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27 Abstract:

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The global food system is a major energy user and a relevant contributor to climate change. To date, the literature on the energy profile of food systems addresses individual countries and/or food products, and therefore a comparable assessment across regions is still missing. This paper uses a global multi-regional environmentally-extended input-output database in combination with newly constructed net energy use accounts to provide a production and consumption-based stock-take of energy use in the food system across different world regions for the period 2000-2015.

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Overall, the ratio between energy use in the food system and the economy is slowly decreasing.
Likewise, the absolute values point towards a relative decoupling between energy use and food
production, as well as to relevant differences in energy types, users and consumption patterns
across world regions. The use of (inefficient) traditional biomass for cooking substantially
reduces the expected gap between per capita figures in high- and low-income countries.

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43 The variety of energy profiles and the higher exposure to energy security issues compared to the 44 total economy in some regions suggests that interventions in the system should consider the 45 geographical context. Reducing energy use and decarbonizing the supply chains of food products 46 will require a combination of technological measures and behavioral changes in consumption

47 patterns. Interventions should consider the effects beyond the direct effects on energy use, since
48 changing production and consumption patterns in the food system can lead to positive spillovers
49 in the social and environmental dimensions outlined in the Sustainable Development Goals.

51 **1. INTRODUCTION**

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The global food system is a major energy user, responsible for between 15 and 20% of total energy use (Beckman et al. 2013). Energy is used in different forms throughout all the life cycle stages of food. Diesel serves as fuel in agricultural machinery and transportation activities, natural gas is a key input in the production of fertilizers, electricity is used to store and prepare food, etc. As a result, food systems are connected to several environmental impacts through the use of energy, most notably climate change.

59

60 The overall use of energy in the food system is shaped by several global factors. Growing 61 populations and increasing affluence has resulted in large increases in food consumption and 62 significant changes in dietary compositions, both of which impact heavily on energy inputs in the 63 food system (Behrens et al. 2017). Increasing consumption volumes often require either the 64 development of new arable land (requiring further energy input), or increasing yields (often 65 resulting in increased fertilizers and energy inputs) (Woods et al. 2010). Changes in dietary 66 composition, on the other hand, are driven by rising affluence, a process commonly termed the 67 nutrition transition whereby diets move from vegetal staples to increasing amounts of animal 68 products and processed foods (Popkin 2006). This increased emphasis on animal products 69 increases the dependence on energy inputs as they are generally less efficient than vegetal 70 alternatives (Pelletier et al. 2008). These trends have driven large developments in food system 71 energy use (Canning et al. 2010) and will continue to do so for the foreseeable future (Woods et 72 al. 2010).

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74 The global food system is also characterized for being very inefficient with regard to waste. 75 Currently, a third of all edible food is discarded globally along the life cycle stages of food 76 (Gustavsson et al. 2011). Acting on it, as foreseen under the Sustainable Development Goal 12.3 77 ("By 2030, halve per capita global food waste at the retail and consumer levels and reduce food 78 losses along production and supply chains, including post-harvest losses") (UN 2015b), could 79 result in important environmental savings, including energy resources (Usubiaga et al. 2018). A 80 further trend in food systems has been towards greater volumes of trade between nations, with 81 increasing percentages of environmental impacts embodied in traded agricultural goods. For 82 example, a quarter of all agricultural emissions are traded (Kander et al. 2015), along with 22% 83 of all freshwater withdrawal (Dalin et al. 2012).

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85 The combined pressures of increasing population and wealth will continue and intensify during a 86 period in which society is under increasing pressure to transition to renewable and low-carbon 87 technologies. The food system will need to transition but will face specific technological and 88 social challenges distinct from those seen in other sectors. Compounding this is the need for 89 heavy mobile machinery for production and pre-processing steps (ploughing, reaping, threshing, 90 winnowing etc.), which require large mobile sources of energy to operate using high energy 91 density fuels such as diesel. For example, while 15% of the overall electricity mix in the 92 European Union was from renewable sources in 2015, this drops to only 7% in food systems 93 (Monforti-Ferrario et al. 2015). Socio-economic challenges to transitioning to more efficient 94 food production systems in some producer nations include the lack of financial and human 95 resources, and inertia due to conservative approaches of producers. This comes on top of the 96 existing barriers to changing food consumption patterns (Mozaffarian et al. 2018).

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101 food systems and identifies the research gaps addressed in this paper. Section 3 describes the 102 methodology. Section 4 and 5 present and discuss the results, while section 6 concludes. 103 104 2. PREVIOUS ASSESSMENTS OF THE ENERGY PROFILE OF THE FOOD SYSTEM 105 106 Given these developments, research on the energy use in food systems has become increasingly 107 relevant from a policy perspective. Although different methodological approaches have been 108 used (Coley et al. 1998; Eshel and Martin 2006), the dominant approach has been Life-Cycle 109 Assessments (LCA) (Pelletier et al. 2011). There have been several investigations of the large 110 differences between food products (Foster et al. 2006; Pimentel et al. 2008; Laso et al. 2018). 111 LCA assessments have been combined to form a basket of goods which may represent a typical 112 diet (Monforti-Ferrario et al. 2015). They have also been used to investigate the energy 113 requirements of different nutrients (González et al. 2011). Assessments of energy use in specific 114 parts of the supply chain have been prominent, particularly on food miles and the regionality of 115 production (Pretty et al. 2005; Hauwermeiren et al. 2007). Assessments of other areas of the 116 supply chain have been less numerous due to methodological difficulties, for instance in 117 packaging (Sanjuán et al. 2014; Molina-besch et al. 2019). Indirectly, many LCA studies have 118 some consideration of energy consumption by focusing on greenhouse gas emissions, but the 119 underlying composition of energy inputs into food is obscured (Tilman and Clark 2014).

