# 1 Sharing tableware reduces waste generation, emissions and water

# 2 consumption in China's takeaway packaging waste dilemma

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30 China has a rapidly growing online food delivery and takeaway market, serving 406 million 31 customers with 10. 0 billion orders and generating 323 kt of tableware and packaging waste in 32 2018. Here we use a top-down approach with city-level takeaway-order data to explore the 33 packaging waste and life-cycle environmental impacts of takeaway industry in China. The ten 34 most wasteful cities, with just 7% of the population, in terms of per capita waste generation 35 produced 30% of the country's takeaway waste, 27%-34% of country's pollutant and 30% of 36 water consumption. We defined one paper-substitution and two sharing tableware scenarios 37 to simulate the environmental mitigation potentials. The results of scenario simulations find that sharing tableware could reduce up to 92% waste generation, more than two-thirds of 38 39 environmental emissions and water consumption. Such a mechanism provides a potential 40 solution to address the food packaging waste dilemma and a new strategy for promoting 41 sustainable and zero-waste lifestyle.

## 42 Introduction

43 The digital revolution and changing lifestyles are reshaping the takeaway industry <sup>1,2</sup>. In China, 44 online food delivery platforms such as Meituan, Ele.me and Baidu are undergoing rapid 45 development and traditional food shopping habits are changing with advances in e-commerce and mobile terminal technology <sup>3,4</sup>. It is estimated that users of online takeaway platforms in 46 47 China increased in number from 60 million in 2011 to 416 million in 2019<sup>5</sup>. China's online food 48 delivery and takeaway market value has experienced an estimated increase from 22 billion yuan in 2011 to 285 billion yuan in 2019<sup>5</sup>, and the proportion of online takeaway turnover in the 49 50 total catering industry in China increased from 1.4% in 2015 to 10.6% in 2018<sup>6</sup>.

51 The negative impacts of production and disposal of single-use plastic packaging on the

52 environment and human health are growing global concerns <sup>7-9</sup> and in China, the 20 million

53 takeaway orders placed per day across the three online food delivery platforms are associated

54 with the use of 7.3 billion single-use plastic tableware sets per year<sup>10</sup>. China is now the world's

largest plastic and waste producer, generating 60.4 million tonnes (Mt) of plastic products in

56 2018 <sup>11</sup> and an estimated 553 kilotonnes (kt) of municipal solid wastes (MSW) per day <sup>12</sup>.

57 Packaging accounts for one-third of MSW.

58 A number of initiatives in China have sought new solutions for MSW management and plastic

reduction, including the MSW sorting implementation plan jointly issued by National

60 Development and Reform Commission (NDRC) and Ministry of Housing and Urban-Rural

61 Development (MHURD) in March 2017<sup>13</sup>, the "zero-waste city" pilot program by General Office

62 of the State Council in January 2019<sup>14</sup>, and a national-wide single-use plastic ban by DNRC and

63 Ministry of Ecological Environment (MEE) in January 2020<sup>15</sup>. In terms of the priority areas of

64 plastic pollution such as from e-commerce and the takeaway industry, Shanghai Association of

65 Food Contact Materials has, for example, released three non-binding food packaging standards

to encourage replacement of plastic food containers and bags with paper bowls and bag, and

67 biodegradable sacks<sup>16-18</sup>. The standards were implemented on a trail basis by three online food

68 delivery platforms in three districts of Shanghai since June 2018<sup>19</sup>. Shanghai was the first pilot

cities to implement the national MSW sorting policy, and the first mandatory regulation on
domestic waste management in China has been acted upon in Shanghai on July 1<sup>st</sup>, 2019,
mentioning that restaurant and food delivery business could not provide single-use chopsticks

72 and cutlery, if not requested by consumers  $^{20}$ .

73 In terms of sustainable management strategies, a number of studies have focused on the 74 environmental impacts of food tableware or packaging (e.g. container <sup>21-28</sup>, cutlery <sup>28-30</sup>, and bag <sup>28,31,32</sup>) with different materials (e.g. petroleum-based polymers <sup>21-26,30-33</sup>, and bio-based 75 polymers <sup>21,24,27,29,30,32,34,35</sup>) and lifecycle processes. For example, within its lifespan a 76 77 Tupperware reusable food saver was shown to balance out the life cycle impacts of single-use plastic takeaway food containers made from aluminium or extruded polystyrene <sup>26</sup>. When life-78 79 cycle energy use and environmental emissions were compared between one-way and 80 returnable food packaging systems in the European context, reusable packaging systems offered 81 potential environmental and economic benefits over single-trip solutions <sup>36,37</sup>. Circular solutions 82 associated with innovative reuse models, such as reusable packaging can be effective alternatives in minimising negative externalities of plastic packaging <sup>38,39</sup>. 83

