climate warms and the disease season is lengthened. Shifting the
531 timing of frog disease outbreaks will alter the life history stages at risk. If common frog tadpoles
532 become affected, the abundance of susceptible hosts will be increased with concomitant impacts
533 on the ranavirus basic reproductive number (R_0) (Altizer et al., 2006); i.e. the dynamics of
534 outbreaks will be fundamentally changed, making predictions of their impacts more challenging.

535 Additional research on the effect of climatic changes on the breeding phenology of UK
536 amphibians is needed, but it has been suggested that the common frog differs from other UK
537 species in having shown no response in the timing of breeding to increasing absolute
538 temperatures (Beebee, 1995). Therefore, any compensatory changes in common frog behavior to
539 future climate change may be negligible. Common frog populations in regions where
540 temperatures become suitable for severe disease outbreaks during the larval stage might
541 experience reduced recruitment and a subsequent reduction in their capacity to persist in the
542 presence of infection. Additionally, altering the timing of outbreaks could create opportunities
543 for host jumps if other potential host species have increased contact with (high levels of) virus
544 (Hoberg & Brooks, 2015). The results highlight how species with complex life cycles might

545 undergo sudden shifts in the level of threat posed by an infectious disease if gradual changes to
546 the climate result in greater exposure/susceptibility of alternative life history stages when a
547 favorable environment had previously buffered them against the most severe impacts of disease.
548 However we did not have information about the inflection point in the viral growth curve or how
549 frog populations will respond if temperatures exceed such an inflection point, so predictions of
550 how individual outbreaks would proceed through the whole season were not attempted and it is
551 possible that impacts might be reduced in some populations due to exposure to very high
552 temperatures in the future.

553 Our laboratory findings that virulence is reduced at lower temperatures, that frogs might be
554 better able to manage infections at lower temperatures, together with field records showing a
555 mitigating effect of shading, pond volume and vegetation (which might also be due to lowered
556 temperatures of frogs) on incidence and severity of ranavirosis, point to possible steps for
557 mitigation. Thermoregulatory behavior leading to an increase in body temperature above normal
558 range (“behavioral fever”) is known to be important for disease mitigation in a range of
559 ectotherms (Boltaña et al., 2013; Elliot, Blanford, & Thomas, 2002; Monagas & Gatten, 1983).
560 Whether thermoregulatory behavior serves as an amphibian strategy for managing infections
561 remains to be elucidated (Sauer et al., 2018) but shading has been proposed as an environmental
562 feature that might have negatively impacted the ability of newts to clear chytrid infections
563 because it limited the availability of warm-water patches (Raffel et al., 2010). Staying cool can
564 be a more difficult challenge than getting warm for many ectotherms (Kearney, Shine, & Porter,
565 2009), but the provision of suitable opportunities for behavioral regulation of body temperature
566 in the form of shading, log piles and larger ponds might help to manage the severity of future
567 outbreaks and warrants further study.

568 Ranaviriosis has had a major impact on common frog populations in south-east England (Teacher
569 et al., 2010) and the current study suggests that these impacts may become greater and more
570 widespread (in the UK and elsewhere) if future climate change projections are realized. Our
571 results - as well as the predictions that follow from them - are strengthened through the use of a
572 model system that allows us to investigate possible drivers of field epidemiology using
573 laboratory experiments, both at the cellular and whole animal levels. Together, our results
574 present a clear case of the environment modulating an important host-pathogen interaction. Few
575 previous studies have convincingly shown how climate change affects disease emergence in wild
576 animal populations, but we have been able to demonstrate a historic impact of warming in the
577 wild and then tease apart relationships between the environment, host and pathogen in the
578 laboratory. These results underline the importance of accounting for the effects of local
579 environmental drivers to predict the dynamics of an invading pathogen (Cohen et al., 2018,
580 2017; Raffel et al., 2013).

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585 before any work was undertaken and procedures were performed under UK Home Office
586 licenses P8897246A & 80/2214. The authors declare that there are no conflicts of interest. All
587 data and code required to reproduce the analyses in this article are included in the Supporting
588 information files (Data S1 to Data S8).

589 **Author contributions**

590 SJP & AAC designed the *in vitro* experiments, the lab work was performed by CO & WL, and
591 analysis was performed by SJP with help from CO. SJP designed the *in vivo* experiments with
592 help from TWJG & RAN, the lab work was performed by SJP, WL & CS, and the results were
593 analyzed by SJP. AAC oversaw collection of the epidemiological datasets, SJP, RP, TWJG,
594 RAN & FB planned the analyses which were conducted by SJP and RAN. SJP and RP planned
595 the climate change projections, which were performed by SJP. SJP & FB wrote the first draft of
596 the manuscript which was edited by all authors.