Against this background, the paper intends to provide a stock-take of energy used in the global

follows. Section 2 provides an overview of previous assessments of the energy requirements of

food system and shed light on the energy profile of regional food systems. It is structured as

120	However, there are several weaknesses of LCA for how energy is used across food systems (as
121	opposed to individual product analysis). Firstly, being a bottom-up analysis, decisions on
122	boundary settings, allocation choices, and background data makes results difficult to standardize
123	and compare across studies (Ayres 1995). Secondly, there are estimation challenges when it
124	comes to truncation errors, that is, where the boundaries of the system are drawn (Ward et al.
125	2017). Third, while there is increasing attention on the regionalization of data within LCAs,
126	many use averages in nations rather than including different production factors across nations in
127	the food supply chain (Yang and Heijungs 2016).
128	
129	At a higher level of aggregation, encompassing broader sectors or product groups,
130	environmentally extended Input-Output Analysis (EEIOA) has been used to estimate direct and
131	indirect energy consumption across an entire economy. However, these analyses until now have
132	been based on national investigations (Ozkan et al. 2004b; Bekhet and Abdullah 2010; Canning
	been based on national investigations (Ozkan et al. 20040, Deknet and Abdunan 2010, Canning
133	et al. 2010; Cao et al. 2010; Reynolds et al. 2015; Sherwood et al. 2017; Song et al. 2019) or
133 134	et al. 2010; Cao et al. 2010; Reynolds et al. 2015; Sherwood et al. 2017; Song et al. 2019) or highly aggregated food sectors (Alcántara and Duarte 2004). For national analyses, EEIOA
133 134 135	et al. 2010; Cao et al. 2010; Reynolds et al. 2015; Sherwood et al. 2017; Song et al. 2019) or highly aggregated food sectors (Alcántara and Duarte 2004). For national analyses, EEIOA studies have had to be complemented with exogenous data for supply chains outside the nation
 133 134 135 136 	et al. 2010; Cao et al. 2010; Reynolds et al. 2015; Sherwood et al. 2017; Song et al. 2019) or highly aggregated food sectors (Alcántara and Duarte 2004). For national analyses, EEIOA studies have had to be complemented with exogenous data for supply chains outside the nation of investigation, leading to a number of simplifications and assumptions (Monforti-Ferrario et al.



There is also a 'geographical-gap' in studies as both LCA and EEIOA studies have focused predominantly on high-income nations (de Haes 2004; Aleksandrowicz et al. 2016). An exception is Turkey, for which several studies of different food types have been made (Ozkan et al. 2004b; Ozkan et al. 2004a; Hatirli et al. 2005; Kizilaslan 2009). A key challenge is to expand these analyses for other nations in a comparable manner which will incorporate heterogeneities in the amount and types of energy used in the food system (Pelletier et al. 2011).

148 Using environmentally-extended input-output methods generally trades product specificity in 149 LCAs for a broader, global scope. Here we present, to our knowledge, the first comprehensive 150 analysis of energy in food systems using a global environmentally-extended multi-regional 151 input-output (EEMRIO) database. We analyze the use of energy in food supply chains using 152 EXIOBASE, an environmentally-extended database with a high resolution in both food products, 153 energy types, and also in related activities (Stadler et al. 2018). The level of product 154 disaggregation available allows us to isolate the energy demands (in amount and type) of 155 different food groups, while following this energy use through the supply chain. The database 156 gives information on pre-production (i.e. energy for fertilizer inputs), production, processing, 157 transport, consumption, and disposal. The database represents 10 middle-income nations as well 158 as 34 high-income nations (with the remaining nations represented as five aggregated regions). 159 160 This work addresses three key research gaps: firstly, the inclusion of several middle-income 161 nations broadens our knowledge of food systems outside high-income nations; secondly, the 162 coverage of different food and energy types allows for increased insight into how energy is used 163 at an international level; finally, the inclusion of a time-series and GRMIO allows for the 164 investigation of the evolution of energy use in international food supply chains in a way that 165 previously has not been possible.

166

167 **3. METHODS**

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169 In this paper, we characterize the energy profile of regional food systems both from the

170 production and consumption perspectives for the period 2000-2015. To this end, we use a global

171 EEMRIO database with high sectoral detail, which in this context can provide policy-relevant 172 insights on energy mixes, drivers, energy self-sufficiency, etc. The following subsections define 173 the system under study, and describe the methodology and main data sources used in the 174 analysis. 175 176 3.1. Food system 177 178 Here we assimilate the food system to the part of the economy that is activated to produce the 179 food (including beverages) purchased directly or in restaurants and hotels by final consumers 180 such as households, governments, NGOs and similar organizations, as well as to produce the 181 energy products these final consumers use in food-related activities such as cooking and 182 refrigeration. 183 184 The activities involved in the production of food are not restricted to the agricultural sector, food 185 processing, packaging and distribution. They also cover the life-cycle stages of the inputs 186 required to support each of these activities (e.g. fertilizer and pesticide production, extraction of 187 raw materials, manufacturing industries, energy production and distribution, service industries, 188 etc.). This approach ensures that all the elements involved directly or indirectly in food 189 production for and consumption of final consumers are accounted for. Purchases made in other 190 food-related industries (e.g. hospitals, universities, schools, prisons, stadiums, cinemas, etc.) are 191 not included in this analysis. 192

3.2. Data sources

195	The energy profile of the food system can be assessed from two sides: production and
196	consumption. The production side shows the domestic energy supply or use associated with the
197	food system. The consumption side, on the other hand, depicts the upstream energy demand
198	related to food consumption activities, independent from where energy is used. The upstream
199	energy demand of consumption is commonly referred to as energy footprint. We use the term
200	'energy foodprint' to refer to the energy footprint of food systems.
201	
202	EEMRIO databases provide the means necessary to assess both the production and consumption
203	perspectives. Here we use EXIOBASE 3.6 (Stadler et al. 2018) as the core data source. The
204	monetary structure of EXIOBASE represents 200 product groups for 44 countries that account
205	for more than 90% of the world's GDP. The remaining countries are grouped in five 'rest of
206	world' regions. For ease of reporting, we aggregate countries, food products, energy users and
207	energy products as shown in Table 1. Details on the mapping of the EXIOBASE countries and
208	products classification to the groups represented in this paper are available in the supporting
209	material.

211 Table 1: Regions, food product, user and energy product groups used in this paper

Regions	Food products	Energy user	Energy products
Europe	Meat	Agriculture	Coal electricity
North America	Fish	Fishing	Gas electricity

Latin America	Dairy and other animal products	Other Primary	Oil electricity
Africa	Grains	Food Processing	Nuclear electricity
Middle East	Vegetables, fruits and nuts	Chemicals	Renewable electricity
High-income Asia and Pacific	Other	Other manufacturing	Biomass/waste electricity
Other Asia and Pacific		Electricity/heat	Heat
		Transport	Coal
		Services	Gas
		Households	Oil products
			Nuclear fuels
			Biomass/waste

In its current publicly available version (v3.4), EXIOBASE contains detailed industry- and product-specific energy accounts. The database includes primary energy accounts (supply) and gross energy accounts (supply and use) for around 60 energy products. In contrast to primary energy accounts, gross energy accounts represent certain energy flows twice (e.g. coal for electricity production and the electricity itself), which makes them inadequate for footprint calculations (Arto et al. 2016). The use of primary accounts as environmental extension avoids this double accounting problem.