84 As the sharing economy has the potential to promote shifts in collective consumption 85 behaviour<sup>40</sup>, sharing tableware may effectively decrease single-use plastic packaging and 86 enhance sustainability of the takeaway industry. Here we quantify the takeaway packaging 87 waste and seven environmental indicators of China's takeaway industry. We use a top-down 88 approach that divides the national packaging consumption into 353 cities based on city-level 89 takeaway order data collected from Meituan, the largest Chinese online food delivery platform, 90 http://waimai.meituan.com. Mitigation scenarios, such as paper-substitution and tableware-91 sharing, are compared with the baseline scenario and we show that sharing tableware is a 92 potential solution to reduce takeaway packaging waste and a new strategy for promoting 93 sustainable and zero-waste lifestyles.

# 94 Results

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## 95 Waste generated by online takeaway orders

96 Chinese online food delivery and takeaway industry served 406 million customers with 10.0 billion orders <sup>41</sup>, and generated 323 kt of tableware and packaging waste (218 kt plastic waste) 97 98 in 2018 (visualized in Extended Data Figure 1), which is equal to three-fifths of China's overall 99 MSW generation per day, 13 days of MSW generation in Beijing and one month of MSW 100 generation in Dongguan (a city in Guangdong province) <sup>12</sup>. The national average per capita 101 takeaway waste generated is 0.24 kg per year, and that generated in cities is shown in Figure 102 **<u>1</u>Figure 1**. Wuxi (a city in Jiangsu province) has the largest per capita takeaway waste (1.46 kg 103 per year), 6 times higher than the national average, and 5.12 million times higher than that of 104 Diging (a city in Yunnan province).

105 Figure 1 Takeaway packaging waste generated in China, 2018. The colours show the annual per capita 106 waste generated by cities, and darker regions have higher waste. The takeaway packaging wastes are 107 estimated in a top-down approach that downscales the national packaging consumption into the city-108 level with takeaway order collected from Meituan online food delivery platform. Takeaway waste 109 generated in Chinese cities vary significantly notably, there is no takeaway restaurant information in the 110 Shennongjia region (in Hubei province), Tongchuan (in Shannxi province), Gannan Tibetan Autonomous 111 Prefecture (in Gansu province), Tibetan Autonomous Prefecture of Guoluo, Huangnan, Hainan, and Yushu 112 (in Qinghai province), Guyuan (in Ningxia province), and Atux (in Xinjiang province).

113 The ten most 'wasteful' cities, shown in <u>Figure 2</u>, produce 30% (97.5 kt) of the country's

- overall takeaway waste. As the largest packaging producer (21.8 kt), Shanghai ranked the
- seventh in per capital packaging waste (0.90 kg). Wuxi was the fifth packaging waste producer
- 116 (9.6 kt) but contributed the largest per capital packaging waste (1.46 kg), indicating that people
- in Wuxi prefer ordering more takeaway than other cities. Generally, cities on the east coast (e.g.
- 118 nine of the top ten cities) have a greater economy in takeaways and produce the highest

amount of waste per capita, followed by the cities in the central and western regions (e.g. all

- the bottom ten cities as ranked by waste generation in **Figure 2**Figure 2). Food containers,
- 121 chopsticks, and plastic bags make up 44%, 19%, and 17% of the total takeaway waste,
- 122 respectively.
- **Figure 2 Takeaway packaging waste generated per capital in Chinese cities.** Cities are ranked by per capital takeaway packaging waste after dividing city takeaway packaging wastes by the population.

125 The bar charts show the per capita takeaway packaging waste of top and bottom 10 cities, and 126 contribution of each tableware and packaging is shown in different colours.

# 127 Environmental impacts of online takeaway orders

128 China's online takeaway ordering produced 709 kt of CO<sub>2</sub>, 2.0 kt of SO<sub>2</sub>, 2.6 kt of NO<sub>x</sub>, 485 t of

129 PM<sub>2.5</sub>, 436 mg of dioxin, 2.8kt of COD, and consumed 2.5 million m<sup>3</sup> of water in 2018. Single-use

130 food container, plastic bag, and tissue have higher environmental impacts (85% on average)

131 compared with other tableware. Food containers are the largest contributor to CO<sub>2</sub> (57% of the

total CO<sub>2</sub>), SO<sub>2</sub> (52%), NO<sub>x</sub> (48%), PM<sub>2.5</sub> (48%), and dioxin (46%) emissions from tableware and

are responsible for the greatest river water consumption (47%) from tableware. Plastic bag is

the second-greatest contributor of emissions of CO<sub>2</sub> (25%), NOx (18%), PM<sub>2.5</sub> (39%) and dioxin

- 135 (17%). Napkin makes up the largest share of COD emission (59%) and the second-largest share
- of SO<sub>2</sub> emission (18%) and water consumption (20%). The results from tableware and life cycle
- processes are presented in <u>Table 1</u>. From a lifecycle process perspective, the production

138 of raw material and tableware contributes more than four-fifths of the whole life-cycle

environmental impacts (i.e. 96% of SO<sub>2</sub>, 92% of PM<sub>2.5</sub>, 89% of COD, and 80% of water).