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862

863 **Supporting information Appendix S1**

864 **Supplementary methods: *In vivo* investigation of dose response and disease** 865 **progression**

866 **Acclimation, housing and husbandry:** Common frogs (*Rana temporaria*) were reared from
867 eggs (collected at sites with no history of ranavirosis) through metamorphosis and overwintered
868 once. Thirty individuals were randomly allocated to each of three experimental treatments
869 (“high” dose, “low” dose, and a sham control). They were weighed, transferred to individual 1.6
870 L plastic boxes with a non-airtight lid, a moist paper substrate, and a plastic plant pot for cover
871 and acclimated to the experimental housing for two days prior to viral exposure. The room was
872 climate controlled with lighting set to a constant 12-hour day/night cycle. Environmental
873 conditions (temperature and humidity) were monitored daily for two weeks prior to and
874 throughout the experiment. Mean temperature in the room was 20.2°C (18.6-21.1°C) and mean
875 humidity was 59% (47-79%). Boxes were rotated daily to ensure there were no persistent effects
876 of location on frog responses. Animals were fed brown crickets on alternate days and cleaned
877 and weighed on every fourth day.

878 **Exposure:** Ranavirus isolate RUK13 was passaged three times on confluent fathead minnow
879 (FHM) cells in maintenance medium (EMEM plus 10% FBS plus 1% L-glutamine) and spun at
880 800g to remove cell debris. Virus titer was calculated using a TCID₅₀ (50% Tissue Culture
881 Infectious Dose) protocol in a 96-well flat-bottomed tissue culture plate. TCID₅₀ for the virus
882 stock was then calculated using the Reed and Muench method (Reed & Muench, 1938). Stock
883 virus and sham media were initially diluted in cell culture medium prior to a second dilution step
884 in aged tap-water to ensure that the inoculum for each treatment contained the same total volume
885 of cell culture medium. Viral titers of inocula were $1 \times 10^{4.6}$ TCID₅₀ mL⁻¹ of virus for the high

886 dose treatment and $1 \times 10^{3.3}$ TCID₅₀ mL⁻¹ for the low dose treatment. Animals were bath-
887 exposed in individual 0.07 L plastic boxes for 7 hours prior to returning individuals to their
888 housing.

889 **Sampling:** Animals were monitored daily and scored for changes in spontaneous behaviour,
890 behaviour during handling (when applicable), food intake, stools, and body condition. Animals
891 were also monitored for signs of disease - including reddening of skin, petechial hemorrhaging,
892 ulceration, bleeding, emaciation and lethargy - as well as other physical changes. Swab samples
893 from the pericloacal region of each animal were collected at day 8 using sterile dry swabs
894 (MWE) by rolling the swab tip forwards and backwards across the cloacal entrance three times.
895 Day 8 after exposure was predicted to be the approximate midpoint of the experiment based on
896 previous infections in adult common frogs (Cunningham, Hyatt, Russell, & Bennett, 2007).
897 Approximately half of the animals remaining in the control and low dose treatments were
898 selected at random for euthanasia by a schedule 1 method for amphibians. Five of the animals in
899 the low dose treatment died overnight prior to euthanasia so eight (of 26) animals were
900 euthanized along with 11 (of 22) control animals. Dead animals were examined within twelve
901 hours of death for skin lesions and abnormalities, and post-mortem pericloacal swabs taken. The
902 carcasses (frozen or within three days of death) were examined post-mortem for signs of disease.
903 Liver and kidney samples (approximately 20mg) and the third toe of right hind-foot were frozen
904 at -20°C for DNA extraction and PCR.

905 **Viral load estimation:** tissue samples (and positive and negative extraction control tissues) were
906 disrupted by beating with 5mm beads at 15Hz for 20s in a Qiagen Tissue Lyser II. The tips of
907 swabs were cut off using a new scalpel blade for each sample and placed in individual
908 microcentrifuge tubes with the Promega Wizard digestion mix plus 0.03-0.04g of 0.5mm silica
909 microbeads and beaten for 45s at 30Hz in the tissue lyser alongside extraction controls. All

910 samples were digested overnight for 18 hours at 56°C prior to extraction with a Promega Wizard
911 SV96 Genomic DNA Purification kit according to the manufacturer's guidelines. DNA was
912 eluted in a single step leading to a total volume of approximately 250µL. DNA concentrations
913 were measured using on a NanoDrop 2000 spectrophotometer.

914 A quantitative polymerase chain reaction was used to estimate viral load. Primers from
915 Holopainen, Honkanen, Jensen, Ariel, & Tapiovaara (2011) were used to amplify a 93bp
916 fragment of the largest DNA polymerase subunit. SYTO13 was used as the dsDNA reporter dye
917 (Monis, Giglio, & Saint, 2005). Each sample was screened in duplicate wells using a reaction
918 mix of 12.5µL of Promega GoTaq Hot Start Colorless Mastermix, 0.625µL of 10µM forward
919 primer, 0.625µL of 10µM reverse primer, 0.25µL SYTO13 (500µM stock), 0.5µL ROX, 5.5µL
920 of nuclease free water and 5µL of template DNA. Samples were run on 96-well plates with
921 negative controls (duplicate wells with nuclease free water as template) and standards. Standards
922 were generated from a cultured stock of ranavirus isolate RUK11 (Cunningham, Hyatt, Russell,
923 & Bennett, 2007) - quantified following the same protocol described above for RUK13, prior to
924 DNA extraction of a 300µL aliquot of stock virus with a DNAeasy Blood and Tissue spin
925 column following the manufacturer's protocol for 'Purification of Total DNA from Animal
926 Blood or Cells'. Standards were added to each plate in a dilution series in duplicates. Plates were
927 run on an Applied Biosystems Step One Plus thermocycler using the same settings as
928 Holopainen et al. (2011). If one of the replicates failed to amplify or there were large standard
929 deviations of mean CT values samples were rerun in duplicate. Baselines for background
930 fluorescence, thresholds for calculation of CT scores, and exclusion of samples with unusual
931 amplification or melt curves was performed using the AB Step One Plus software. Standard
932 curves were checked to ensure R² scores greater than 0.99 and efficiency scores in the range 80-

933 120%. To account for differences in extraction efficiency, virus quantity scores were
934 standardized by dividing by the concentration of DNA in each sample extract.