Primary energy accounts can contain primary energy supply (PES) – domestic extraction of
energy – and net energy use (NEU) – end use of energy products (excluding exports) plus all
losses of energy – data. Each type of account is intended to address a different set of research
questions (Owen et al. 2017). For instance, energy footprints based on PES data are best suited to
shed light into the origin of the energy associated with final consumption activities, while
footprints based on NEU data are more appropriate to attribute the actual energy use to industry
sectors.

228

229 Because EXIOBASE only contains data on PES, we have generated NEU accounts to be used as 230 environmental extension following the guidance provided in the official energy accounting 231 manuals (Eurostat 2014; UN 2015a). This required the conversion of IEA extended energy 232 balances (IEA 2017b, 2017a), from the territory to the residence principle (see Usubiaga and 233 Acosta-Fernández (2015) for more details), filtering the net energy use data and allocating the 234 resulting energy use to the EXIOBASE product and industries following the allocation procedure 235 in Stadler et al. (2018). A more detailed explanation is available in the supporting information. 236 237 Given that our definition of the food system covers food-related activities that take place within 238 the household, we have also estimated the direct energy use required for cooking and 239 refrigeration within the household. To this end, we have extracted the product-specific 240 percentages of residential energy devoted to such activities from the TIMER model (Daioglou et 241 al. 2012) and incorporated it in the NEU extension as described in the supporting information.

242

3.3. Energy profile of the food system

We have computed production- and consumption-based accounts (footprints) for the whole economy and for the food system using both PES and NEU as environmental extensions. In the figures we use net energy use to refer to the production-based energy use, and energy foodprint for the consumption perspective. The mathematical formulation is the same irrespective of the extension used. In the equations below, bold lower case refers to vectors, bold upper case to matrices and italics to scalars. The dimensions of all the variables are given in the supporting information).

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253 Production-based accounts for the economy as a whole are given by the environmental extension. In the case of food systems (FS superscript), these are a function of the demand of food 254 255 by final consumers such as households, government, etc., and energy products used by 256 households in food-related activities such as cooking and refrigeration, which is shown in equation 1, where x represents output, L is the Leontief inverse, y^F the final demand of food, z^{F-R} 257 258 the direct input of food products associated with final consumers' purchases in restaurants and hotels, and $\mathbf{y}^{\mathbf{E}-\mathbf{F}}$ the final demand of energy products for cooking and refrigeration purposes. The 259 last two elements are calculated as shown in equations 2 and 3. In equation 2, A^F describes the 260 261 input coefficient matrix where the non-food input coefficients are converted to zero and y^{R} the 262 final demand of hotels and restaurants. Equation 3 shows the element-wise multiplication of the final demand of energy products (y^E) and the share of each product that is used for cooking and 263 264 refrigeration (w^{E-F}).

265

266
$$\mathbf{x}^{FS} = \mathbf{L} \left(\mathbf{y}^{F} + \mathbf{z}^{F-R} + \mathbf{y}^{E-F} \right)$$
(1)

- $z^{F-R} = \mathbf{A}^F \mathbf{y}^R \quad (2)$
- 269
- 270 $\mathbf{y}^{\mathbf{E}-\mathbf{F}} = \mathbf{y}^{\mathbf{E}} \circ \mathbf{w}^{\mathbf{E}-\mathbf{F}}$ (3)
- 271

In equation 4, **D_prod** and **S** represent production-based accounts and the stressor (primary energy supply or net energy use) intensity respectively. The element **fh**^{FS} refers to the direct food-related energy use of households in physical terms. This is a positive value when using net energy use as extension and equals 0 when using primary energy supply, for the extraction of primary energy products is not undertaken by final consumers.

- 277
- 278

$$\mathbf{D}_{\mathbf{FS}} = \mathbf{S} \operatorname{diag}(\mathbf{x}^{\mathbf{FS}}) + \mathbf{fh}^{\mathbf{FS}} \quad (4)$$

279

The calculation of the energy footprint of country *i* (equation 5) and of its food consumption (equation 6) is carried out using the standard formula for EEIOA, where **D_cons** denotes the energy foodprint and **fh**^{FS} the direct food-related energy use of households. This last item is 0 when using PES as extension.

284

285
$$\mathbf{D}_{\mathbf{cons}_{i}} = \mathbf{S} \mathbf{L} \operatorname{diag}(\mathbf{y}_{i}) + \mathbf{fh}_{i} \qquad (5)$$

286

287
$$\mathbf{D}_{\mathbf{cons}_{i}^{FS}} = \mathbf{S} \mathbf{L} \operatorname{diag}(\mathbf{y}_{i}^{F} + \mathbf{z}_{i}^{F-R} + \mathbf{y}_{i}^{E-F}) + \mathbf{fh}_{i}^{FS} \quad (6)$$

We have also compared the import dependency (i_dep) of different energy products for the food system and the whole economy. The equation below shows the dependency for the economy where j and k refer to energy products and industries respectively. The import dependency of the food system is calculated the same way using the **D_prod^{FS}** and **D_cons^{FS}** matrices instead. In this case, the production- and consumption-based indicators use PES as extension.

294

295
$$i_{dep_{i}} = 100 * \frac{\sum_{j,k} \mathbf{D}_{-} \mathbf{cons_{i}} - \sum_{j,k} \mathbf{D}_{-} \mathbf{prod_{i}}}{\sum_{j,k} \mathbf{D}_{-} \mathbf{prod_{i}}}$$
(7)

296

297 **4. RESULTS**

298

299 Overall, the food system accounts for approximately 13% of the global net energy use (dropping 300 from over 15% in the early 2000s, see Figure S1). Between 2000 and 2015, absolute net energy 301 use in food systems has increased by 14% approximately (Figure 1). Small absolute reductions 302 are seen in high-income regions (Europe, high-income Asian and Pacific countries (APAC)) with 303 the exception of North America. Larger absolute increases are seen across predominantly 304 middle- and lower-income regions. The trends in middle- and lower-income regions are partially 305 explained by population growth, as absolute increases are much higher than those in per capita 306 terms. Some middle- and lower-income regions actually show reductions in per capita terms (e.g. 307 Africa).

309 Figure 1: Overview of the absolute (a) and per capita (b) net energy use in the global food







The large per capita energy intensity gap between North America (mainly the United States) and the rest of the world has narrowed slightly over the period. While this fell for many countries from 2000 to 2015, it fell more rapidly in North America (see Figure 1). However, North American energy inputs into the food system are almost double than the next closest highincome region, in this case Europe.

