Spatially explicit global hotspots driving China's mercury related health impacts

3	Yumeng Li ^{1,#} , Long Chen ^{2,#} , Sai Liang ^{3,1,*} , Jianchuan Qi ^{3,1} , Haifeng Zhou ¹ ,
4	Cuiyang Feng ¹ , Xuechun Yang ¹ , Xiaohui Wu ¹ , Zhifu Mi ⁴ , Zhifeng Yang ^{3,5}
5	¹ State Key Joint Laboratory of Environment Simulation and Pollution Control,
6	School of Environment, Beijing Normal University, Beijing, 100875, P. R. China
7	² Key Laboratory of Geographic Information Science (Ministry of Education), School
8	of Geographic Sciences, East China Normal University, Shanghai, 200241, P. R.
9	China
10	³ Key Laboratory for City Cluster Environmental Safety and Green Development of
11	the Ministry of Education, Institute of Environmental and Ecological Engineering,
12	Guangdong University of Technology, Guangzhou, Guangdong, 510006, P. R. China
13	⁴ The Bartlett School of Construction and Project Management, University College
14	London, London, WC1E 7HB, UK
15	⁵ Guangdong Provincial Key Laboratory of Water Quality Improvement and
16	Ecological Restoration for Watersheds, Institute of Environmental and Ecological
17	Engineering, Guangdong University of Technology, Guangzhou, 510006, People's

18 Republic of China

19 *#* These two authors contributed equally.

20 * Corresponding author: <u>liangsai@gdut.edu.cn</u> (Sai Liang).

21 ABSTRACT

22 Over 100 nations signed the Minamata Convention on Mercury to control the adverse 23 effects of mercury (Hg) emissions on human beings. A spatially explicit analysis is needed to identify the specific sources and distribution of Hg-related health impacts. 24 This study maps China's Hg-related health impacts and global supply chain drivers 25 26 (i.e., global final consumers and primary suppliers) at a high spatial resolution. Here we show significant spatial heterogeneity in hotspots of China's Hg-related health 27 impacts. Approximately 1% of the land area holds only 40% of the Chinese 28 29 population but nearly 70% of the fatal heart attack deaths in China. Moreover, 30 approximately 3% of the land area holds nearly 60% of the population but 70% of the intelligence quotient (IQ) decrements. The distribution of hotspots of China's 31 Hg-related health impacts and global supply chain drivers are influenced by various 32 factors including population, economy, transportation, resources, and dietary intake 33 habits. These spatially explicit hotspots can support more effective policies in various 34 35 stages of the global supply chains and more effective international cooperation to reduce Hg-related health impacts. This can facilitate the successful implementation of 36 37 the Minamata Convention on Mercury.

38

39 KEYWORDS

40 Mercury; health impacts; supply chain; input-output analysis; spatially explicit
41 analysis; trade; footprint; Minamata Convention

42 SYNOPSIS

43 Spatially explicit analyses for China's Hg-related health impacts and relevant global
44 economic drivers can facilitate the smooth implementation of the Minamata
45 Convention on Mercury.

46

47 GRAPHICAL ABSTRACT/ TOC



49

50 1. INTRODUCTION

Mercury (Hg) is a global pollutant that can be transported worldwide and is 51 difficult to be removed from the environment.^{$\frac{1}{2}$, $\frac{2}{2}$ One of its forms, methylmercury} 52 (MeHg), is highly toxic to the nervous, digestive, and immune systems of human 53 body, as well as the development of the child in utero.^{3, 4} Given the adverse impacts of 54 Hg on human beings, the World Health Organization (WHO) has listed Hg as one of 55 the top ten chemicals with primary public health concern.³ To control global 56 anthropogenic Hg emissions and their adverse effects on human beings, over 100 57 58 nations signed the Minamata Convention on Mercury (Minamata Convention)⁵. This

59

60

convention entered into force on 16 August 2017, which is the first global agreement to protect human health and the environment in the recent decade.⁶

61 In order to reduce the human health damage of global Hg pollution, lots of studies have been conducted to evaluate Hg-related health impacts via MeHg intake and 62 dose-response relationship of MeHg exposure.^{7, 8} Studies on prenatal MeHg exposure 63 64 indicate that MeHg intake can have severe impacts on neurodevelopment in children, thus affecting their intelligence quotient (IQ) and cognitive levels.⁹ In addition, MeHg 65 exposure is closely associated with the risks of cardiovascular diseases.¹⁰ Given the 66 67 long-distance transport of Hg, the assessment of Hg-related health impacts should be based on a complete Hg cycle (e.g., atmospheric Hg emissions¹¹⁻¹³, atmospheric Hg 68 transport and deposition^{14, 15}, MeHg bioaccumulation^{4, 16}, and Hg-related health risks¹⁷, 69 ¹⁸). Recently, studies have combined the economic systems with traditional 70 biogeochemical Hg cycle to identify the impacts of economic activities and 71 international trade on Hg emissions and related health impacts. For example, previous 72 studies have characterized the impacts of local Hg emissions on domestic human 73 74 health in the U.S and China.^{17, 19,20} In addition to local Hg emissions, remote economic activities can also influence local Hg emissions through global supply chains^{17,21,22} and 75 subsequently influence human health^{13,17,18}. For instance, our recent study has revealed 76 China's regional Hg-related health impacts driven by the economic activities of other 77 regions in China.¹⁷ In addition to domestic economic drivers, international economic 78 activities also influence domestic Hg-related health impacts via global supply chains. 79 This requires international cooperation which is an important component of the 80

*Minamata Convention*⁵. However, the global economic drivers (including global final
consumers and primary suppliers) of local Hg-related health impacts have not been
well characterized yet.

Moreover, Hg-related health impacts and economic activities are both highly 84 localized and have significant spatial heterogeneity.²³ Existing studies have nationally 85 86 or sub-nationally averaged out the spatial characteristics of Hg-related health impacts and economic activities.^{17, 19} However, they ignored the identification of Hg-related 87 health impacts and economic drivers at a high spatial resolution. It is important to 88 89 develop a highly resolved inventory to explicitly spatialize the supply chain hotspots 90 of Hg-related health impacts (including the occurrence of Hg-related health impacts, global final consumers, and global primary suppliers). This effort can more accurately 91 locate the hotspots for Hg control measures and effectively facilitate the 92 implementation of the Minamata Convention. 93

This study fulfils the above knowledge gaps, taking China as a representative case. 94 95 China is important in global biogeochemical Hg cycle due to its largest atmospheric Hg emissions and population in the world.²⁴ Atmospheric Hg deposition over China 96 has been proven to adversely affect human health through the intake of Hg-containing 97 foods (e.g., rice, fish, vegetables, and meat).^{25, 26} The MeHg intake has resulted in 98 approximately 0.14 points of per-foetus IQ decrements and 7,360 deaths from fatal 99 heart attacks in China in 2010.¹⁷ Moreover, China participates in intensive trade of 100 101 commodities with other nations and plays the role of world factory in global supply

102 chains.²⁷ Identifying the hotspots for China's Hg-related health impacts and global
103 supply chain drivers plays an important role in global Hg control.