935 **Analyses:** Associations between signs of disease and exposure to ranavirus (low and high dose
936 combined) were tested using Fisher's exact tests in R. The effect of dose on the amount of virus
937 in tissues of dead animals was investigated using a linear mixed-effects (LME) model with dose
938 and sample type (liver, kidney, toeclip) as fixed effects and individual id as a random effect to
939 account for repeated measures (multiple tissues) from the same individuals [function *lme*, R
940 package *nlme*; Pinheiro, Bates, DebRoy, Sarkar, & R Core Team (2018)]. Viral loads at death
941 were compared to viral loads in live animals using individuals sampled by swabbing between
942 days 7 and 9 post-exposure and again at death (LME model with sample time (alive/dead) as a
943 fixed effect and individual id as a random effect) as well as sets of animals that either died due to
944 disease or were euthanized part-way through the experiment (LME model with the fate of
945 individuals [euthanized/died] as a fixed effect and id as a random effect). The effect of dose on
946 viral loads in all animals swabbed between days 7 and 9 post-exposure was assessed using a
947 linear model. The effect of dose on the timing of signs of disease (disease progression, measured
948 as the first day following exposure that a sign was observed in an individual) was analyzed using
949 a generalized mixed-effects model [Poisson family; *glmer* function in the R package *lme4*; Bates,
950 Mächler, Bolker, & Walker (2015)] with dose and the type of sign as fixed effects and individual
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952 **Supplementary references**

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972 **Supporting information Appendix S2**

973 **Supplementary methods: Proportion of incidents in Frog Mortality Project** 974 **(FMP) database classified as ranavirosis-consistent**

975 The FMP database contained variables describing the month and year that mortality incidents
976 began as well as georeferences (latitude and longitude on the WGS84 datum; coordinate
977 reference system, EPSG:4326). Monthly averages of the daily maximum temperature for all UK
978 5 x 5 km grid squares covering the study period (1991 - 2010) were downloaded from the Met
979 Office UKCP09 dataset (Met Office, 2017). All individual points from the FMP database were
980 overlaid on the grid of 5 x 5 km grid squares to retrieve the average local maximum temperature
981 at the onset of each mortality incident.

982 **Supplementary references**

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986

987 **Supporting information Appendix S3**

988 **Supplementary methods: *In vivo* investigation of effect of temperature on** 989 **virulence**

990 In order to validate results from cell culture models *in vivo*, 60 common frogs (*Rana temporaria*)
991 were reared from eggs (collected at sites with no history of ranaviriosis) through metamorphosis
992 and overwintered once. Frogs were randomly allocated to one of six treatments (10 animals per
993 treatment); three exposure treatments (sham, RUK11, RUK13) each at two temperatures (20°C
994 and 27°C). The animals were individually housed in 0.7 L volume polypropylene boxes with lids
995 which contained moist paper towels and half a plastic plant pot for cover. Frogs were weighed
996 then acclimated to their boxes at the two incubation temperatures for 72 hours prior to the start of
997 the experiment. They were fed on approximately six small crickets twice weekly after the paper
998 towel substrates had been replaced. During acclimation and throughout the experiment, animals
999 were subjected to a twelve-hour day-night cycle.

1000 Climate controlled chambers were constructed from polystyrene boxes by installing thermostat-
1001 controlled heat mats covered in a layer of vermiculite to distribute heat evenly and computer fans
1002 to aid air flow. Large rectangular holes were cut in one side of each polystyrene box and covered
1003 inside and out with Perspex sheets to create double-glazed windows which let light in whilst
1004 insulating against heat transfer. Fifteen frogs in their individual boxes were housed in each
1005 polystyrene box along with an additional (identical) box containing an iButton temperature
1006 logger (Measurement Systems Ltd, UK) programmed to log the temperature at 30-minute
1007 intervals for the duration of the experiment. Boxes were stacked two or three high at the start of
1008 the experiment and raised above the heat mat on wire baking trays with space around each stack
1009 for airflow. Four polystyrene boxes - two at the “high” temperature (27°C) and two at the “low”

1010 temperature (20°C) - were used with each climate chamber containing five animal boxes from
1011 each of the three exposure treatments. Temperatures in all four climate chambers were stable
1012 around the set points [mean temperatures (standard deviation): Chamber 1 = 26.7 (0.72),
1013 Chamber 2 = 26.5 (0.47), Chamber 3 = 19.9 (0.21), Chamber 4 = 19.9 (0.21)] but were subject
1014 to daily, short-term fluctuations coinciding with removing chamber lids to perform checks
1015 (Figure S1).