From a footprint perspective, the demand of grains for human consumption drives the largest energy inputs in all regions (see Figure 2). This statement should be interpreted carefully though, for although the 'grains' category includes grains and grain-based processed products, the latter often covers processed products with many ingredients such as meat, vegetables, vegetable oils and sugars that belong to other categories, but could be allocated to them (see related limitations in and full product correspondence in the supporting information). This estimate does not include the as much as 36% of all grains in some regions that are directed to livestock rearing

326 (Cassidy et al. 2013). The energy inputs required to produce the grains fed to livestock are 327 embodied in the corresponding category of animal products. The energy inputs for all animal 328 products (meat, fish, dairy and other products) is roughly equivalent to the energy inputs for 329 grains produced for human consumption in some regions such as Europe. In total, between 23% 330 and 31% of all net energy use from European, North American and high-income Asian and 331 Pacific countries' foodprint is linked to animal products. Although, the per capita figures of 332 related to animal products are far from those in high-income countries, their relevance is 333 increasing over time in Latin America, the Middle East and other APAC countries. In this line, 334 most of the foodprint associated with food purchases is driven by consumption within the 335 household, although purchases in restaurants and hotels are not negligible in most high-income 336 regions (Figure S2). Direct energy use for refrigeration and cooking varies widely from region to 337 region, comprising as little as 11% in high-income Asia Pacific nations, to 88% in Africa. As 338 might be expected, higher income nations generally have a more efficient use of direct energy in 339 food supply, as driven by developed electricity grids and improvements in refrigeration and 340 cooking technologies. Because higher income countries tend to use electricity and natural gas 341 within the household, the indirect energy required to produce - especially the former - is higher 342 than in low- and middle-income countries and can represent an important share of its foodprint.

344 Figure 2: Breakdown of per capita energy footprint driven by the purchase of different food





346 given in GJ/cap.

348 Note: The direct energy represents energy required in cooking and refrigeration. Indirect energy 349 use refers to the energy used in the production of the food-related energy products consumed 350 within the household. 'Grains' include, among others, grains and grain-based products such as 351 bread and pasta whether or not cooked or stuffed, as well as other products such as biscuits, 352 pastries and cakes. 'Other' includes sugar products, beverages, oil seeds and other vegetable fats 353 (all plant-based products).

354

Splitting further between where the energy is used in the food chain, direct energy used by the household for food preparation and storage is significant as also shown in Figure 3, even in higher-income nations, varying from 13 to 16% of the total energy used in the food system across North America, Europe and high-income Asia Pacific nations to as much as 55% and 89% in other APAC nations and Africa respectively. In both cases, the domination of in-house energy use is due to inefficient cooking methods and lack of electrification in rural areas (see Figure 3). Other Asia Pacific nations, and the Middle East have seen significant reductions in the energy use by households since 2000. The proportion of energy use used in food processing and in primary cultivation or livestock rearing is similar in most regions, with slightly more energy use in processing within high-income APAC and Latin American nations. Chemical use in the food chain, including those for plastics and fertilizers are larger in the Middle East and high-income APAC nations and has grown larger over time.

367

Figure 3: Net energy use for the food system within different sectors across regions (%), 2000
and 2015 (D_prod^{FS}). The total per-capita figures on the top are given in EJ.



370

371 Note: The chemicals sector includes energy use for both fertilizers and plastics. Households
372 refers to direct use of energy for food use in the home. Services includes construction and non-

373 transport services such as financial services, education, waste management and real state.

374

376 Fuel use in the global food system is dominated by the use of fossil fuels and biomass (see 377 Figure 4). Fossil fuels include their end use (e.g. combustion of diesel in machinery, but not as 378 input of oil in a refinery) and the losses incurred in transformation, transport, etc. A maximum of 379 21% of energy in the food system comes from electricity. Higher-income nations tend to have a 380 lower biomass to fossil fuel ratio, with middle- and lower-income nations the reverse. Between 381 65% to 87% of net energy use is derived from fossil fuels across higher-income nations, with the 382 Middle East reaching 95%. The relative lack of electricity in the food system as compared to 383 other systems highlights the decarbonization challenge for energy in the food sector. There is 384 some growth of renewable energy as a proportion in some regions, with the largest proportional 385 increase in Europe. There are also large proportional increases in biomass for Europe and North 386 America, likely driven by increasing interest in, and expansion of biofuels. Across high-income 387 nations, oil makes up between 31% (Europe) and 43% (high-income APAC nations) of total 388 energy use in the food system. The large amount of direct oil use (i.e. not converted into 389 electricity) in the food sector highlights the challenge for the renewable transition within the food 390 sector.

392 Figure 4: Different types of net energy use in the food system across regions, 2000 and 2015



393 (D_prod^{FS}).



398

399 The large dependency on oil and gas also highlights potential issues of energy security within the 400 food system. Given the high regional dependency on these resources, disruptions to energy 401 supply may influence food systems. Figure 5 shows the difference in energy dependency 402 (modelled through equation 7 above) between energy used in the food system compared to the 403 whole of the economy. European countries see an increased import dependency on all fuels in 404 the food sector (when compared to the rest of the economy) except for renewable energy. In 405 total, Europe sees roughly a 50% higher dependency in the food system than the overall 406 economy. North American dependencies are more mixed, with its large endowments of coal and 407 shale oil/gas reducing dependency. Latin America shows a similar dependency for coal and gas

as Europe, but with less reliance on overseas oil. In Africa, the largest difference between the
food system and the whole economy is seen in renewables and nuclear. The middle east shows
expected trends in domestic supply of oil and gas and import dependency on all others. Highincome APAC countries show heavy energy dependency on imports for all fuels except for coal.
Across all fuels except coal, the food system is more dependent than the rest of the economy.

It should be noted however that trade interdependency may be larger than this picture shows due to the different grades of fuels within energy types. For example, US imports and processes large amounts of heavy crude oil, but exports large amounts of light crude oil produced domestically. These two grades are not easily fungible in the energy system so grouping by energy type can sometimes underestimate the underlying trade in fuels.

420 Figure 5: The import dependency of different energy types as used in the food system and the





423 Note: This figure does not include the category 'biomass and waste' (see limitations in

424 supporting information). The reader should note that individual energy carriers may represent

425 varying portions of absolute energy demand (see Figure 4).