To effectively identify the global supply chain hotspots driving China's 104 Hg-related health impacts at a high spatial resolution, this study constructs a model 105 for the Spatially Explicit Global Tracking of Hg-related Health Impacts (SEGTHI). 106 This model integrates global economic supply chains with China's biogeochemical Hg 107 cycle to track spatially explicit hotspots of global supply chains driving the spatially 108 explicit health impacts of China's Hg emissions. We map China's Hg-related health 109 110 impacts (including IQ decrements and deaths from fatal heart attacks) in 2010 caused by local production, global final demand, and global primary inputs at a high spatial 111 resolution. In this way, we can identify key hotspots of Hg-related health impacts and 112 global supply chain drivers at a finer scale than the national and sub-national levels. 113 These findings can facilitate distinct Hg control strategies and international 114 cooperation to reduce Hg-related health impacts more effectively. It is worth noting 115 that there is a time lag between MeHg intake and the response of heart attacks 116 (generally 6 years). Moreover, the data on population structure (i.e., the Sample 117 Survey Data of 1% of National Population²⁸) used to calculate the fatal heart attack 118 119 deaths are only updated to 2015. Consequently, we use the global MRIO data in 2010 and estimate China's Hg-related health impacts in 2010. 120

121 **2. METHODS AND DATA**

6

The SEGTHI model consists of three components, including (1) the global 122 environmentally extended multi-regional input-output (EE-MRIO) model, (2) the 123 124 biogeochemical Hg cycle, and (3) the spatially explicit mapping for Hg-related health impacts of China caused by domestic production, global final demand, and global 125 126 primary inputs. This study evaluates the roles of 30 Chinese provinces and 140 nations as primary suppliers, direct emitters, and final consumers for China's 127 Hg-related health impacts at a high spatial resolution. Tables S1 and S2 in the 128 Supporting Information (SI) present the list of Chinese provinces, world nations, and 129 130 aggregated regions. The framework of the SEGTHI model in this study is illustrated in Figure S1. Here we provide the simplified methods and the detailed methods and 131 data can be found in section S1 of the SI. 132

133

2.1 Global EE-MRIO model

We use the global EE-MRIO model to evaluate the roles of nations and Chinese 134 provinces as primary suppliers, direct emitters, and final consumers for atmospheric 135 Hg emissions in China. The global EE-MRIO model has been widely used to 136 investigate environmental issues related to interregional trade, such as resource 137 $extraction^{29}$ and $atmospheric emissions^{27,30,31}$. The global EE-MRIO model is 138 constructed with a global multi-regional input-output (MRIO) table and a satellite 139 account of environmental issues (i.e., atmospheric Hg emissions in China in this 140 study). This model is used to calculate production-based, consumption-based, and 141 income-based emissions of nations and Chinese provinces for atmospheric Hg 142 emissions in China. We use the Chinese atmospheric Hg emission inventory compiled 143

in our previous work¹⁴ to construct the satellite account of the global EE-MRIO model. 144 The atmospheric Hg emission inventory is compiled by multiplying energy uses or 145 product yields with corresponding emission factors.¹⁴ The global MRIO table is 146 constructed by Mi et al.³², which links the Chinese MRIO data to the Global Trade and 147 Analysis Project (GTAP) database. The base year in this study is set as 2010. 148 The production-based Hg emissions are direct geographic Hg emissions from the 149 physical emitters. They are the satellite account of the global EE-MRIO table. We 150 define the Hg emission intensity of each sector using Eq. (1). 151 152

153 The notation indicates production-based Hg emissions of sectors in province n. 154 The row vector q indicates the direct Hg emission intensity of each province sector, 155 which represents Hg emissions for unitary output of each province sector. The column 156 vector represents the total output of sectors in province n.

157 Consumption-based Hg emissions of nations and Chinese provinces are calculated 158 by the Leontief MRIO model, which regards the economy as demand-driven. 159 Consumption-based Hg emissions represent China's Hg emissions directly and 160 indirectly driven by the final demand of nations and Chinese provinces^{31, 33}, as shown 161 in Eq. (2).

162

163 The notation represents consumption-based Hg emissions of nation/province n.
164 The row vector q indicates the direct Hg emission intensity of each province sector.

165 The column vector represents the final demand of nation/province *n*. The matrix is 166 the direct input coefficient matrix, in which each element equals to direct input from 167 sector *i* to sector *j* divided by the total output of sector *j*. The matrix **I** is an identify 168 matrix. The matrix is the Leontief Inverse matrix, which describes both direct and 169 indirect inter-sectoral relationships of the outputs and unitary final demand³⁴.

Income-based Hg emissions of nations and Chinese provinces are calculated by
the Ghosh MRIO model, which regards the economy as supply-push. Income-based
Hg emissions represent China's Hg emissions directly and indirectly enabled by
primary inputs of nations and Chinese provinces^{35, 36}, as shown in Eq. (3).

174

175 The notations represents income-based Hg emissions of nation/province *n*. The row vector indicates the primary inputs of nation/province *n* (values for sectors of 176 nation/province *n* are non-zero and values for other nation/province sectors are zeros). 177 178 The column vector q' represents the transposition of the Hg emission intensity vector q. The matrix **B** is the direct output coefficient matrix, in which each element equals 179 to direct input from sector *i* to sector *j* divided by the total output of sector *i*. With an 180 identity matrix I, the matrix is the Ghosh Inverse matrix, which represents both 181 direct and indirect inter-sectoral relationships of the outputs and unitary primary 182 inputs.³⁴ 183

184 **2.2 Biogeochemical cycle of Hg**

9

The methods for modelling the biogeochemical cycle of Hg in this study refer to 185 those in our previous work^{1/2}. Here we give a brief overview of each component, while 186 specific methods are described in SI S1.2. We also provided the details on related 187 methods and data on our website (https://www.cgeed.net/cmstm-model). The 188 biogeochemical cycle of Hg consists of five components which occur over mainland 189 China and the coastal seas, including a Chinese anthropogenic Hg emission inventory, 190 an atmospheric transport model, changes in food MeHg resulting from atmospheric 191 deposition, a human intake inventory of MeHg, and human health impacts due to 192 193 MeHg intake. Meanwhile, two external components include Hg emissions from natural sources and foreign anthropogenic sources, and imports of food products from 194 foreign nations. 195

We use the Chinese atmospheric Hg emissions inventory complied in our previous 196 work as the production-based emission inventory.¹⁴ The production-based emissions 197 are distributed as point and non-point sources, which serve as the input of an 198 atmospheric transport model. Then, we use the GEOS-Chem chemical transport 199 model (version 9-02; http://geos-chem.org) to simulate atmospheric Hg cycle over 200 201 China. A nested simulation at a horizontal resolution of $1/2^{\circ} \times 2/3^{\circ}$ over China is 202 conducted in a boundary condition provided by a global simulation at a horizontal resolution of $4^{\circ} \times 5^{\circ}$. Changes in Hg deposition simulated by the model are used as a 203 measure of changes in Hg concentrations in Hg-containing food products, including 204 ten categories of food products (i.e., rice, wheat, beans, vegetables, pork, poultry, 205 milk, eggs, marine fish, and freshwater fish) referred from previous studies $\frac{25}{37}$. Next, 206

we compile an intake inventory at the provincial scale to present a map of MeHg intake for the population over China in the model. The data regarding the compilation of intake inventory of MeHg and the evaluation of human health impacts are from our previous study¹⁷. In addition to the intake inventory, the model involves the dose-response relationships between dietary intake of MeHg and Hg-related health impacts to evaluate the health impacts. The dose-response relationships referred to in previous studies are based on epidemiological studies.^{10,19,38}