140 Production of raw material is the major source of CO<sub>2</sub> emissions (59%), followed by incineration

141 (34%). Incineration accounts for the largest dioxin emission (62%). Transportation contributes

the least to environmental impacts (less than 13% except for NOx emission, which is 54%).

143 Table 1 Takeaway environmental impacts by tableware and life cycle processes in China, 2018. The 144 environmental impacts of the takeaway industry are the sum of life-cycle phases of eight types of 145 tableware and packaging. The environmental impact of each packaging is estimated by multiplying the 146 annual packaging consumption by the life-cycle emission factor. Six life-cycle phases including production 147 of raw material ("Material production"), transportation of raw materials to production sites, production 148 and packaging of tableware and packaging ("Tableware production"), distribution of tableware and 149 packaging products to suppliers, takeaway delivery to consumers, utilization of tableware, and final 150 disposal ("Incineration" and "Landfill") are considered, while the transportation of raw materials for 151 tableware production, tableware production for suppliers and takeaway delivery were aggregated into 152 "Transportation" phase. There is no additional environmental impact in the tableware utilization phase 153 under baseline scenario.

Indicator	CO <sub>2</sub>	SO <sub>2</sub>	NOx	PM2.5	Dioxin	COD	Water
Unit	kt	t	<u>∓t</u>	t	mg	t	$10^3 m^3$

By tableware

Food container	406.09	1,057.12	1,241.83	231.45	202.41	708.07	1,157.91
Spoon	62.33	166.47	165.96	10.88	22.44	37.65	141.92
Chopsticks	4.56	45.59	333.65	23.55	50.04	65.29	307.57
Toothpick	0.12	1.19	8.88	0.60	1.35	1.81	8.69
Napkin	24.93	354.89	267.49	21.80	55.43	1,627.69	493.13
Cutlery wrapper	35.13	62.08	93.39	7.26	17.97	22.38	98.53
Toothpick wrapper	1.58	6.60	13.56	0.35	14.27	19.64	71.26
Plastic bag	174.65	321.45	467.94	189.19	72.50	280.63	184.14
By life cycle process							
Material production	417.07	1,339.84	1,009.60	392.19	81.31	2,281.58	1,053.17
Transportation	3.39	58.76	1412.92	16.71	0.47	60.16	365.54
Tableware production	45.29	591.21	118.19	53.85	81.94	184.56	917.91
Incineration	243.14	23.09	50.26	21.94	268.86	85.83	119.65
Landfill	0.51	2.01	1.74	0.19	3.83	153.03	5.88
Total	709.39	2,015.39	2,592.71	484.88	436.42	2,763.15	2,463.16

154 There are large regional differences in the environmental impacts of the takeaway industry in 155 Chinese cities (see Supplementary Table 6 for each environmental impact). We find that 156 relatively few cities are responsible for a disproportionately large share of the total emissions 157 and water consumption. For example, the ten most 'wasteful' cities contribute 32% of the 158 county's CO<sub>2</sub> emissions and 30% of the county's water consumption from tableware packaging, 159 but have just 7% of the population (pollutant emissions can be found in Supplementary Table 6). 160 As the most developed regions in China, city clusters of Jing-Jin-Ji, Yangtze River Delta, and Pearl 161 River Delta owing approximately one-seventh of the country's cities, are responsible for 53% of 162 the country's CO<sub>2</sub> emissions and 48% of the county's water consumption from takeaway 163 packaging, and have 24% of the population. Rich and tourist cities have larger environmental 164 impacts from takeaway orders than others (see Extended Data Figure 2). See Extended Data 165 Figure 2(b) of top 10 cities in per capita  $CO_2$  emissions as an example. As popular tourist cities 166 Qinhuangdao in Hebei province (2.5 kg per capita), Kunming in Yunnan province (2.0 kg per 167 capita), Sanya in Hainan province (1.9 kg per capita) have large  $CO_2$  emissions from takeaway.