426

427 **5. DISCUSSION**

428

429 With global population expected to be close to 10 billion people in 2050 (UN 2017), a zero 430 hunger goal will inevitably require more food to be produced in the future. The key to ensuring 431 that food production can be reconciled with the biophysical limits of the planet will be to 432 decouple food production from the inputs of natural resources as much as possible. These natural 433 resources include energy, and its associated environmental impacts, most notably climate 434 change. Energy use patterns vary widely both across regions with different and similar income 435 levels and there is variation in both the proportion of energy used in different food production 436 and consumption stages and the types of energy used. Because of this, measures for improving 437 efficiency and facilitating the low-carbon energy transition should be adapted to each 438 geographical context.

439

Across all high-income nations there is the large opportunity to reduce food waste, particularly at the point of consumption (Gustavsson et al. 2011). Such efforts have large upstream benefits. For instance, halving consumer food waste could potentially reduce the environmental foodprint of Europe by 10-11% on average (Usubiaga et al. 2018). In less industrialized countries, most food is lost in the production, processing, storage and transportation stages before it reaches the

445 consumer (Gustavsson et al. 2011), which also offers substantial possibilities to increase the446 efficiency of the system.

447

448 Across lower-income regions, the reduction of direct energy use while concurrently improving 449 refrigeration and reducing food losses in the production chain are key options. Low efficiency 450 cooking and heating using traditional biomass leads to large energy use in regions that rely on it, 451 leading to the counterintuitive result that per-capita food-energy use in Europe and Africa are 452 closer than expected (Figure 1). Latin America, the Middle East and other APAC nations also 453 have some reliance on traditional biomass but all except Africa have seen significant progress 454 from 2000 to 2015 in reducing those energy inputs. While the deployment of clean cookstove 455 efforts has been partially successful (Rosenthal et al. 2017) a prominent lesson from these efforts 456 is the need to ensure that solutions are location specific so that options address local differences 457 in cookware (i.e. flat or curved pans), cooking habits (i.e. appropriate for local dishes and 458 cultures), yet are still scalable at the same time (Diehl et al. 2019). A further important factor at 459 the African household level may be the continuing reduction in solar energy costs and potential 460 for electric refrigeration (N'Tsoukpoe et al. 2014). Improving the diffusion of clean refrigeration 461 and cooking technologies is a key task in achieving numerous SDGs relating to poverty, health, 462 gender equality, and maintenance of environmental services (Oparaocha and Dutta 2011; Rao 463 and Pachauri 2017; Fuso Nerini et al. 2018).

464

465 Significant embodied dependencies in the food system have been found for virtual water and
466 other resources (Dalin et al. 2012), but to the best of our knowledge there has been no estimate of
467 embodied energy in global food trade. Similar to water security issues driven through trade,

468 energy security through trade has important implications. We find that the European food system 469 has a higher exposure to imported energy embodied in food than the rest of the economy across 470 all energy types except for renewables. This suggests there may be an underappreciated food 471 supply risk benefit in further decarbonization of the European energy system. That is, further 472 development of European renewables may improve food supply security as well as energy 473 security (although there are concerns about the material requirements for renewable energy and 474 their potential supply risks). The interplay between energy and food security is one that is 475 relatively understudied, with nexus studies often focusing on water as the coordinating resource 476 (Lawford et al. 2013).

477

478 There is an urgent need for a low-carbon energy transition across all regions and sectors. 479 Progress to decarbonize the electricity system, although insufficient, has been much faster than 480 other energy sectors (Davis et al. 2018). The food system commonly lags behind in the 481 penetration of renewables compared to the rest of the economy (Monforti-Ferrario et al. 2015) 482 because the use of energy tends to be more diffuse than in other sectors with a particular focus on 483 transport and heating fuels. With new renewable energy installations in many countries at, or 484 cheaper than, the price of existing fossil fuel generation (McKinsey 2019) we can assume that 485 electricity in the food system (Figure 4) can be made renewable with relative ease at low cost. 486 Much harder is the use of oil in food production and transportation. Despite the fact that 487 transport currently makes up a small amount of the total energy used in food systems it may 488 become a dominant proportion for food-system emissions as the rest of the energy services 489 decarbonize. There is increasing innovation in electrifying farming equipment (Monforti-Ferrario 490 et al. 2015), but the electrification of long-distance transport still poses a significant challenge,

491 especially in shipping (Davis et al. 2018). Potential exists to reduce the energy use and emissions
492 of long distance freight through modal shift and fuel switching in freight, but the required
493 changes in logistics and infrastructure are not expected to be widespread in the short term
494 (McCollum et al. 2009; Kaack et al. 2018).

495

496 Monforti-Ferrario et al. (2015) have also documented options to increase the amount of 497 renewable energy in the production of ammonia and hydrogen for fertilizer production. The ban 498 of single-use plastics – included in the chemical sector in Figure 3 –, which includes unnecessary 499 packaging in the food industry also offers benefits. Irrigation practices also represent an 500 interesting example, where switching from open channel flow delivery systems to pressurized 501 networks can lead to significant water savings at the expense of higher energy use (Díaz et al. 502 2009). In these cases, important energy savings can be achieved by optimizing the operation of 503 the pumping station (Díaz et al. 2009; Lamaddalena and Khila 2012).

504

505 Changes in dietary trends and consumer behavior – especially in high-income nations – and can 506 also offer large benefits in several environmental aspects (Behrens et al. 2017) – including 507 energy – and health systems (Willett et al. 2019). Benefits are not only linked to changes in the 508 dietary mix, but also to reducing total food intake in some regions (Alexander et al. 2017). 509 Changing consumption and a focus on local and seasonal food products are likely to significantly 510 reduce the demand for freight transport, refrigeration, and fertilizers. Furthermore, diets with 511 reduced meat consumption limit the demand of, and consequent emissions from, land-use and 512 have been highlighted as a potentially important aspect of climate change mitigation pathways 513 (van Vuuren et al. 2018).

515

516 sector and putting in place taxes that help food products reflect their true environmental cost and 517 nutritional value are options that could go a long way, but seem harder to implement. 518 519 6. CONCLUSIONS 520 521 So far, most analyses of the energy profile of the food systems have a national or product-level 522 focus. This paper uses a global EEMRIO database to characterize the energy profiles of regional 523 food systems around the globe and their evolution between the years 2000 and 2015. By using a 524 single database, the analysis, which shows energy users, drivers and energy types across world

Economic instruments such as removing environmentally harmful subsidies in the agricultural

regions, is carried out in a comparable manner, which is something missing in the literature todate.