214

2.3 Spatially explicit mapping

The spatially explicit mapping of Hg-related health impacts in China includes the 215 impacts caused by domestic production, global final demand, and global primary 216 inputs. The grid maps of Hg-related health impacts in China caused by domestic 217 production show the occurrence of foetal IQ decrements and deaths from fatal heart 218 attacks. For grid maps of Hg-related health impacts in China caused by global final 219 demand and primary inputs, data in each grid represent the foetal IO decrements and 220 221 deaths from fatal heart attacks occurring in China caused by the final demand and primary inputs in this grid. 222

For Hg-related health impacts in China caused by domestic production, we use the gridded food consumption expenditure data in China as the proxy to spatialize the foetal IQ decrements and deaths from fatal heart attacks in China. Due to the lack of basic gridded data, we multiply the per-capita food consumption expenditure data of China at the city level with gridded population to obtain the gridded data for food

consumption. We select China's gridded population in 2010 and 2015 at the 1km \times 228 1km grid scale as the basic grid data, which are from the Institute of Geographic 229 Sciences and Natural Resources Research, Chinese Academy of Sciences³⁹. The 230 gridded population in 2010 is used in the evaluation of the foetal IO decrements, and 231 the gridded population in 2015 is used to investigate Hg-related fatal heart attacks to 232 reflect the time lag between MeHg intake and the response of heart attacks.¹⁷ The 233 gridded population data are calibrated with the official demographic statistics^{40,28}. 234 Notably, when calculating the foetal IQ decrements, we multiply the birth rate in 2010 235 236 to the gridded population data to obtain the gridded data of new-born population. The 237 birth rate data can be obtained from the China Statistical Yearbook⁴⁰.

For grid maps of China's Hg-related health impacts driven by global final demand, 238 we construct a grid map for global final demand as the proxy to spatialize the health 239 impacts. The grid map for global final demand is constructed with the final demand of 240 nations and Chinese provinces in the global MRIO table and the global nighttime 241 lights data at the 1km \times 1km grid scale in 2010. The grid data for global nighttime 242 lights can be obtained from the Defense Meteorological Satellite Program's 243 Operational Linescan System (DMSP-OLS) Nighttime Lights Time Series⁴¹. They can 244 245 directly reflect the intensity of human activities and describe the socioeconomic development level of a region.⁴² We simulated the relationship between the global 246 final demand and the values of nighttime lights. We observed a significant positive 247 248 correlation relationship between them, and the fitting degree is high (SI Figure S5). Therefore, we assume that the spatial distribution of nighttime lights is consistent with 249

that of the final demand. Subsequently, we use the values of nighttime lights at the 1km \times 1km grid scale as the proxy variable to spatialize global final demand at the national/subnational scale. In this way, we can investigate global final consumers driving China's Hg-related health risks at a higher spatial resolution to support more targeted policy decisions.

The primary inputs are also called value added, whose sum in each region is equal to the Gross Domestic Product (GDP). The main variable that influences the primary inputs is GDP. Thus, for grid maps of Hg-related health impacts in China enabled by global primary inputs, we use the global gridded GDP data at the 1km \times 1km scale in 2010 developed by United Nations Environment Programme⁴³ as the proxy variable to spatialize the health impacts. See section S1.3 of the SI for specific methods.

261

3. RESULTS AND DISCUSSION

262 **3.1 Spatially explicit health impacts of atmospheric Hg emissions in China**

Overall, anthropogenic atmospheric Hg emissions in China led to 8.48×10^{-2} points 263 of per-foetus IQ decrements and 4,593 deaths from fatal heart attacks in 2010. These 264 impacts generally increase from northwestern inland areas to southeastern coastal 265 areas (Figure 1a-b). Figure 1c shows that atmospheric Hg emissions of Central China 266 lead to the most national health impacts, with 2.19×10^{-2} points of per-foetus IQ 267 decrements and 1,196 deaths from fatal heart attacks. In terms of sectoral sources, the 268 electricity and heat power sector is the major sectoral source of national Hg-related 269 270 health impacts in eastern regions (including the Northeast, Beijing-Tianjin Region, Central Coast, and South Coast), contributing approximately 40% in each region. The 271

metallurgy industry plays a dominant role in western regions, including the Northwest 272 and Southwest. In addition, for the North and Central, the nonmetal products, 273 274 metallurgy, and electricity and heat power sectors are all important sources of national Hg-related health impacts. Specially, a significant spatial heterogeneity was observed 275 276 in hotspots of China's Hg-related health impacts. Approximately 1% of the land area holds 40% of the Chinese population but nearly 70% of the deaths from fatal heart 277 attacks in China. Moreover, approximately 3% of the land area holds nearly 60% of 278 the Chinese population but 70% of the IQ decrements. For example, approximately 80% 279 280 of the hotspots in Sichuan Province are located in the Chengdu-Chongqing District (occupying only 32% of the land area in Sichuan) of eastern Sichuan. Few hotspots 281 have been observed in western Sichuan because western Sichuan is mountainous with 282 283 lower population and less industrial activities. This indicates the necessity of spatially explicit mapping to accurately identify the hotspots and effectively reduce Hg-related 284 health impacts in China. Moreover, it shows that the hotspots of economic activities 285 286 and Hg-related health impacts are driven by various factors, not just the population 287 density.

The spatial distribution of Hg-related health impacts is affected by many factors, including population, economy, resources, and dietary intake habits. Many hotspots of Hg-related health impacts are distributed in urban agglomerations. This pattern is the most obvious in the Yangtze River Delta (including Shanghai, South Jiangsu, North Zhejiang, and some regions in Anhui), Beijing-Tianjin-Hebei Region, Pearl River Delta (including Guangzhou, Foshan, Shenzhen, and Zhuhai in Guangdong Province),

and Chengdu-Chongqing District. For example, the land area of the Yangtze River 294 Delta and Pearl River Delta urban agglomerations accounts for only 4% of the whole 295 country, but holds approximately 30% of China's Hg-related health impacts. Most of 296 these cities are provincial capitals or municipalities, such as Beijing, Shanghai, 297 Nanjing, Hangzhou, Guangzhou, and Chengdu (see SI Figure S5a for provincial 298 boundaries). These cities have large populations and high levels of socioeconomic 299 development, inducing large amounts of anthropogenic Hg emissions. Meanwhile, 300 their populations consume large amounts of marine fish and meat with relatively high 301 Hg concentration. This subsequently results in high Hg-related health impacts in these 302 cities. 303



304

305 Figure 1. Spatially explicit human health impacts caused by anthropogenic atmospheric Hg emissions in China in 2010. Graphs a and b describe spatially 306 307 explicit deaths from fatal heart attacks due to methylmercury intake and total foetus IQ decrements in each grid. It is worth noting that graph b shows the total IQ 308 decrements for the total new-born population in each grid. Graph c shows national 309 per-foetus IQ decrements and deaths from fatal heart attacks caused by anthropogenic 310 atmospheric Hg emissions in each region of China. The pie chart shows the 311 contributions of sectors in each region to the national health impacts caused by 312 313 atmospheric Hg emissions in this region.