1016 The viral inocula were generated by expanding each isolate in EPC cells at 27°C and titers
1017 estimated using TCID₅₀ (Reed & Muench, 1938). Sham exposure inoculum consisted of
1018 supernatant from EPC culture media (Eagle's minimum essential media supplemented with 10%
1019 fetal bovine serum and 1% L-glutamine) with cells removed by centrifugation for ten minutes at
1020 800g. Titers of the two virus isolates were not equalized; RUK13 was used at 1.58×10^5 TCID₅₀
1021 mL⁻¹ and RUK11 at 1.58×10^7 TCID₅₀ mL⁻¹. Animals were transferred to 0.07 L polypropylene
1022 boxes which allowed animals to move around but prevented them from climbing. Boxes were
1023 individually numbered, and individuals allocated to treatments according to a random number
1024 draw without replacement made ahead of time using R. The inoculum - 10 mL volume sufficient
1025 to cover the base of the box and ensure that an animal's ventrum was always in contact with the
1026 media - was added by pipetting through a hole in the lid. All animals were subjected to a six-
1027 hour exposure period at room temperature before being returned to their boxes in the climate-
1028 controlled chambers. Individuals were monitored twice daily for signs of disease for the entire
1029 duration of the experiment. Investigators were not blinded to temperature treatment but were
1030 blinded to exposure treatment by using coded IDs. Individuals reaching humane endpoints and
1031 all surviving individuals at the end of the experiment (day eight post-exposure) were euthanized
1032 following a schedule 1 method for amphibians. The number of hours post-exposure that animals
1033 were found dead or at endpoints were recorded for survival analysis.

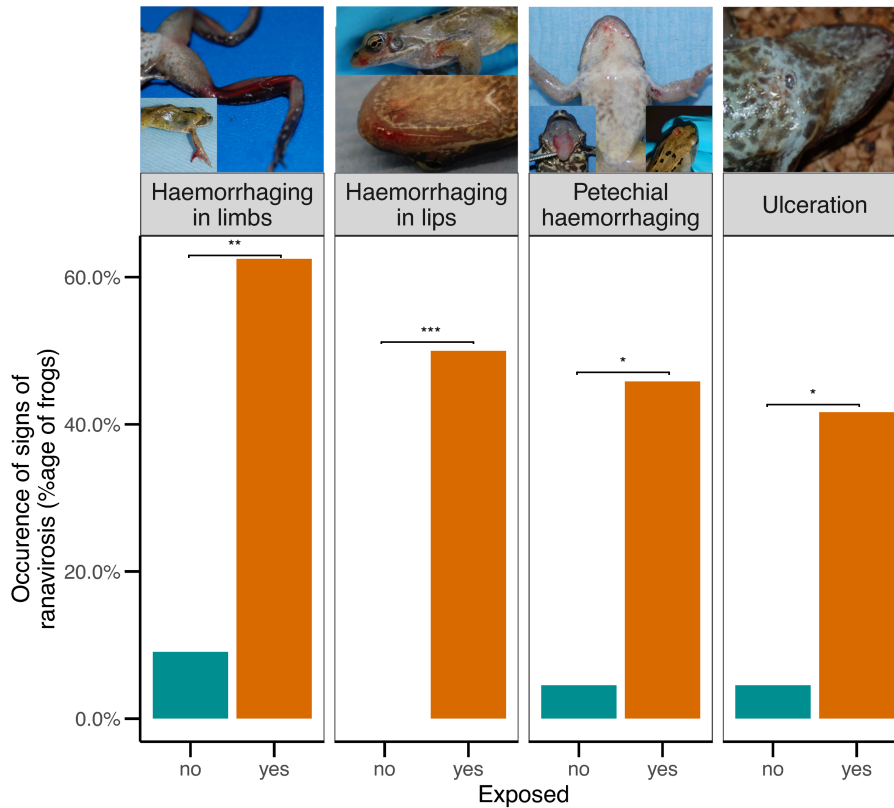
1034 **Supplementary references**

1035 Reed, L. J., & Muench, H. (1938). A Simple Method of Estimating Fifty Per Cent Endpoints.

1036 *American Journal of Epidemiology*, 27(3), 493–497. Retrieved from

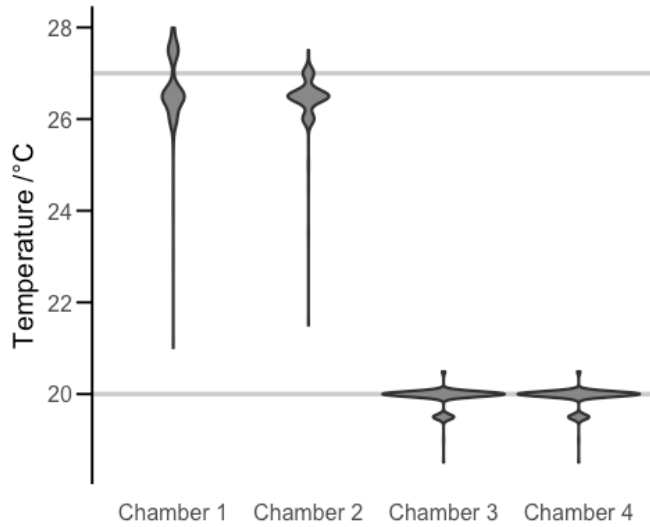
1037 <http://aje.oxfordjournals.org/content/27/3/493>