527

528 Overall, the ratio between energy use in the food system and the economy is slowly decreasing. 529 Current trends also point to a relative decoupling between net energy use (up 12% between 2000 530 and 2013) and food production (up 23% in the same period, (FAO 2017)). There is a myriad of 531 factors affecting this effect, including changes in population, diets, yields, electrification, energy 532 efficiency, food waste, access to food, etc., which make it very difficult to disentangle the main 533 drivers behind this phenomenon. The magnitude of the decoupling effect can be influenced by 534 the fact that our energy figures do not consider some of the food consumed outside the household 535 (e.g. cinemas, hospitals, canteens, etc.).

536

537 The energy profiles of world regions vary widely in terms of the energy types used, the energy 538 users, the food products driving consumption, or the dependency of imported energy in the food 539 system. This diversity in profiles arises from the variety of energy services demanded in the food 540 sector ranging across processes (production, processing, storage) and relevant actors (producers, 541 distributors, consumers). The difference in how food is produced, prepared and consumed across 542 the world suggests that interventions will need to prioritize different parts of the system 543 depending on the location. This implies that the solutions for reducing energy use, increasing 544 energy efficiency and decarbonizing energy supply in the food system will be substantially 545 different across regions and has an intimate interaction with the rest of the economy.

546

547 As a general observation, in high income regions energy use in food production is spread over 548 the supply chain, with production, processing, manufacturing and household energy use all 549 contributing significant amounts of energy demand. This demand is satisfied mainly through the 550 use of significant volumes of oil products (transport, packaging), as well as electricity and gas 551 (processing, cooking). In lower income regions, the nature of food production and distribution, 552 as well as the use of inefficient cooking fuels (i.e. traditional biomass) leads to energy use being 553 concentrated at the 'end-use'. Thus, electrification and access to cookstoves is key in reducing 554 biomass use for cooking in less industrialized countries. Thus, in higher income regions broader 555 strategies are required concerning electricity production, freight transport and heating 556 technologies.

557

558 Technological solutions will need to be complemented with changes in consumer behavior,

specially in industrialized countries with carbon-intensive diets and high food waste figures.

560	Interventions should, in any case, consider the effects beyond the immediate effects on energy
561	use, for changes in how food is produced and consumed can have spillovers and positive
562	synergies with respect to the social and environmental dimensions outlined in the SDGs.
563	
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- 570 **REFERENCES**
- 571
- Alcántara, V. and R. Duarte. 2004. Comparison of energy intensities in European Union
 countries. Results of a structural decomposition analysis. *Energy Policy* 32(2): 177-189.
 Aleksandrowicz, L., R. Green, E. J. M. Joy, P. Smith, and A. Haines. 2016. The Impacts of
- Aleksandrowicz, E., K. Green, E. J. M. Joy, T. Shifti, and A. Hanles. 2010. The impacts of
 Dietary Change on Greenhouse Gas Emissions, Land Use, Water Use, and Health: A
 Systematic Review. *PLoS One* 11(11): e0165797.
- Alexander, P., C. Brown, A. Arneth, J. Finnigan, D. Moran, and M. D. A. Rounsevell. 2017.
 Losses, inefficiencies and waste in the global food system. *Agricultural Systems* 153:
 190-200.
- Arto, I., I. Capellán-Pérez, R. Lago, G. Bueno, and R. Bermejo. 2016. The energy requirements
 of a developed world. *Energy for Sustainable Development* 33(Supplement C): 1-13.
- Ayres, R. U. 1995. Life cycle analysis: A critique. *Resources, Conservation and Recycling* 14(3):
 199-223.
- Beckman, J., A. Borchers, and C. A. Jones. 2013. Agriculture's Supply and Demand for Energy
 and Energy Products. (112): 35-35.
- Behrens, P., J. C. Kiefte-de Jong, T. Bosker, J. F. D. Rodrigues, A. de Koning, and A. Tukker.
 2017. Evaluating the environmental impacts of dietary recommendations. *Proceedings of the National Academy of Sciences* 114(51): 13412-13417.
- Bekhet, H. A. and A. Abdullah. 2010. Energy Use in Agriculture Sector: Input-Output Analysis.
 International Business Research 3(3): 111-121.
- Canning, P., A. Charles, S. Huang, K. R. Polenske, and A. Waters. 2010. Energy Use in the U.S.
 Food System. *Energy*(94).
- Cao, S., G. Xie, and L. Zhen. 2010. Total embodied energy requirements and its decomposition
 in China's agricultural sector. *Ecological Economics* 69(7): 1396-1404.
- 595 Cassidy, E. S., P. C. West, J. S. Gerber, and J. A. Foley. 2013. Redefining agricultural yields :
 596 from tonnes to people nourished per hectare.
- 597 Coley, D. A., E. Goodliffe, R. Jennie Macdiarmid, and J. Macdiarmid. 1998. The embodied
 598 energy of food: the role of diet. *Energy Policy* 26(6): 455-459.
- Daioglou, V., B. J. van Ruijven, and D. P. van Vuuren. 2012. Model projections for household
 energy use in developing countries. *Energy* 37(1): 601-615.
- Dalin, C., M. Konar, N. Hanasaki, A. Rinaldo, and I. Rodriguez-Iturbe. 2012. Evolution of the
 global virtual water trade network. *Proceedings of the National Academy of Sciences* 109(16): 5989-5994.
- Davis, S. J., N. S. Lewis, M. Shaner, S. Aggarwal, D. Arent, I. L. Azevedo, S. M. Benson, T.
 Bradley, J. Brouwer, Y.-M. Chiang, C. T. M. Clack, A. Cohen, S. Doig, J. Edmonds, P.
- Fennell, C. B. Field, B. Hannegan, B.-M. Hodge, M. I. Hoffert, E. Ingersoll, P. Jaramillo,
 K. S. Lackner, K. J. Mach, M. Mastrandrea, J. Ogden, P. F. Peterson, D. L. Sanchez, D.
- Sperling, J. Stagner, J. E. Trancik, C.-J. Yang, and K. Caldeira. 2018. Net-zero emissions
 energy systems. *Science* 360(6396): eaas9793.
- de Haes, H. A. U. 2004. Life-Cycle Assessment and Developing Countries. *Journal of Industrial Ecology* 8(1-2): 8-10.