314

Moreover, parts of the hotspots of China's Hg-related health impacts are 315 316 associated with the distribution characteristics of mineral and water resources. China's mineral resources are mainly distributed in the North, Northwest, and Southwest,⁴⁴ 317 where many hotspots of Hg-related health impacts are located. Famous metal mining 318 319 areas include iron mines in Panzhihua of Sichuan, Hg mines in Tongren of Guizhou, and gold mines in Zhaoyuan of Shandong (SI Figure S5b). For fossil fuels, Shanxi, 320 Xinjiang, and Sichuan are the major producers of coal, oil, and natural gas, 321 322 respectively. These resource-abundant areas discharge large amounts of atmospheric Hg emissions due to intensive mining, smelting, and energy combustion activities. 323 This subsequently results in significant Hg-related health impacts. Figure 1c shows 324 that the *metallurgy* industry is an important source of national Hg-related health 325 impacts for these resource-based regions, including the Northwest (58%), Southwest 326 (52%), and North China (16%). In addition to mineral resources, Hg-related health 327 impact hotspots are also associated with water resources because water resources are 328 essential for industrial production.⁴⁵ For instance, many hotspots are located in areas 329

330 of the Yellow River catchment across the Gansu, Ningxia, and Shanxi provinces (SI 331 Figure S5b). Moreover, Hg-related health impacts in southeast coastal areas are 332 generally higher than other areas. This situation might be related to the higher 333 consumption of fish, because fish consumption is a major pathway of MeHg 334 exposure³⁷.

335

3.2 Spatially explicit global final consumers driving Hg-related health impacts

Domestic consumption (mainly in Central China, Central Coast China, and 336 Northwest China) contributes 82% of the Hg-related health impacts in China, while 337 foreign consumption (mainly in North America, Western Europe, Japan, India, and 338 the Middle East) contributes the remaining 18% of impacts (see SI Figure S6 for 339 details of domestic and foreign consumers). The U.S. is the largest foreign consumer, 340 contributing 3.2×10^{-3} points of per-foetus IQ decrements and 171 deaths from fatal 341 heart attacks. At the sectoral level, domestic demand of construction activities is the 342 largest source. In most foreign regions, the consumption of mechanical and electrical 343 products is the largest source (contributing over 20%). However, in the Middle East 344 and North Africa, the final demand of construction activities has the leading role 345 (contributing 25%). The demand for building materials (e.g., cement, steel, and 346 ceramic tiles) in the Middle East (especially Saudi Arabia) is highly dependent on 347 China.46 348

Figure 2 shows a significant spatial heterogeneity in global consumers driving
Hg-related health impacts in China. The consumption activities in approximately 0.1%

17

of the global land area holds nearly 10% of the global population but drives 70% of 351 the Hg-related health impacts in China. Overall, the domestic consumption in China 352 353 has driven the majority of China's Hg-related health impacts. Since domestic hotspots in China are much larger than other global hotspots, we have adopted another 354 355 classification to obtain more details about hotspots in China (SI Figure S7). The hotspots in China are mainly concentrated in North China, South Coast China, 356 Middle-Lower Yangtze Region, and Pearl River Delta. They are the most developed 357 regions in China with large populations and intensive consumption activities. They 358 359 drive large Hg-related health impacts in China through global supply chains. The details are shown in the section S2 of the SI for the consumption of specific Chinese 360 regions driving China's Hg-related health impacts. In addition, we observe significant 361 362 influences of transportation and resources on the distribution of hotspots, mainly in developing areas (e.g., Yunnan, Guizhou, Guangxi, Eastern Sichuan, and the Yellow 363 River Basin). These hotspots distribute along roads and rivers. 364



Figure 2. Spatially explicit global consumers driving China's Hg-related health
impacts in 2010. Abbreviations: NAM – North America; LAM – Latin America and
the Caribbean; WEU – Western Europe; MNA – the Middle East and North Africa;
PAS – Southeast Asia and the Pacific; POECD – Pacific-OECD-1990 nations
(including Japan, Australia, and New Zealand); EIT – Economies in Transition
(including Eastern Europe and the former Soviet Union); EAS – East Asia (China
excluded).

373

Foreign consumers driving China's Hg-related health impacts mainly include the U.S., Western Europe, Japan, and South Korea (See section S2 of the SI for more international pairs). They are developed nations and regions, where the hotspot locations are based on urban agglomerations centred on large cities. Cities are aggregation areas of consumption activities. They consume large amounts of commodities that directly and indirectly lead to atmospheric Hg emissions and related health impacts in China. For instance, in the eastern U.S., hotspots are concentrated in

the megalopolises of New York, Boston, Washington D.C., Atlanta, and Chicago, 381 forming a dotted distribution. The consumption of mechanical and electrical products 382 383 in these regions results in significant Hg-related health impacts in China through global supply chains (Figure 2). In Western Europe, the consumption of Netherlands, 384 Belgium, western Germany, and southern United Kingdom leads to the most 385 Hg-related health impacts in China. In addition, some regions in Southeast Asia (e.g., 386 Hanoi and Ho Chi Minh City in Vietnam, Bangkok in Thailand, Kuala Lumpur in 387 Malaysia, and Singapore) have close trade ties with China. Their consumption has 388 389 driven large Hg-related health impacts in China.

390 The transportation is also an important factor affecting the distribution of global consumers driving China's Hg-related health impacts. This distribution feature is 391 mainly manifested as a coastal distribution. The geographical location of coastal areas 392 facilitates international trade activities. For example, consumption hotspots on the 393 west and east coast of the U.S., the west coast of Latin America and Caribbean, and 394 395 the south coast of West Africa all follow the coastal distribution patterns. Another typical area of the transportation-based pattern is the Mid-Eastern of the U.S. There 396 397 are major cities working as freight transportation nodes, and the hotspots are radiantly distributed on the dominating transportation arteries. For example, the Ohio state 398 ranks second among all the U.S. states in terms of China's Hg-related health impacts 399 driven by the final demand of various states in the U.S. However, the total value of 400 Ohio's nighttime lights ranks only 7th. There are not too many nighttime lights in 401 Ohio, but China's Hg-related health impacts caused by its final demand are high. One 402

possible reason is the influence of transportation factors. As the third most developed 403 state in manufacturing in the U.S., Ohio plays an important role as freight 404 transportation hub.⁴⁷ For instance, the freight volume on the Ohio River is twice that 405 of the Panama Canal.⁴⁸ Another area worthy of attention is Minnesota (especially 406 Minneapolis and St. Paul). It ranks sixth among all the U.S. states in terms of China's 407 Hg-related health impacts driven by the final demand of various states in the U.S. 408 However, its population only ranks 21st among the U.S. states. Minnesota is adiacent 409 to the Great Lakes and Mississippi River. It also borders Canada. This makes it an 410 important port and rail transportation center.⁴⁸ These findings indicate that particular 411 areas responsible for high Hg-related health impacts are not necessarily big cities with 412 large populations or nighttime lights. Convenient traffic conditions can also make 413 414 final demand activities gather here, resulting in high Hg-related health impacts in China through global supply chains. 415

416 3.3 Spatially explicit global primary suppliers enabling Hg-related health 417 impacts