168 Figure 3 Life-cycle takeaway CO<sub>2</sub> emission and takeaway Engle's coefficient of China, 2018. The blue 169 dots represent the takeaway carbon emission per capita of the cities. The larger the dots are, the larger 170 the per capita  $CO_2$  emission estimated by dividing life-cycle  $CO_2$  emissions of eight takeaway packaging 171 by the population. City' colour show their takeaway Engle's coefficient (TEC), defined as the proportion spent on takeaway of the household expenses. Annual takeaway spending of the city is determined by 172 173 multiplying annual takeaway order volume with associated sale price. Darker red colours represent 174 higher proportions of income spent on takeaway. We examine the Pearson correlation coefficients 175 between the TEC and per capita  $CO_2$  emission in cities (0.817, p-value 0.000). There are strong

203 correlations between the variables at the 0.01 significance level (2-tailed), indicating that the per capita
 204 takeaway CO<sub>2</sub> emission is closely related to the TEC.

205 We define takeaway Engle's coefficient (TEC), as shown in Figure 3Figure 3, to further explore 206 the city-level takeaway spending and lifestyle differences. A higher TEC (darker red in Figure 207 **3Figure 3**) indicates proportionately greater spending on takeout. We find that tourist and rich 208 cities have larger TECs than others, indicating their residents are willing to pay more on 209 takeaway food. Among the top ten cities with high TECs, six are tourist cities, such as Liaoyang 210 (in Liaoning province), Behai (in Guangxi province), Sanya (in Hainan province), Kelamayi (in 211 Xinjiang province), Xiamen (in Fujian province), and Tongliao (in Inner Mongolia province). The 212 remaining four cities (Wuxi and Suzhou in Jiangsu province, Wuhu in Anhui province, and 213 Shenzhen in Guangdong province) are rich, coastal cities. The less-developed cities in the 214 western region (e.g. Loudi in Hunan province and Wuwei in Gansu province) have lower TECs. 215 The TEC of Wuxi is 0.88%, which is 5.2 times higher than the national average (0.17%) and 2640 216 times higher than that of Loudi, and the takeaway CO<sub>2</sub> emission of Wuxi is 4.01 kg/cap, which is 217 8 times higher than the national average (0.52 kg/cap) and 236,239 times higher than that of 218 Loudi. High-income cities in developed areas with high TECs contribute larger takeaway CO<sub>2</sub> 219 emission than do low-income cities, and these large cities face greater environmental burdens.

# **Tableware sharing to mitigate impacts of online takeaway orders**

With the fast-development of circular and sharing economy<sup>40,42</sup>, paper alternatives and reusable
tableware provide potential solutions to mitigate the environmental impact of the takeaway
industry in China. To evaluate the mitigation potentials of different management strategies for
the Chinese takeaway industry, we define two scenarios (see scenario design, Extended Data
Figure 1 and Supplementary Table 1 for more details):

(1) Paper-substitution scenario: a set of tableware that includes a polyethylene (PE)-coated
 kraft paper container; a kraft paper bag; single-use cutlery package, comprising a
 polypropylene (PP) spoon, a pair of wooden chopsticks, a wooden toothpick and its
 wrapper, napkin and a biaxially oriented polypropylene (BOPP) chopstick wrapper.

Formatte

Formatte

230 (2) Tableware-sharing scenario: a reusable and returnable tableware set that includes a silicone 231 container (Partita); a reusable high-density polyethylene (HDPE) non-woven bag; a cutlery 232 package (wrapped by napkin), comprising a reusable silicone spoon, a pair of reusable 233 wooden chopsticks, a recycled napkin and a wooden toothpick and its recycled wrapper. 234 Two different takeback mechanisms are considered, including centralized takeback 235 mechanism whereby all tableware will be collected by courier and hand-washed in the 236 restaurant separately, and decentralized takeback mechanism that assumes all the reusable 237 tableware are returned to collection points by consumers and machine-washed in central 238 cleaning stations.

239 Figure 4 and Figure 5 show the life-cycle environmental emissions and water consumption by 240 tableware and processes under different scenarios, and different scales are used side by side for 241 the same indicator. The paper-substitution measure can reduce plastic waste by 57% (183kt) 242 and  $CO_2$  emissions by 49% (365 kt), but it creates an additional 493 kt of paper waste, 243 corresponding to 1.5 times the waste generated in the baseline scenario. Since pulp and paper production is one of the most energy-intensive manufacturing sectors <sup>43</sup>, paper-substitution 244 245 produces 79% more NO<sub>x</sub>, 465% more dioxin, and 89% more COD emissions and consumes an 246 additional 41% of water.