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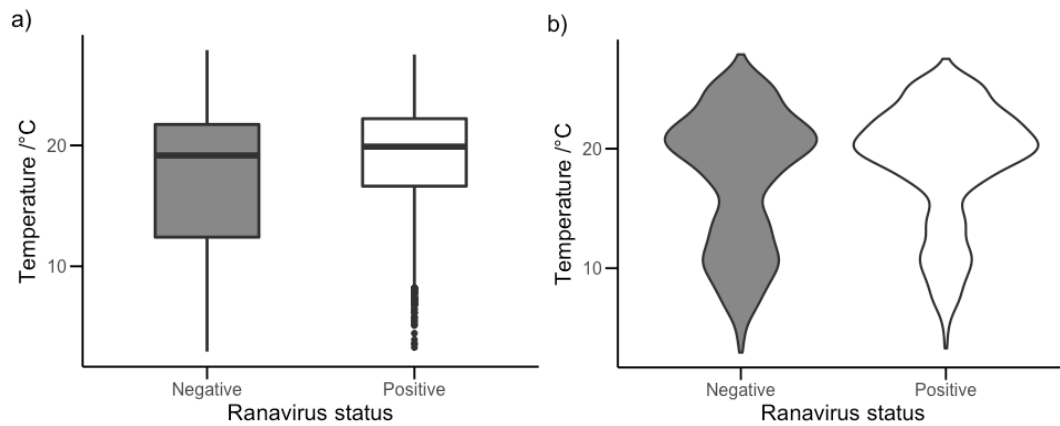
1040 **Figure S1. Association of pathologies with exposure to ranavirus.** Each panel shows the
 1041 proportion of experimental animals with observations of signs of disease when juvenile common
 1042 frogs (*Rana temporaria*) were exposed to “low” and “high” doses of ranavirus isolate RUK13
 1043 (“exposed”) as well as a sham control. Asterisks (*) above plots represent p-values from Fisher’s
 1044 exact tests of association of signs with exposure treatment: * - $0.01 < p < 0.05$; ** - $0.001 < p < 0.01$;
 1045 *** - $p < 0.001$). A subset of pathologies expected to be identifiable by untrained citizen scientists
 1046 were selected, plotted and are represented in the photographs above plots.



1047

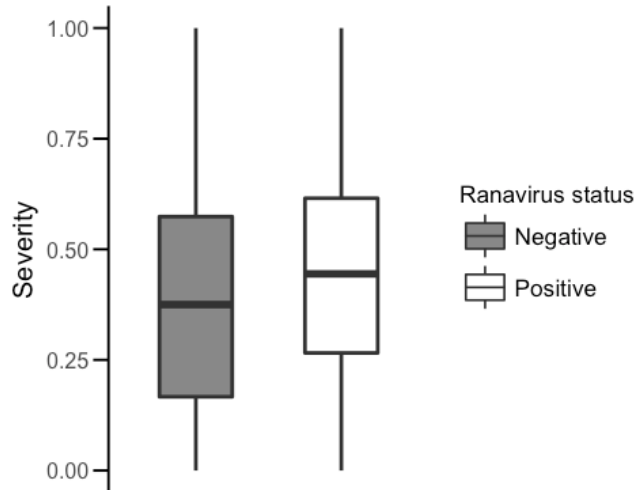
1048 **Figure S2. Temperature distribution in each of four climate-controlled chambers across the**
 1049 **duration of the *in vivo* temperature experiment.** The horizontal lines at 20°C and 27°C show
 1050 the set points for chambers.

1051



1052

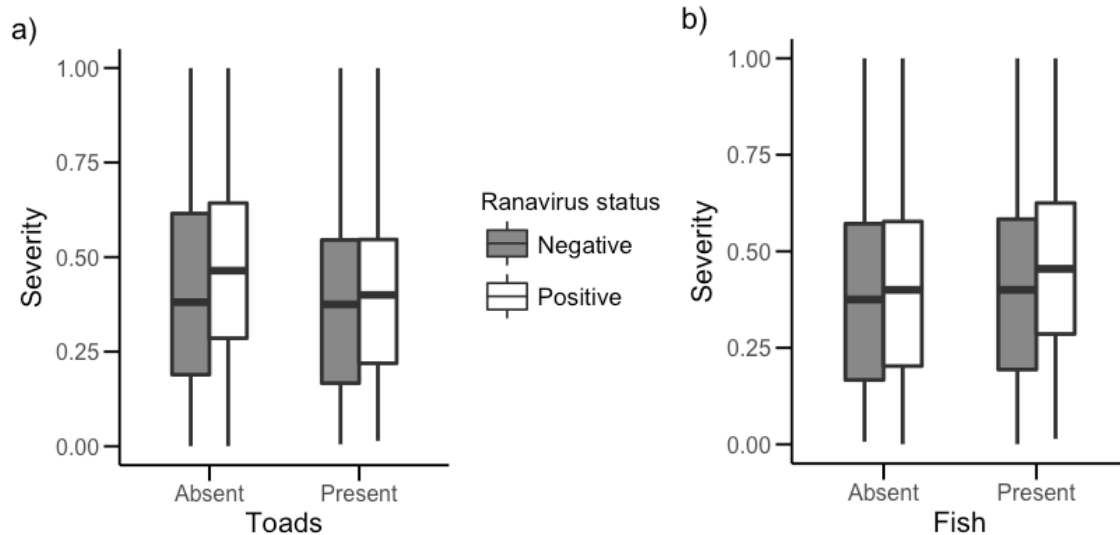
1053 **Figure S3. Temperature distribution (average daily maximum temperature) in month of**
 1054 **onset of frog mortality for ranavirosis incidents compared to those caused by other factors.**
 1055 Plots comparing distributions: a) Boxplot (lower quartile, median, upper quartile and
 1056 interquartile range (upper quartile - lower quartile; central 50% of the data); whiskers extend to
 1057 the most extreme data point which is no more than 1.5 times the interquartile range from the box;
 1058 outliers shown as individual points). b) Violin plot.