Figure 3 shows the spatially explicit global primary suppliers enabling China's Hg-related human health impacts in 2010. Domestic primary inputs (mainly in Central China, Northwest China, and North China) enable approximately 93% of the Hg-related health impacts in China, while foreign primary inputs (mainly in North America, Western Europe, and Japan) enable the remaining 7% of impacts (SI Figure S8). For relatively developed regions such as the Central Coast, major sectoral source enabling China's Hg-related health impacts is the *scientific research and other*

services sector. For the Northwest with abundant mineral resources, primary inputs in 425 the *metallurgy* sector takes the leading role. In the Central, primary inputs in the 426 427 nonmetal products sector contributes the most to Hg-related health impacts in China (SI Figure S3). Foreign primary suppliers enabling significant Hg-related health 428 impacts in China involve the Pacific-OECD-1990 nations (including Australia, Japan, 429 and New Zealand), Latin America and the Caribbean, and the Middle East and North 430 Africa. See section S2 of the SI for more international and intranational trade pairs 431 inducing China's Hg-related health impacts. In particular, Australia is the largest 432 primary supplier, enabling 1.6×10^{-3} points of IQ decrements and 85 deaths from fatal 433 heart attacks in China. The metal and nonmetal mining sector and petroleum and gas 434 sector are the major sectoral sources for foreign primary suppliers. The metal and 435 436 nonmetal mining sector contributes nearly 50% of China's Hg-related health impacts enabled by primary inputs of Pacific-OECD-1990 nations, as well as Latin America 437 and the Caribbean. The petroleum and gas sector contributes 76% for the Middle East 438 439 and North Africa. The reason is that China imports large amounts of mineral resources from Australia and Latin America, and oil and gas resources from the 440 Middle East.⁴⁹ 441

Figure 3 also shows a significant spatial heterogeneity in global primary suppliers enabling Hg-related health impacts in China. The distribution of global primary suppliers is significantly influenced by the distribution of resources, including mineral, oil, gas, and water resources. Similarly, we provide a spatially explicit map of domestic primary suppliers in China driving Hg-related health impacts (SI Figure S9).

In China, the resource-based distribution feature is obvious for hotspots in the Central 447 (e.g., Henan), North (e.g., Hebei, and Shanxi), and Northwest (e.g., Shaanxi, Gansu, 448 449 and Inner Mongolia). These regions have abundant fossil fuel and mineral resources (e.g., coal in Shanxi and Inner Mongolia, nonferrous metals in Henan and Hebei, and 450 oil in Xinjiang). They provide raw materials to downstream producers and enable 451 downstream Hg-related health impacts. There are also hotspots in the Yangtze River 452 and Yellow River catchments driven by the availability of water resources. Moreover, 453 hotspots in Jiangsu Province have a distinct distribution feature along the Yangtze 454 River and Taihu Lake. For the rest of the world, hotspots also have obvious 455 characteristics of resource-based distributions. For example, hotspots in the Middle 456 East are associated with the distribution of oil fields along the Persian Gulf. Southeast 457 458 Asia is abundant in minerals, including coal mines in Vietnam and large gold mines in Indonesia.⁵⁰ Thus, primary inputs in the *coal mining* and *metal and nonmetal mining* 459 sectors are major hotspot sources in Southeast Asia. The distribution of foreign 460 461 primary suppliers is also influenced by water resources (e.g., the Nile in Egypt, the Amazon in northern Latin America, and the Rhine in Germany). 462



Figure 3. Spatially explicit global primary suppliers enabling China's Hg-related
health impacts in 2010. Abbreviations: NAM – North America; LAM – Latin
America and the Caribbean; WEU – Western Europe; MNA – the Middle East and
North Africa; PAS – Southeast Asia and the Pacific; POECD – Pacific-OECD-1990
nations (including Japan, Australia, and New Zealand); EIT – Economies in
Transition (including Eastern Europe and the former Soviet Union); EAS – East Asia
(China excluded).

471

In addition, many hotspots of primary suppliers are distributed based on urban agglomerations, dominated by the service industries. The *leasing and business services* sector in Beijing-Tianjin Region, Chengdu-Chongqing District, North America and Western Europe provides leasing and business services to downstream production activities and enables large Hg-related health impacts in China through global supply chains. Moreover, hotspots in the east coast of Australia and the northeast and west coasts of Latin America form the transportation-oriented distribution. Hotspots in coastal areas provide transportation services to downstream
producers and enable large Hg-related health impacts in China through global supply
chains. This indicates the significant role of transportation in the distribution of
primary suppliers.

483 4. IMPLICATIONS

This study constructs a SEGTHI model, which integrates global economic supply 484 chains with China's biogeochemical Hg cycle. We map China's Hg-related health 485 impacts, associated global final consumers, and global primary suppliers at a high 486 spatial resolution. We observed significant spatial heterogeneity in hotspots of China's 487 Hg-related health impacts and global supply chain drivers. Results show that 488 approximately 1% of the land area holds 40% of the population but nearly 70% of the 489 490 fatal heart attack deaths in China. This indicates that the distribution of hotspots may be influenced by many factors, such as economy, transportation, resources, and 491 dietary intake habits, not just the population density. This finer-scale study can help 492 more accurately locate the hotspots of Hg-related health impacts and economic 493 activities. It can facilitate more targeted policy decisions for the implementation of the 494 Minamata Convention. 495

496 **Comparisons with previous studies.** We compared our results with previous studies. 497 Compared with the gridded Hg emissions in China in 2010⁵¹, we find that Hg 498 emissions in the southeast coastal areas are relatively low, but their Hg-related health 499 impacts are relatively high. This is related to the high consumption of marine fish in 500 coastal areas and also reflects the Hg outflow from inland areas to coastal areas 501 through atmospheric transport. Conversely, Hg emissions in Henan province are high, 502 but the Hg-related health impacts are not so serious. This may explain the role of 503 atmospheric Hg transport and may be related to the low-rice and high-wheat dietary 504 habit in Henan province. The consumption of rice is one of the most important 505 pathways to MeHg exposure.²⁵

Moreover, compared with our previous study, spatially explicit results help 506 identify the specific sources and distribution of global hotspots. They can guide more 507 508 accurate and effective measures to reduce Hg-related health impacts. Compared with 509 the sub-national results in our previous study, this study observed many areas where Hg-related health impacts are overestimated or underestimated (SI Figure S12). For 510 example, the total Hg-related health impacts in Beijing, Tianjin, and Shanghai are 511 relatively low. However, they have many spatially explicit hotspots of Hg-related 512 health impacts. In addition, the total Hg-related health impacts in Hebei and Shandong 513 are not much high. However, special attention should be paid to Shijiazhuang, 514 Tangshan, Jinan, and other industrial cities, because of their large Hg emissions and 515 516 high Hg-related health impacts. On the other hand, although the total Hg-related 517 health impacts in Jiangsu, Zhejiang, Guangdong, and Sichuan are relatively high, most of the hotspots in these regions are concentrated in limited areas including the 518 Yangtze River Delta, Pearl River Delta, and Chengdu-Chongqing District. The 519 520 spatially explicit findings of this study can support the successful implementation of the Minamata Convention from two aspects: providing spatially explicit hotspots for 521

522 policy decisions in various stages of global supply chains and for international 523 cooperation.