Paper bags and paper food containers are the primary sources of CO<sub>2</sub> (62%), SO<sub>2</sub> (70%), NO<sub>x</sub>
(82%), PM<sub>2.5</sub> (87%), dioxin (93%), COD (66%) emissions, and water consumption (68%). Dioxins
are mainly by-products of industrial processes, especially chlorine bleaching of paper pulp,
production of raw material (e.g. kraft paper) is responsible for the largest share of the dioxin
emissions (58%). Raw material production contributes the most to the COD emissions (66%),
followed by landfill (17%) and tableware production (13%).

The results could be attributed to the fact that withstanding the same pressure and having the
same volume, the paper bag has more mass, about seven times more than the plastic bag.
Paper bag production consumes 1.1 times energy and four times the amount of water, leads to
14 times eutrophication of water bodies, and produces 2.7 times solid waste it takes to make

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- 257 plastic bags<sup>44</sup>. For those areas without formal waste collection and recycling systems, paper
- substitution is not the optimal option for addressing takeaway packaging waste dilemma.

259 Figure 4 Life-cycle takeaway environmental impacts (air) by tableware and packaging under scenarios. 260 These bar charts indicate the  $CO_2$  and four air-pollutant emissions by six life-cycle phases and eight 261 tableware and packaging under baseline (SC-baseline), paper-substitution (SC-paper), and two 262 tableware-sharing scenarios. "SC-Decentral washing" denotes sharing tableware collection with manual 263 washing in restaurants. "SC-Central washing" implies the decentral collection of sharing tableware with 264 machine washing. "Material prod" means production of raw material, and "Tableware prod" denotes 265 production of tableware and packaging. "Transportation" represents material transport to tableware 266 manufacturers, tableware transport to suppliers, and the food delivery to consumers. "Incineration" and 267 "Landfill" represent the end-of-life process for single-use tableware and packaging, and "recycle" shows the final disposal for reusable items. "Washing" and "Takeback" belong to the utilization of sharing 268 269 tableware phase, respectively indicating water, electricity, and detergent consumption during the 270 washing process, as well as transport from decentralized tableware collection points to central cleaning 271 centres and send back to restaurants. "Cutl. W." means the cutlery wrapper. "Toot. W." refers to the 272 toothpick wrapper. 273 Figure 5 Life-cycle takeaway environmental impacts (water) by tableware and packaging under 274 scenarios. These bar charts indicate COD emission and water consumption by six life-cycle phases and 275 eight tableware and packaging under baseline (SC-baseline), paper-substitution (SC-paper), and two 276 tableware-sharing scenarios. The abbreviation for scenarios and life-cycle phases are the same as Figure 277 4. 278 Tableware-sharing scenarios have stronger mitigation effects on environmental impacts, 279 reducing takeaway waste by 92% (295 kt including 217 kt plastic waste, 63 kilotons disposable 280 chopsticks, and 13 kt paper waste) and environmental impacts by more than two-third (97% of 281 CO<sub>2</sub>, 93% of SO<sub>2</sub>, 68% of NO<sub>x</sub>, 89% of PM<sub>2.5</sub>, 84% of dioxin, 95% of COD and 67% of water for 282 decentralized takeback) compared with the baseline scenario. The use of recycled napkins can 283 mitigate more than one-half of environmental impacts (i.e. 73% of CO<sub>2</sub>, 52% of SO<sub>2</sub>, 17% of NO<sub>x</sub>, 284 38% of PM<sub>2.5</sub>, 61% of dioxin, and 96% of COD for decentralized takeback) and 67% of water 285 consumption compared with the use of virgin napkins.

- 286 The production of material and tableware generates the largest environmental emissions (CO<sub>2</sub>,
- dioxin, COD), followed by transportation (including takeback logistics) and washing phase. For
- 288 SO<sub>2</sub>, NO<sub>x</sub> and PM<sub>2.5</sub> emissions and water consumption, transportation is the main contributor.
- 289 Life-cycle water consumption of a reusable tableware set is 21 times higher than that of one-
- 290 way tableware set (see Supplementary Table 9). The water consumption of reusable tableware

is only 30% of cumulative one-way tableware in a year period. There are similar tendencies for
 other indicators, indicating that reusable tableware has resource-saving benefit and
 environmental mitigation potential.

294 The decentralized collection scenario has larger  $SO_2$ ,  $NO_X$ , COD emissions than centralized 295 takeback due to the extra impacts of takeback logistics. Takeback transportation contributes 4% 296 of CO<sub>2</sub> emissions, less than 16% of air pollutant emissions (SO<sub>2</sub>, PM<sub>2.5</sub>, dioxin) and water 297 consumption, and 21% of COD emissions, but contribute the largest  $NO_x$  emissions (75%). 298 Compared with centralized collection with manual washing, the decentralized collection with 299 machine washing can save another 31,617 kWh of electricity, 2000 m<sup>3</sup> water, and 1.4 kt 300 detergent, corresponding to reducing more than one third of environmental impact of washing 301 process (i.e. 34% of CO<sub>2</sub>, SO<sub>2</sub>, NO<sub>x</sub>, PM<sub>2.5</sub>, and 35% dioxin, COD and water).