- Díaz, J. A. R., R. L. Luque, M. T. C. Cobo, P. Montesinos, and E. C. Poyato. 2009. Exploring
 energy saving scenarios for on-demand pressurised irrigation networks. *Biosystems Engineering* 104(4): 552-561.
- Diehl, J. C., S. Van Sprang, J. Alexander, and W. Kersten. 2019. A Scalable Clean Cooking
 Stove Matching the Cooking Habits of Ghana and Uganda. Paper presented at GHTC
 2018 IEEE Global Humanitarian Technology Conference, Proceedings, 2019.
- Eshel, G. and P. A. Martin. 2006. Diet, Energy, and Global Warming. *Earth Interactions* 10(9).
- Eurostat. 2014. *Physical Energy Flow Accounts (PEFA) Manual 2014. draft version 15 May* 2014. Luxembourg: Eurostat.
- 621 FAO. 2017. *Food Balance Sheets*. Food and Agriculture Organization of the United Nations.
- 622 Foster, C., K. Green, M. Bleda, P. Dewick, B. Evans, A. Flynn, and J. Mylan. 2006.
- *Environmental Impacts of Food Production and Consumption*. London: Department for
 Environment, Food and Rural Affairs.
- Fuso Nerini, F., J. Tomei, L. S. To, I. Bisaga, P. Parikh, M. Black, A. Borrion, C. Spataru, V.
 Castán Broto, G. Anandarajah, B. Milligan, and Y. Mulugetta. 2018. Mapping synergies and trade-offs between energy and the Sustainable Development Goals. *Nature Energy* 3(1): 10-15.
- González, A. D., B. Frostell, and A. Carlsson-kanyama. 2011. Protein efficiency per unit energy
 and per unit greenhouse gas emissions : Potential contribution of diet choices to climate
 change mitigation. *Food Policy* 36: 562-570.
- Gustavsson, J., C. Cederberg, U. Sonesson, R. v. Otterdijk, and A. Meybeck. 2011. *Global food losses and food waste Extent, causes and prevention*. Rome: Food and Agriculture
 Organization of the United Nations.
- Hatirli, S. A., B. Ozkan, and C. Fert. 2005. An econometric analysis of energy input–output in
 Turkish agriculture. *Renewable and Sustainable Energy Reviews* 9(6): 608-623.
- Hauwermeiren, A. V., H. Coene, G. Engelen, and E. Mathijs. 2007. Energy Lifecycle Inputs in
 Food Systems : A Comparison of Local versus Mainstream Cases. *Journal of Environmental Poicy and Planning* 9:1(December 2014): 37-41.
- EA. 2017a. Energy Balances of non-OECD Countries (2016 edition). Paris: International
 Energy Agency.
- 642 IEA. 2017b. Energy Balances of OECD Countries (2016 edition). Paris: International Energy
 643 Agency.
- Kaack, L. H., P. Vaishnav, M. G. Morgan, I. L. Azevedo, and S. Rai. 2018. Decarbonizing
 intraregional freight systems with a focus on modal shift. *Environmental Research Letters* 13(8): 083001.
- Kander, A., M. Jiborn, D. D. Moran, and T. O. Wiedmann. 2015. National greenhouse-gas
 accounting for effective climate policy on international trade. *Nature Climate Change* 5:
 431.
- Kizilaslan, H. 2009. Input–output energy analysis of cherries production in Tokat Province of
 Turkey. *Applied Energy* 86(7): 1354-1358.
- Lamaddalena, N. and S. Khila. 2012. Energy saving with variable speed pumps in on-demand
 irrigation systems. *Irrigation Science* 30(2): 157-166.
- Laso, J., H. Daniel, M. Margallo, I. Garcia-Herrero, L. Batlle-Bayer, A. Bala, P. Fullana-i Palmer, I. Vazquez-Row, A. Irabien, and R. Aldaco. 2018. Assessing Energy and

656 Environmental Efficiency of the Spanish Agri-Food System Using the LCA/DEA 657 Methodology. energies 11: 2295-2295. 658 Lawford, R., J. Bogardi, S. Marx, S. Jain, C. P. Wostl, K. Knüppe, C. Ringler, F. Lansigan, and 659 F. Meza. 2013. Basin perspectives on the Water-Energy-Food Security Nexus. McCollum, D., G. Gould, and D. Greene. 2009. Greenhouse Gas Emissions from Aviation and 660 661 Marine Transportation: Mitigation Potential and Policies. Solutions White Paper Series. 662 McKinsey. 2019. Global Energy Perspective 2019 : Reference Case. McKinsey. Energy Insights. 663 Molina-besch, K., F. Wikström, and H. Williams. 2019. The environmental impact of packaging 664 in food supply chains — does life cycle assessment of food provide the full picture ? The 665 International Journal of Life Cycle Assessment 24: 37-50. 666 Monforti-Ferrario, F., J.-F. Dallemand, I. P. Pascua, V. Motola, M. Banja, N. Scarlat, H. 667 Medarac, L. Castellazzi, N. Labanca, P. Bertoldi, D. Pennington, M. Goralczyk, E. M. 668 Schau, E. Saouter, S. Sala, B. Notarnicola, G. Tassielli, and P. Renzulli. 2015. Energy use 669 in the EU food sector: State of play and opportunities for improvement. Luxembourg: 670 Joint Research Centre - Institute for Energy and Transport and Institute for Environment 671 and Sustainability. 672 Mozaffarian, D., S. Y. Angell, T. Lang, and J. A. Rivera. 2018. Role of government policy in nutrition—barriers to and opportunities for healthier eating. BMJ 361: k2426. 673 674 N'Tsoukpoe, K. E., D. Yamegueu, and J. Bassole. 2014. Solar sorption refrigeration in Africa. 675 Oparaocha, S. and S. Dutta. 2011. Gender and energy for sustainable development. Current 676 Opinion in Environmental Sustainability 3(4): 265-271. 677 Owen, A., P. Brockway, L. Brand-Correa, L. Bunse, M. Sakai, and J. Barrett. 2017. Energy 678 consumption-based accounts: A comparison of results using different energy extension 679 vectors. Applied Energy 190(Supplement C): 464-473. 680 Ozkan, B., A. Kurklu, and H. Akcaoz. 2004a. An input-output energy analysis in greenhouse 681 vegetable production: a case study for Antalya region of Turkey. *Biomass and Bioenergy* 682 26(1): 89-95. 683 Ozkan, B., H. Akcaoz, and C. Fert. 2004b. Energy input-output analysis in Turkish agriculture. 684 Renewable Energy 29(1): 39-51. 685 Pelletier, N., N. Arsenault, and P. Tyedmers. 2008. Scenario modeling potential eco-efficiency 686 gains from a transition to organic agriculture: Life cycle perspectives on Canadian 687 canola, corn, soy, and wheat production. Environmental Management 42(6): 989-1001. Pelletier, N., E. Audsley, S. Brodt, T. Garnett, P. Henriksson, A. Kendall, K. J. Kramer, D. 688 689 Murphy, T. Nemecek, and M. Troell. 2011. Energy Intensity of Agriculture and Food 690 Systems. Annu. Rev. Environ. Resour 36: 223-246. 691 Pimentel, D., S. Williamson, C. E. Alexander, O. Gonzalez-Pagan, C. Kontak, and S. E. Mulkey. 2008. Reducing energy inputs in the US food system. Human Ecology 36(4): 459-471. 692 693 Popkin, B. M. 2006. Global nutrition dynamics: the world is shifting rapidly toward a diet linked 694 with noncommunicable diseases. The American Journal of Clinical Nutrition 84(2): 289-695 298. 696 Pretty, J. N., A. S. Ball, T. Lang, and J. I. L. Morison. 2005. Farm costs and food miles : An 697 assessment of the full cost of the UK weekly food basket. Food Policy 30: 1-19. 698 Rao, N. D. and S. Pachauri. 2017. Energy access and living standards: some observations on 699 recent trends. Environmental Research Letters 12(2): 025011.