Spatially explicit policy implications. Three perspectives are widely recognized to 524 identify hotspots of global supply chains, comprising production-side, demand-side, 525 and supply-side perspectives. Production-side control measures are effective for Hg 526 emissions and related health impact hotspots, including improving energy use 527 efficiency, promoting low-Hg energy sources, and establishing Hg removal facilities.⁵² 528 Spatially explicit measures should focus on hotspots including power plants in the 529 530 Yangtze River Delta and Pearl River Delta, the *nonmetal products* sectors of urban agglomerations in the middle reaches of the Yangtze River (especially in Wuhan, 531 Changsha, and Nanchang), and the *metallurgy* sector in certain large mining areas. 532

533 Demand-side control measures, mainly aiming at the optimization of consumption behaviours through economic instruments, are effective in hotspot areas where the 534 final demand induces significant upstream health impacts through global supply 535 536 chains.⁵³ According to the results of this study, consumption behaviour optimization should focus on the construction sector in the urban agglomerations of 537 Beijing-Tianjin Region, Middle-Lower Yangtze River, and Pearl River Delta in China. 538 For foreign hotspots, relevant measures should be aimed at the eastern U.S. (mainly 539 New York, Boston, Washington D.C., and Chicago), Cologne and Munich in 540 Germany, and London and Manchester in the United Kingdom. Moreover, the 541 542 demand and productivity of construction activities in Saudi Arabia should be 543 concerned.

27

Supply-side measures, focusing on the optimization of primary input behaviours 544 such as the adjustment of tax rates and financial incentives on selling low-Hg 545 546 products, are effective in critical geographic and sectoral sources whose primary inputs enable large downstream health impacts.³⁵ Decision makers could limit loan 547 supply and subsidies to the *metal and nonmetal mining* and *metallurgy* sectors in the 548 coastal areas of Australia and Latin America (mainly in Brisbane, Melbourne, Rio de 549 Janeiro, and Santiago). For the Persian Gulf coast in the Middle East and the large oil 550 fields in Eastern Europe, authorities can choose to invest in dominant petroleum and 551 552 gas enterprises with less income-based Hg-related health impacts. In addition, primary inputs of the *leasing and commercial services* sector in Western Europe (especially 553 southwest Germany) enable large amounts of Hg-related health impacts in China. 554 555 These regions could encourage their leasing and commercial enterprises to preferentially serve enterprises with lower income-based Hg-related health impacts. 556

Different from previous studies, the findings of this study can identify spatially 557 558 explicit geographical and sectoral hotspots for the implementation of these three types of measures. Considering the spatial heterogeneity of hotspots, control strategies 559 560 should be conducted in regions where hotspots are relatively concentrated (e.g., developed urban agglomerations). Notably, almost all of the Hg-related health 561 impacts in Sichuan are concentrated in its eastern part and connected with Chongqing. 562 Thus, measures in this region can only focus on Chengdu, Chongqing, and their 563 surrounding areas. This can significantly improve the effectiveness of measures 564 controlling Hg-related health impacts across the whole province. Moreover, some 565

hotspots distribute along the transportation arteries and large rivers, which are not
limited to administrative boundaries. Thus, the measures should consider
cross-boundary cooperation and may potentially be coordinated with city,
transportation, and resource management.

International cooperation is an important component of the *Minamata Convention*⁵. 570 The results of this study provide scientific foundations for this component by 571 identifying spatially explicit international trade pairs from the demand and supply 572 perspectives (SI S2). From the demand perspective, measures should focus on the 573 574 international pair of east-central U.S. and coastal areas in China (mainly the Yangtze River Delta and Pearl River Delta). From the supply perspective, international 575 cooperation should focus on the interconnections between resource-exporting regions 576 and corresponding Chinese areas, such as coastal areas in Australia -577 Chengdu-Chongqing District and the Middle East – Yangtze River Delta. Potential 578 measures include transferring technologies and related capital from these global 579 580 consumers and primary suppliers to hotspots of Hg-related health impacts in China. For instance, dominant enterprises of the *metal and nonmetal mining* sector in the 581 582 coastal areas of southeastern Australia could transfer technologies and related capital to the Chengdu-Chongqing District in China to reduce Hg-related health impacts there. 583 In addition, intranational cooperation is critical for reducing the overall Hg-related 584 health impacts in China. Herein, we find that Northwest (mainly in the Yellow River 585 Basin) - Southwest (mainly in Chengdu-Chongqing District) is an important trade 586

pair leading to the most Hg-related health impacts in China. Intranational cooperation
between these two regions should be promoted by national governments.

Limitations and uncertainties. The methods of this study have certain limitations. 589 On one hand, due to the limitation of computing resources and the lack of 590 high-resolution meteorological data, the resolution of this SEGTHI model is limited 591 by the horizontal resolutions in the chemical transport model. To analyze China's 592 Hg-related health impacts and the economic drivers at a higher spatial resolution, we 593 spatialize these factors based on certain regional heterogeneity features (e.g., 594 595 population, socioeconomic status, food consumption, and income). Therefore, our 596 finer-scale model is a parametric model developed with multiple parameters. It can reflect the influence of multiple factors, not just population. Specifically, we use the 597 gridded population data at the 1km \times 1km scale and city-level dietary intake data to 598 spatialize the production-based Hg-related health impacts in China, reflecting the 599 influence of population density and dietary intake. In terms of China's Hg-related 600 601 health impacts driven by global final demand and primary inputs, we use gridded data of global nighttime lights and GDP at the $1 \text{km} \times 1 \text{km}$ scale as the proxy variables to 602 603 spatialize global final demand and primary inputs, respectively. This can reflect the 604 influence of socioeconomic factors and reveal the global economic drivers at a finer scale than the national/subnational level. This study can reveal the spatially explicit 605 features of global final consumers and primary suppliers (e.g., city-level, basin-based, 606 607 and cross-boundary features of global economic drivers). Our SEGTHI model is currently refined and improved at the output terminal of Hg-related health impacts, 608

while the front-end emission data and mid-end transport model are limited by the lack 609 of gridded data, computing resources, and the background of meteorological data. It is 610 611 difficult to unify the resolution of the whole model. Improving this limitation depends on the refinement of Hg emission data and meteorological data. On the other hand, at 612 present, there are debates about whether Hg-related heart attacks are fatal. Some 613 scholars believe that MeHg intake may lead to deaths from fatal heart attacks, while 614 other scholars hold objections. Giang and Selin¹⁹ calculated both the non-fatal heart 615 attacks and fatal heart attacks in the U.S., while Chen et al.¹⁷ evaluated the deaths 616 from fatal heart attacks in China. According to the data availability, this study uses the 617 deaths from fatal heart attacks to characterize Hg-related health impacts. 618

This study is subject to certain uncertainties. Uncertainties mainly originate from 619 the estimations of China's atmospheric Hg emissions, global MRIO data, chemical 620 transport model simulations, the intake inventory of MeHg, and the evaluation of 621 human health impacts. We conducted a Monte Carlo simulation with 10,000 samples 622 623 of these factors. The P10 and P90 values of the statistical distributions are set as the lower and upper limits of the uncertainty range. For example, ranges for the national 624 625 Hg-related health impacts of per-foetus IQ decrement points caused by atmospheric Hg emissions from Central China are $(0.39 \times 10^{-2}, 4.31 \times 10^{-2})$. For deaths from fatal 626 heart attacks, the uncertainties are estimated at (433, 2693). The uncertainties in the 627 points of per-foetus IQ decrements and deaths from fatal heart attacks driven by the 628 final demand of the U.S. are $(0.06 \times 10^{-2}, 0.61 \times 10^{-2})$ and (70, 374). From the supply 629 perspective, the uncertainties enabled by the primary inputs of Australia are 630

- $(0.03 \times 10^{-2}, 0.30 \times 10^{-2})$ and (35, 182). See SI S3 for more details of the uncertainties,
- 632 including the specific sources and related improvement measures.