## 302 Discussions and policy implications

To deal with the problem of takeaway packaging waste in China, policy-makers need specific information on the environmental impacts of the takeaway industry. We develop a top-down approach to estimate the takeaway waste generation and the life-cycle environmental impacts in China with city-level meal-ordering data from Meituan. The potential environmental impacts of different management strategies are indicated that tableware sharing is an effective and sustainable way to lessen the environmental impact of the takeaway industry.

309 Results of the sensitivity analysis demonstrated that life-cycle inventory datasets from different

β10 geographic regions have significant-notable impacts on the results (see Supplementary Table 7).

311 The baseline scenario is less sensitive than the paper-substitution scenario. The effects of life

bla cycle inventory (LCI) datasets on baseline results of CO<sub>2</sub>, COD and water are within 10% of each

B13 <u>other</u>. SO<sub>2</sub>, NO<sub>x</sub>, PM<sub>2.5</sub>, and dioxin emissions are more sensitive than other indicators.

- 314 Transportation contributing to the largest effects of CO<sub>2</sub>, NOx, COD, and dioxin emissions is
- 315 more sensitive than other lifecycle phases. If the weights of food container and bag were
- increased by 5%, their environmental impacts would increase by 1% to 4% (see Supplementary
- Table 8). Paper containers and bags are more sensitive to plastic ones for packaging weights.

318 The shared tableware and packaging could balance out the  $CO_2$ ,  $SO_2$ ,  $NO_x$ ,  $PM_{2.5}$  and COD319 emissions of the same amount of single-use plastic packaging in the baseline scenario after 320 being reused 14 times (39 times for water consumption and 91 times for dioxin emission, as 321 shown in Supplementary Table 9). Even under 90% and 75% of return rate, shared tableware 322 requires 20 reuses to offset the impact of the disposable item in baseline and paper-323 substitution scenarios (43 times for water consumption and 122 times for dioxin emission. 324 The sustainable model of sharing tableware needs to be established to achieve win-win 325 amongst government, restaurants, food delivery platforms, and consumers. Measures for the supervision and administration of takeaway food safety <sup>45</sup> and food safety operation 326 327 specifications <sup>46</sup> have been acted upon in the online takeaway services of China since 2018, and 328 the government should propose incentives and punitive schemes for the adoption and safe use 329 of sharing tableware. The online food delivery platforms should be responsible for the 330 distribution and inspect-monitor the usage of shared items. The restaurants and the consumers 331 could increase star ratings and receive subsidies by using and returning the reusables. Public 332 education and guidance encourage consumers to make sustainability a key factor in using and 333 returning sharing items. The sharing tableware should be used as a pilot in cities that have large 334 takeaway customer bases. With joint efforts and mutual cooperation, the sharing packaging 335 mechanism can not only accelerate the transition to a zero-waste takeaway future, but also be 336 promoted to the industry of retail, catering, and logistics to create a zero-waste society. By 337 comparing life-cycle environmental impacts of sharing takeaway packaging with single-use 338 items, we hope that tableware-sharing can serve as a feasible solution for reducing food 339 delivery packaging waste that many cities around the globe struggle with, help integrated 340 policy-making for the sustainable development of the takeaway industry.

There are uncertainties and limitations in this study. We made assumptions to simplify the type, material, and size of tableware and packaging. The city-level meal ordering data were collected from Meituan platform, and the possible asymmetries existing in the remaining takeaway market were not considered. The resource consumption during the washing process may be

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different among shared items, we calculate them as a tableware set due to the data limitation.

346 Life-cycle inventories for seven environmental indicators were compiled, impact category

347 indicators are quantified to assess the effects of takeaway industry on the environment and

348 human health. We only focus on environmental impacts of takeaway packaging, and the food

349 waste are excluded. A population's acceptance and human behavioral change under the sharing

350 mechanism is a good point to explore the environmental impacts of food waste.