- Reynolds, C. J., J. Piantadosi, J. D. Buckley, P. Weinstein, and J. Boland. 2015. Evaluation of
 the environmental impact of weekly food consumption in different socio-economic
 households in Australia using environmentally extended input–output analysis.
 Ecological Economics 111: 58-64.
- Rosenthal, J., K. Balakrishnan, N. Bruce, D. Chambers, J. Graham, D. Jack, L. Kline, O. Masera,
 S. Mehta, I. R. Mercado, G. Neta, S. Pattanayak, E. Puzzolo, H. Petach, A. Punturieri, A.
 Rubinstein, M. Sage, R. Sturke, A. Shankar, K. Sherr, K. Smith, and G. Yadama. 2017.
 Implementation science to accelerate clean cooking for public health. *Environmental Health Perspectives*.
- Sanjuán, N., F. Stoessel, and S. Hellweg. 2014. Closing Data Gaps for LCA of Food Products:
 Estimating the Energy Demand of Food Processing. *Environmental Science and Technology* 48: 1132-1140.
- Sherwood, J., R. Clabeaux, and M. Carbajales-Dale. 2017. An extended environmental input–
 output lifecycle assessment model to study the urban food–energy–water nexus.
 Environmental Research Letters 12(10): 105003-105003.
- Song, F., T. Reardon, X. Tian, and C. Lin. 2019. The energy implication of China's food system
 transformation. *Applied Energy* 240: 617-629.
- Stadler, K., R. Wood, T. Bulavskaya, C.-J. Södersten, M. Simas, S. Schmidt, A. Usubiaga, J.
 Acosta-Fernández, J. Kuenen, M. Bruckner, S. Giljum, S. Lutter, S. Merciai, J. H.
 Schmidt, M. C. Theurl, C. Plutzar, T. Kastner, N. Eisenmenger, K.-H. Erb, A. d. Koning,
 and A. Tukker. 2018. Developing a time series of detailed Environmentally Extended
 Multi-Regional Input-Output tables. *Journal of Industrial Ecology* 22(3): 502-515.
- Tilman, D. and M. Clark. 2014. Global diets link environmental sustainability and human health.
 Nature 515: 518.
- UN. 2015a. System of Environmental-Economic Accounting for Energy. SEEA-Energy. Final
 Draft. United Nations Statistics Division and United Nations Department of Economic
 and Social Affairs.
- UN. 2015b. *Transforming our world: the 2030 Agenda for Sustainable Development* United
 Nations General Assembly. Resolution adopted by the General Assembly.
- 729 UN. 2017. 2017 Revision of World Population Prospects. United Nations.
- Usubiaga, A. and J. Acosta-Fernández. 2015. Carbon Emission Accounting in MRIO Models:
 The Territory vs. the Residence Principle. *Economic Systems Research* 27(4): 458-477.
- Usubiaga, A., I. Butnar, and P. Schepelmann. 2018. Wasting food, wasting resources: Potential
 environmental savings through food waste reductions. *Journal of Industrial Ecology* 22(3): 574-584.
- van Vuuren, D. P., E. Stehfest, D. E. H. J. Gernaat, M. van den Berg, D. L. Bijl, H. S. de Boer,
 V. Daioglou, J. C. Doelman, O. Y. Edelenbosch, M. Harmsen, A. F. Hof, and M. A. E.
 van Sluisveld. 2018. Alternative pathways to the 1.5 °C target reduce the need for
 negative emission technologies. *Nature Climate Change* 8(5): 391-397.
- Ward, H., L. Wenz, J. C. Steckel, and J. C. Minx. 2017. Truncation Error Estimates in Process
 Life Cycle Assessment Using Input-Output Analysis. *Journal of Industrial Ecology* 00(0).
- Willett, W., J. Rockström, B. Loken, M. Springmann, T. Lang, S. Vermeulen, T. Garnett, D.
 Tilman, F. DeClerck, A. Wood, M. Jonell, M. Clark, L. J. Gordon, J. Fanzo, C. Hawkes,
 R. Zurayk, J. A. Rivera, W. De Vries, L. Majele Sibanda, A. Afshin, A. Chaudhary, M.

745	Herrero, R. Agustina, F. Branca, A. Lartey, S. Fan, B. Crona, E. Fox, V. Bignet, M.
746	Troell, T. Lindahl, S. Singh, S. E. Cornell, K. Srinath Reddy, S. Narain, S. Nishtar, and
747	C. J. L. Murray. 2019. Food in the Anthropocene: the EAT-Lancet Commission on
748	healthy diets from sustainable food systems. The Lancet 393(10170): 447-492.
749	Woods, J., A. Williams, J. K. Hughes, M. Black, and R. Murphy. 2010. Energy and the food
750	system. Philosophical Transactions of the Royal Society B: Biological Sciences
751	365(1554): 2991-3006.
752	Yang, Y. and R. Heijungs. 2016. LCI METHODOLOGY AND DATABASES A generalized
753	computational structure for regional life-cycle assessment. The International Journal of
754	Life Cycle Assessment.
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757	SUPPORTING INFORMATION

- 758
- 759 The Word version of the supporting information provides an additional figure, an extended

760 description of the net energy use accounts generated and a description of the limitations of the

- results. The Excel file contains the correspondence tables used and the data represented in each
- 762 figure.