633 AUTHOR CONTRIBUTIONS

- 634 Conceptualization, S.L.; Methodology, S.L., Y.L., L.C., and J.Q.; Investigation, Y.L.,
- 635 L.C., J.Q., and H.Z.; Writing and revisions, all authors; Supervision, S.L.

636 **COMPETING INTERESTS**

637 The authors declare no competing interests.

638 ACKNOWLEDGEMENTS

This study is financially supported by the National Natural Science Foundation ofChina (71874014 and 41701589).

641 SUPPORTING INFORMATION

- 642 Additional information on (1) the complete methods and data sources (S1); (2)
- 643 international and intranational trade inducing Hg-related health impacts in China (S2);
- 644 (3) the limitations and uncertainties of this study (S3); (4) supplementary data and
- results (S4, Figures S4-S12); (6) the list of aggregated Chinese and world regions, as
- 646 well as regions in the GATP 9 Database (Tables S1 and S2). This information is
- available free of charge via the Internet at http://pubs.acs.org.

648 **REFERENCES**

Driscoll, C. T.; Mason, R. P.; Chan, H. M.; Jacob, D. J.; Pirrone, N. Mercury as a
 global pollutant: sources, pathways, and effects. *Environ. Sci. Technol.* 2013, 47,
 4967-4983.

- Selin, N. E. Global biogeochemical cycling of mercury: a review. Ann. Rev. *Environ. Resources* 2009, 34, 43-63.
- 6543.WHO.Mercuryandhealth:Keyfacts;655https://www.who.int/en/news-room/fact-sheets/detail/mercury-and-health.
- 4. Lavoie, R. A.; Jardine, T. D.; Chumchal, M. M.; Kidd, K. A.; Campbell, L. M.
 Biomagnification of mercury in aquatic food webs: a worldwide meta-analysis. *Environ. Sci. Technol.* 2013, 47, 13385-13394.
- 5. UNEP. *Minamata Convention on Mercury*; United Nations Environment
 Programme: Geneve, Switzerland, 2013.
- 6. Lin, Y.; Wang, S. X.; Steindal, E. H.; Wang, Z. G.; Braaten, H. F. V.; Wu, Q. R.;
 62 Larssen, T. A Holistic Perspective Is Needed To Ensure Success of Minamata
 63 Convention on Mercury. *Environ. Sci. Technol.* 2017, *51*, 1070-1071.
- Axelrad, D. A.; Bellinger, D. C.; Ryan, L. M.; Woodruff, T. J. Dose-response
 relationship of prenatal mercury exposure and IQ: an integrative analysis of
 epidemiologic data. *Environ. Health Perspect.* 2007, *115*, 609-615.
- 8. Mahaffey, K. R.; Clickner, R. P.; Bodurow, C. C. Blood organic mercury and
 dietary mercury intake: National Health and Nutrition Examination Survey, 1999
 and 2000. *Environ. health perspect.* 2003, *112*, 562-570.
- Myers, G. J.; Davidson, P. W.; Cox, C.; Shamlaye, C. F.; Palumbo, D.;
 Cernichiari, E.; Sloane-Reeves, J.; Wilding, G. E.; Kost, J.; Huang, L.-S.;
 Clarkson, T. W. Prenatal methylmercury exposure from ocean fish consumption
 in the Seychelles child development study. *The Lancet* 2003, *361*, 1686-1692.
- Roman, H. A.; Walsh, T. L.; Coull, B. A.; Dewailly, É.; Guallar, E.; Hattis, D.;
 Mariën, K.; Schwartz, J.; Stern, A. H.; Virtanen, J. K.; Rice, G. Evaluation of the
 cardiovascular effects of methylmercury exposures: current evidence supports
 development of a dose-response function for regulatory benefits analysis. *Environ. Health Perspect.* 2011, *119*, 607-614.
- Pirrone, N.; Cinnirella, S.; Feng, X.; Finkelman, R. B.; Friedli, H. R.; Leaner, J.;
 Mason, R.; Mukherjee, A. B.; Stracher, G. B.; Streets, D. G.; Telmer, K. Global
 mercury emissions to the atmosphere from anthropogenic and natural sources. *Atmos. Chem. Phys.* 2010, *10*, 5951-5964.
- Pacyna, E. G.; Pacyna, J. M.; Sundseth, K.; Munthe, J.; Kindbom, K.; Wilson, S.;
 Steenhuisen, F.; Maxson, P. Global emission of mercury to the atmosphere from
 anthropogenic sources in 2005 and projections to 2020. *Atmos. Environ.* 2010,
 44, 2487-2499.
- 13. Streets, D. G.; Hao, J.; Wu, Y.; Jiang, J.; Chan, M.; Tian, H.; Feng, X.
 Anthropogenic mercury emissions in China. *Atmos. Environ.* 2005, *39*, 7789-7806.
- Chen, L.; Meng, J.; Liang, S.; Zhang, H.; Zhang, W.; Liu, M.; Tong, Y.; Wang,
 H.; Wang, W.; Wang, X.; Shu, J. Trade-induced atmospheric mercury deposition
 over China and implications for demand-side controls. *Environ. Sci. Technol.* **2018**, *52*, 2036-2045.
- Liu, M.; Chen, L.; Wang, X.; Zhang, W.; Tong, Y.; Ou, L.; Xie, H.; Shen, H.; Ye,
 X.; Deng, C.; Wang, H. Mercury Export from Mainland China to Adjacent Seas