#### 351 Methods

352 Life-cycle environmental impacts of China's takeaway industry were estimated under three scenarios 353 (see scenario design), while potential environmental mitigation strategies with different packaging 354 materials and management mechanisms were explored. System boundary and function unit was 355 production, packaging, transportation, utilization, and disposal of annual tableware and packaging 356 consumed in China's takeaway industry (see Extended Data Figure 3), and the production of machinery 357 and infrastructure was excluded. Since cutlery, napkins, and chopsticks are habitually bundled with 358 takeaway orders, and each takeaway is assumed to be equipped with a set of tableware and packaging 359 (see Extended Data Figure 1 and Supplementary Table 1 for more details). Based on the life cycle 360 thinking method and ISO 14040/44 methodological guidelines<sup>47,48</sup>, the annual environmental impacts 361 was calculated by multiplying the consumption of tableware and packaging by the corresponding 362 emission factor (see Equation 1).

$$EF_{s,k} = \sum_{i=1}^{I} \sum_{j=1}^{J} AD_{s,i} \cdot CF_{s,k,i,j}$$
 Equation 1

363 where  $EF_{k,s}$  represents the environmental emission and water resource consumption of environmental 364 indicator k under scenario s;  $AD_{s,i}$  denotes the annual takeaway or packaging i consumption related 365 to takeaway order amount under scenario s;  $CF_{s,k,i,i}$  indicates the emission factor of environmental 366 indicator k and tableware and packaging type i in life cycle process j under scenario s; Index j shows 367 the life cycle phase; k represents different environmental or resource indicators, including carbon 368 dioxide, sulfur dioxide, nitrogen oxides, particulates less than 2.5 µm, dioxin (measured as 2,3,7,8-369 tetrachlorodibenzo-p-dioxin), river water consumption, and chemical oxygen demand; s expresses 370 different tableware management scenarios; *i* represents five types of tableware and cutlery (food

container, spoon, wood chopsticks, wooden toothpick, napkin), three types of packaging (packaging bag,
cutlery wrapper, toothpick wrapper) and one transport packaging (corrugated carton).

#### 373 Takeaway data collection

374 As there are no publicly available and comprehensive data on the amount of online takeaway order, the 375 street-level takeaway order data was collected from one of the largest Chinese online takeaway 376 platforms, Meituan (waimai.meituan.com), making up 59% of the China's takeaway market share in 2018 377 and having more than 250 million users <sup>49</sup>. The platform recorded every takeaway food order for each 378 restaurant in each street within each city over the past 30 days, and we accessed Meituan website at the 379 beginning of each month (from March to August 2018). The six-month takeaway order information was 380 downloaded and compiled in Microsoft Excel using a web crawler. 2.8 billion street-level takeaway order 381 volumes covered 430,000 restaurants in 353 Chinese cities between February 2018 to July 2018. To 382 better discuss the takeaway environmental impacts in the city level, we aggregated the street takeaway 383 order data to the city-level.

- 384 The average daily online takeaway transactions come to 1,534,000, which covers 88% of the actual
- transaction volume of Meituan in 2018 <sup>50</sup>. 82.6% of users choose takeaway ordering service through the
- online platform, and 64.1% consumers order takeaway from Meituan, followed by Ele.me (25%)<sup>41</sup>,
- 387 indicating Meituan takeaway order data is representative for exploring city-level order behavior
- 388 difference and associated environmental impacts of China's online takeaway industry. Assuming the
- 389 takeaway order volume follows a uniform distribution over time, six-month takeaway order volume of
- 390 Meituan is expanded two-fold to represent the annual takeaway order volume, and the takeaway order
- in the whole industry is determined based on Meituan's market share (see Supplementary Table 5).

### 392 Scenario design

### 393 Baseline scenario

394 The baseline scenario is designed from the current packaging material and waste disposal patterns.

- 395 Plastic single-use food containers are extensively used in China, occupying 90% of total (polypropylene
- 396 (PP) and polystyrene (PS) each half) <sup>51,52</sup>, while the polyethylene (PE)-coated paper box contributes 10%.
- 397 The environmental impacts of food container are calculated by the weighted sum based on their market
- 398 shares. The spoon is made of PP, and chopsticks and toothpicks are made of birch wood. The packaging

399 bag is made of low-density polyethylene (LDPE), the napkin is made of virgin bleached chemical pulp, 400 and the cutlery wrapper and chopstick wrapper are respectively made of biaxially oriented 401 polypropylene (BOPP) and printing paper. A corrugated carton is considered for the primary packaging 402 for tableware transportation and its specification is listed in Supplementary Table 2. A takeaway is 403 delivered by a courier with the electric bike. In China, only Shanghai and Beijing have enforced the waste classified collection policy since July 2019 and May 2020<sup>20,53</sup>. The post-consumer takeaway packaging 404 405 waste was mixed with municipal solid waste and ended up at an incineration or landfill site, and no 406 waste was recycled.