1059

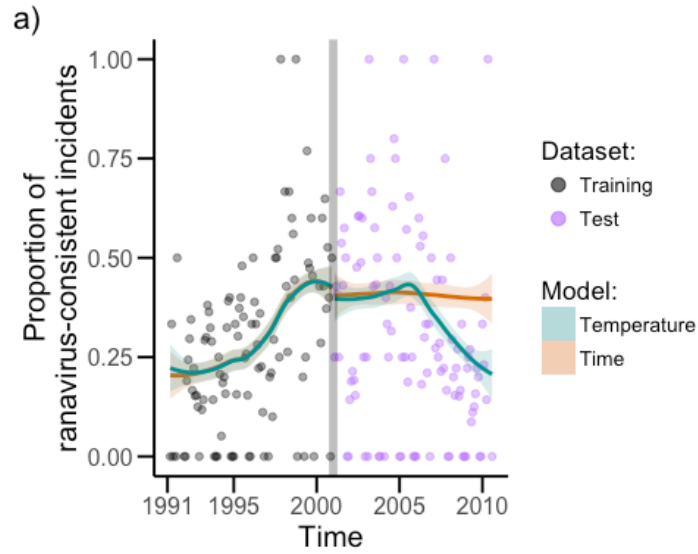
1060 **Figure S4. Severity (estimated proportion of population that died) of mortality incidents**
 1061 **caused by ranavirus was higher than for those caused by other factors.** Boxes represent
 1062 lower quartile, median, upper quartile and interquartile range (upper quartile - lower quartile;
 1063 central 50% of the data); whiskers extend to the most extreme data point which is no more than
 1064 1.5 times the interquartile range from the box.

1065

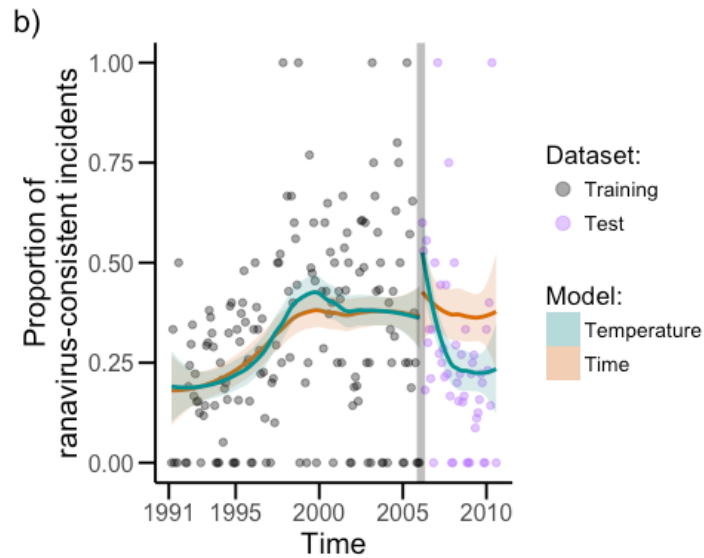


1066

1067 **Figure S5. Effect of the presence of other species on the severity (estimated proportion of**
 1068 **population that died) of ranavirus outbreaks in common frogs.** a) Toads reduced the severity
 1069 of ranavirus mortality incidents. b) Fish increased the severity of ranavirus mortality incidents.
 1070 Boxplots represent lower quartile, median, upper quartile and interquartile range (upper quartile -
 1071 lower quartile; central 50% of the data); whiskers extend to the most extreme data point which is
 1072 no more than 1.5 times the interquartile range from the box.