and Its Influence on the Marine Mercury Balance. Environ. Sci. Technol. 2016, 696 50, 6224-6232. 697 16. Clayden, M. G.; Kidd, K. A.; Wyn, B.; Kirk, J. L.; Muir, D. C. G.; Odriscoll, N. 698 J. Mercury biomagnification through food webs Is affected by physical and 699 chemical characteristics of lakes. Environ. Sci. Technol. 2013, 47, 12047-12053. 700 701 17. Chen, L.; Liang, S.; Liu, M.; Yi, Y.; Mi, Z.; Zhang, Y.; Li, Y.; Qi, J.; Meng, J.; Tang, X.; Zhang, H.; Tong, Y.; Zhang, W.; Wang, X.; Shu, J.; Yang, Z. 702 Trans-provincial health impacts of atmospheric mercury emissions in China. Nat. 703 Commun. 2019, 10, 1484. 704 18. Sunderland, E. M.; Li, M.; Bullard, K. Decadal Changes in the Edible Supply of 705 Seafood and Methylmercury Exposure in the United States. Environ. health 706 perspect. 2018, 126, 029003. 707 708 19. Giang, A.; Selin, N. E. Benefits of mercury controls for the United States. Proc. Natl. Acad. Sci. U. S. A. 2016, 113, 286-291. 709 710 20. Wu, Q.; Li, G.; Wang, S.; Liu, K.; Hao, J. Mitigation options of atmospheric Hg emissions in China. Environ. Sci. Technol. 2018, 52, 12368-12375. 711 712 21. Hui, M.; Wu, Q.; Wang, S.; Liang, S.; Zhang, L.; Wang, F.; Lenzen, M.; Wang, 713 Y.; Xu, L.; Lin, Z.; Yang, H.; Lin, Y.; Larssen, T.; Xu, M.; Hao, J. Mercury flows in China and global drivers. Environ. Sci. Technol. 2017, 51, 222-231. 714 22. Qi, J.; Wang, Y. F.; Liang, S.; Li, Y.; Li, Y. M.; Feng, C. Y.; Xu, L. X.; Wang, S. 715 X.; Chen, L.; Wang, D. F.; Yang, Z. F. Primary suppliers driving atmospheric 716 mercury emissions through global supply chains. One Earth 2019, 1, 254-266. 717 23. Angot, H.; Hoffman, N.; Giang, A.; Thackray, C. P.; Hendricks, A. N.; Urban, 718 N.; Selin, N. E. Global and local impacts of delayed mercury mitigation efforts. 719 Environ. Sci. Technol. 2018, 52, 12968-12977. 720 24. UNEP. Global Mercury Assessment 2018. United Nations Environment 721 Programme: Geneva, Switzerland, 2019. 722 25. Zhang, H.; Feng, X.; Larssen, T.; Qiu, G.; Vogt, R. D. In inland China, rice, 723 724 rather than fish, is the major pathway for methylmercury exposure. Environ. Health Perspect. 2010, 118, 1183-1188. 725 26. Feng, X.; Li, P.; Qiu, G.; Wang, S.; Li, G.; Shang, L.; Meng, B.; Jiang, H.; Bai, 726 W.; Li, Z. Human exposure to methylmercury through rice intake in mercury 727 mining areas, Guizhou Province, China. Environ. Sci. Technol. 2008, 42, 728 729 326-332. 27. Liang, S.; Zhang, C.; Wang, Y.; Xu, M.; Liu, W. Virtual atmospheric mercury 730 emission network in China. Environ. Sci. Technol. 2014, 48, 2807-2815. 731 28. National Statistics Bureau of China, Sample Survey of 1% of population in China, 732 2015; China Statistics Press: Beijing, 2016. 733 29. Wiedmann, T. O.; Schandl, H.; Lenzen, M.; Moran, D.; Suh, S.; West, J.; 734 Kanemoto, K. The material footprint of nations. Proc. Natl. Acad. Sci. U. S. A. 735 736 **2015**, *112*, 6271-6276. 30. Wang, H.; Zhang, Y.; Zhao, H.; Lu, X.; Zhang, Y.; Zhu, W.; Nielsen, C. P.; Li, 737 X.; Zhang, Q.; Bi, J. Trade-driven relocation of air pollution and health impacts 738 in China. Nat. Commun. 2017, 8, 738-738. 739

- Liang, S.; Wang, Y.; Cinnirella, S.; Pirrone, N. Atmospheric mercury footprints
 of nations. *Environ. Sci. Technol.* 2015, *49*, 3566-3574.
- Mi, Z. F.; Meng, J.; Guan, D. B.; Shan, Y. L.; Song, M. L.; Wei, Y. M.; Liu, Z.;
 Hubacek, K. Chinese CO₂ emission flows have reversed since the global
 financial crisis. *Nat. Commun.* 2017, *8*, 1712.
- 745 33. Peters, G. P. From production-based to consumption-based national emission
 746 inventories. *Ecol. Econ.* 2008, 65, 13-23.
- Miller, R. E.; Blair, P. D. *Input-Output Analysis: Foundations and Extensions*. 2
 ed.; Cambridge Univ. Press: Cambridge, 2009.
- Ziang, S.; Qu, S.; Zhu, Z.; Guan, D.; Xu, M. Income-Based Greenhouse Gas
 Emissions of Nations. *Environ. Sci. Technol.* 2017, *51*, 346-355.
- 36. Marques, A.; Rodrigues, J. F. D.; Lenzen, M.; Domingos, T. Income-based
 environmental responsibility. *Ecol. Econ.* 2012, *84*, 57-65.
- 37. Liu, M.; Chen, L.; He, Y.; Baumann, Z.; Mason, R. P.; Shen, H.; Yu, C.; Zhang,
 W.; Zhang, Q.; Wang, X. Impacts of farmed fish consumption and food trade on
 methylmercury exposure in China. *Environ. Int.* 2018, *120*, 333-344.
- 756 38. Rice, G. E.; Hammitt, J. K.; Evans, J. S. A Probabilistic Characterization of the
 757 Health Benefits of Reducing Methyl Mercury Intake in the United States.
 758 *Environ. Sci. Technol.* 2010, 44, 5216-5224.
- 759 39. Xu, X. *lkm*lkm grid data collection of population in China*. 2017.
 <u>http://www.resdc.cn/DOI</u>.
- 40. National Statistics Bureau of China, *China Statistical Yearbook, 2010.* China
 Statistics Press: Beijing, 2011.
- 76341. NCEI. Version 4 DMSP-OLS Nighttime Lights. National Centers for764EnvironmentalInformation,2010.
- 765 <u>https://www.ngdc.noaa.gov/eog/dmsp/downloadV4composites.html</u>.
- 42. Henderson, J. V.; Storeygard, A.; Weil, D. N. A Bright Idea for Measuring
 Economic Growth. *The American Econ. Rev.* 2011, 101, 194-199.
- 43. UNEP/DEWA/GRID-Geneva. Gross Domestic Product 2010.
 https://preview.grid.unep.ch/index.php?preview=data&events=socec&evcat=1&l
 ang=eng.
- 44. Li, C.; Wang, A.; Chen, X.; Chen, Q.; Zhang, Y.; Li, Y. Regional distribution
 and sustainable development strategy of mineral resources in China. *Chin. Geogr. Sci.* 2013, *23*, 470-481.
- 45. Yu, C. China's water crisis needs more than words. *Nature* **2011**, *470*, 307.
- 46. Henders, S.; Persson, U. M.; Kastner, T. Trading forests: land-use change and
 carbon emissions embodied in production and exports of forest-risk commodities. *Environ. Res. Lett.* 2015, *10*, 125012.
- Taylor, G. S. Top 10 states for manufacturing 2019. *Global Trade*, Sep 2, 2019, https://www.globaltrademag.com/top-10-states-for-manufacturing-2019/.
- 48. USDT, *State Transportation Statistics* 2010. United States Department of
 Transportation. 2010, <u>https://www.bts.gov/ctp</u>.
- 49. General Administration of Customs of the People's Republic of China, *China Custons Statistical Yearbook*, 2010; China Customs Press: Beijing, 2011.

- 50. Pombhejara, V. n. 8 Natural Resources and Raw Materials in Southeast Asia. In *Asia and the New International Economic Order*; Jorge A. Lozoya and A.K.
 Bhattacharya, Eds.; Elsevier 1981; pp 164-173.
- 51. Zhang, L.; Wang, S.; Wang, L.; Wu, Y.; Duan, L.; Wu, Q.; Wang, F.; Yang, M.;
 Yang, H.; Hao, J.; Liu, X. Updated emission inventories for speciated
 atmospheric mercury from anthropogenic sources in China. *Environ. Sci. Technol.* 2015, 49, 3185-3194.
- 52. Wu, Q.; Wang, S.; Liu, K.; Li, G.; Hao, J. Emission-Limit-Oriented Strategy To
 Control Atmospheric Mercury Emissions in Coal-Fired Power Plants toward the
 Implementation of the Minamata Convention. *Environ. Sci. Technol.* 2018, *52*,
 11087-11093.
- 53. Liang, S.; Xu, M.; Liu, Z.; Suh, S.; Zhang, T. Socioeconomic drivers of mercury
 emissions in China from 1992 to 2007. *Environ. Sci. Technol.* 2013, 47,
 3234-3240.

798