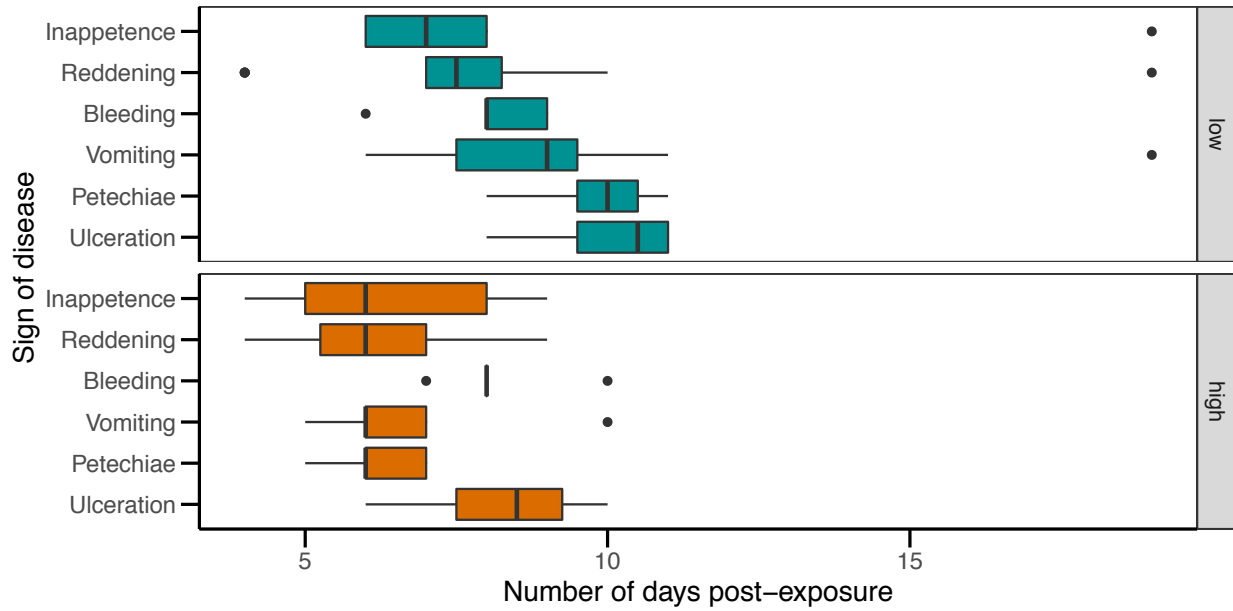


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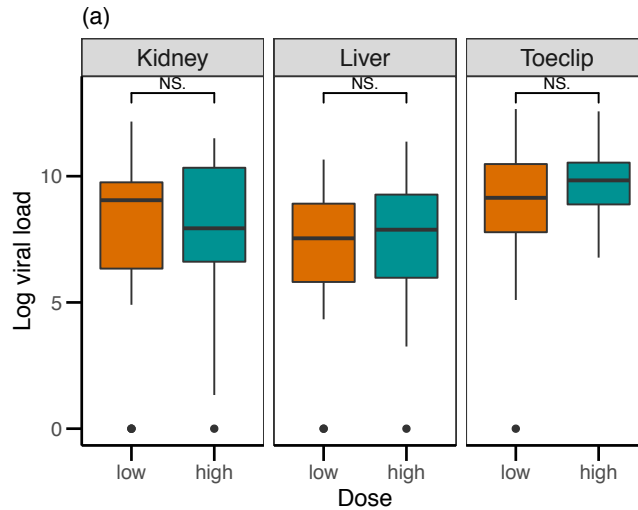
1074

1075 **Figure S6. Comparison of power of generalized additive models incorporating either a**
 1076 **smoothed time trend or smoothed temperature to predict rates of ranavirosis.** a) Models
 1077 trained on data for 1991-2000 and tested on data for 2001-2010. b) Models trained on data for
 1078 1991-2005 and tested on data for 2006-2010. Vertical grey bar represents break between training
 1079 and test data. Points represent data from the training and test set. Lines are loess smoothed trends
 1080 predicted from the models with 95% confidence interval shaded.

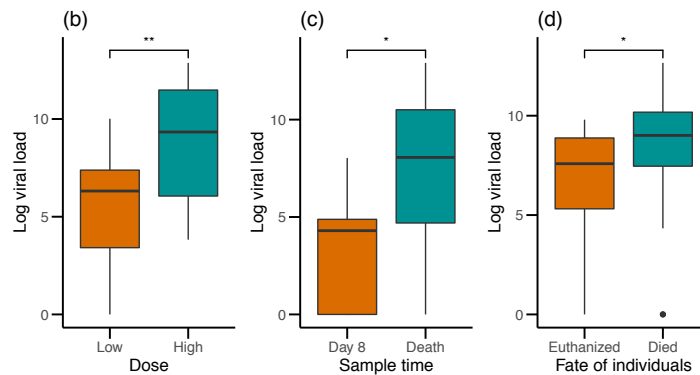


1081

1082 **Figure S7. Effect of exposure dose (high or low) on the timing of signs of ranavirosis.** The
 1083 low dose treatment was subject to a significant lag in the onset of disease compared to high dose
 1084 (generalized linear mixed-effects model, $p=0.004$).



1085



1086

1087 **Figure S8. Comparisons of viral load in experimentally infected live and dead common**
 1088 **frogs suggested a threshold for disease.** (a) At death there is no difference in the viral load in
 1089 tissues between animals receiving different inoculum doses ($p = 0.51$; linear mixed-effects model
 1090 with dose and sample type as fixed effects and individual as a random effect), but (b) loads were
 1091 higher in the high dose treatment at day 8 ($p = 0.006$; linear model, $\log \text{viral load} \sim \text{dose}$) and
 1092 increased to death in (c) a subset of animals sampled over time ($p = 0.01$; linear mixed-effects
 1093 model with sample time as a fixed effect and individual as a random effect) and (d) between
 1094 animals euthanized at day 8 and those that died ($p = 0.0495$; linear mixed-effects model with fate
 1095 of individuals [euthanized/died] as a fixed effect and individual as a random effect), suggesting
 1096 that higher dose increases the initial intensity of infection, giving the virus a head-start, but that
 1097 all animals reached a threshold viral load that resulted in disease.

1098 **Table S1.** Model output from logistic model of incidence of ranavirus outbreaks as function of a
 1099 range of variables describing the local environment of amphibian mortality incidents from Frog
 1100 Mortality Project reports.

Term	Estimate	Std. Error	z-value	Pr(> t)
Intercept	-3.371	0.488	-6.906	<0.001
Av. Max. daily temperature	0.247	0.059	4.225	<0.001
Quadratic term: Av. Max. daily temperature	-0.006	0.002	-3.530	<0.001
Shading (Lots)	-0.278	0.137	-2.029	0.042
Newts (present)	0.268	0.089	3.026	0.002
Fish (present)	0.260	0.099	2.623	0.009

1101

1102 **Table S2.** Model output from linear model of outbreak severity as function of a range of
 1103 variables describing the local environment of amphibian mortality incidents from Frog Mortality
 1104 Project reports.

Term	Estimate	Std. Error	t-value	Pr(> t)
Intercept	-1.247	0.365	-3.422	<0.001
Ranavirus status (positive)	-0.615	0.249	-2.470	0.014
Av. Max. daily temperature	0.030	0.004	6.882	<0.001
log ₂ (pond volume)	-0.075	0.019	-4.063	<0.001
Quadratic term: log ₂ (pond volume)	0.009	0.002	4.280	<0.001
Shading (Lots)	-0.186	0.067	-2.786	0.005
Marginal vegetation (Lots)	-0.205	0.054	-3.762	<0.001
Toads (present)	-0.209	0.045	-4.658	<0.001
Fish (present)	0.224	0.051	4.381	<0.001
Region 2	0.201	0.377	0.533	0.594
Region 3	-0.185	0.413	-0.449	0.653
Region 4	-0.152	0.397	-0.383	0.701
Region 5	-0.101	0.388	-0.261	0.794
Region 6	0.354	0.360	0.985	0.325
Region 7	0.418	0.360	1.159	0.247
Region 8	0.332	0.358	0.928	0.353
Region 9	0.162	0.365	0.444	0.657
Region 10	0.071	0.477	0.149	0.882
Region 11	-0.053	1.154	-0.046	0.963
Interactions				
Ranavirus status (positive): Av. max. daily temp	0.043	0.013	3.401	<0.001

1105

1106 **Table S3.** Model output from GAM component of GAMM to decompose seasonal and across
 1107 year trends in temperature data for study region from 1991 to 2010.

1108 a) parametric component

term	Estimate	Std. Error	t value	Pr(> t)
Intercept	14.855	0.116	127.72	<0.001

1109 a) Smoothed component

term	edf	Ref.df	F	p-value
s(month)	7.018	8.000	272.033	<0.001
s(time)	2.361	2.361	2.786	0.048

1110

1111 **Table S4.** Model output from GAM component of GAMM to decompose seasonal and across
 1112 year trends in ranavirosis data from 1991 to 2010.

1113 a) parametric component

term	Estimate	Std. Error	t value	Pr(> t)
Intercept	0.272	0.014	19.017	<0.001

1114 a) Smoothed component

term	edf	Ref.df	F	p-value
s(month)	4.966	8.000	7.583	<0.001
s(time)	2.974	2.974	7.078	<0.001

1115

1116 **Table S5.** Model output from Generalized Additive Model of ranavirosis rate as a function of
1117 season, time and the across-year trend in temperature.

1118 a) parametric component

term	Estimate	Std. Error	z value	Pr(> z)
Intercept	-0.632	0.099	-6.387	<0.001
time	-0.003	0.001	-4.233	<0.001

1119 b) Smoothed component

term	edf	Ref.df	Chi.sq	p-value
s(month)	7.648	8.000	73.629	<0.001
s(temptrend)	6.647	7.379	137.952	<0.001

1120

1121 **Data S1.** RData file containing all data tables used in analyses.

1122 **Data S2.** Plain text R file containing code required to run analyses.

1123 **Data S3.** Viral growth *in vitro* dataset (CSV file).

1124 **Data S4.** Raw data from data loggers in climate chambers used for *in vivo* test of effect of
1125 temperature (CSV file).

1126 **Data S5.** Frog Mortality Project dataset (CSV file).

1127 **Data S6.** Ranavirosis outbreaks in wild confirmed by molecular methods dataset (CSV file).

1128 **Data S7.** Raw room temperature data for *in vivo* test of effect of temperature (CSV file).

1129 **Data S8.** Survival data from *in vivo* test of effect of temperature (CSV file).

1130