#### 407 <u>Paper substitution scenario</u>

To further discuss the environmental mitigation potential of the takeaway industry, we design a paper substitution scenario based on the practical pilot case of Shanghai. Takeaway plastic containers and bags are substituted by paper ones. If food providers fail to implement the new standards, they will face platform-specific punishments, including lower rankings, and canceling platform subsidies. Food containers and bags are made of kraft paper, and paper box is coated by PE film. Other tableware and packaging materials and their end-of-life are the same as those used in the baseline scenario.

#### 414 <u>Tableware-sharing scenario</u>

415 The tableware-sharing scenario is designed based on ideas of sharing economy. Reusable containers 416 have been successfully adopted in global takeaway industry. For example, the EcoBox initiative based on 417 deposit-return is developed for transporting meals at the restaurant, canteen, and takeaway food outlet 418 in Luxembourg. As the largest lunch box producer in Tokyo, Japan, Tamago-ya company delivers "bento" 419 lunch boxes to local office workers at noon and collects the box in the afternoon by the courier. A 420 restaurant named Yi Kou Liang Shi in Beijing has applied reusable tableware to delivery takeaway food, 421 90% of reusable tableware can be centralizedly collected. The applications in the United States, Europe, 422 Southeast Asia, and Austria have demonstrated the feasibility of the reusable tableware <sup>54</sup>, which set a 423 good example for the sharing tableware mechanism implementation of China. 424 Paper, glass, ceramic, stainless steel, and silicone are alternative materials for food container. Paper

425 container cannot ensure a tight seal and is not suitable for hot liquid food and soup. The reused glass

426 and ceramic containers are safe for microwave and dishwasher. For the same volume, glass and ceramic

427 containers are the heaviest, and they are more prone to breakage during delivery than others. Due to 428 the decreased corrosion and temperature resistance, stainless-steel container may not be suitable for 429 long-term food storage and delivery. Silicone is considered as an ideal material for food container 430 attributed to the superiorities of safety, long-term usage (ten-year lifetime for Partita silicone food 431 container), and easy cleaning. The thermal insulation property could keep takeaway food warm during 432 the delivery. For the above reasons, we selected food-grade silicone as the material for reusable food 433 container and spoon.

434 The container is designed with dual compartments, which can be used to store both staple food (i.e. rice) 435 and dishes, thereby reducing the numbers of food packaging consumption by one-half. A recycled HDPE non-436 woven bag is selected to carry the takeaway as they are tough, durable, cost-effective, and reusable 437 (maximum lifespan of 180 uses). The napkin and toothpick wrapper are made from 100% recycled 438 content. 100% recycled napkin paper is used to wrap the cutlery, and plastic cutlery wrapper is not 439 required. Chopsticks are made of beech wood with a lifetime of two years and should be replaced every 440 six months from the health perspective. The post-consumer toothpick, napkin, cutlery wrapper, 441 corrugated carton, and broken tableware and cutlery were collected and transported to a recycling 442 facility, and the recycling rate is assumed to be 100%.

443 Differentiated takeback mechanisms and cleaning ways are considered: (1) Centralized collection with 444 manual washing. Snacks and fast food are the biggest players in Chinese online catering market, contributing 44% of the total number of restaurants in 2018 <sup>55</sup>. As some snack and fast food providers do 445 446 not have space for dishwasher, sharing tableware is assumed manually washed in the restaurant. The 447 post-consumer tableware is collected at the next delivery and taken back to the restaurant in which the 448 courier picks up a new takeaway order. (2) Decentralized collection with machine washing. Consumers 449 can return the tableware to collection points from where it is delivered to central cleaning stations by 450 diesel truck. The cleaning stations equipped with commercial dishwashers are responsible for cleaning 451 and disinfection of tableware and taking back to the restaurant. Given that shared containers and 452 packaging could be all returned and cleaned on the same day after use, a batch of tableware and 453 packaging with the same amount of average daily takeaway order volume is put on the market and 454 reused for one year. 360 uses for one batch of containers and spoons, and 180 uses for two batches of 455 chopsticks and non-woven bags, are calculated in this scenario. The tableware-sharing scenario is an 456 optimal tableware set and aims to lessen environmental impact.

## 457 Life-cycle inventory

Due to a lack of consistent and systematic life cycle inventory of food packaging products in China, the life-cycle inventories of the takeaway industry were compiled by direct measurements (weight), China life cycle database (CLCD, China-Public 0.8) <sup>56</sup>, peer-reviewed literature and manufacturers' data, and data gaps were filled by the background attributional datasets of Ecoinvent (v3.5) <sup>57,58</sup>. The production of tableware and packaging was considered to be in China (see Extended Data Figure 4 for manufacturer distributions), and the technology level during the production, transportation, and disposal was assumed to be homogenous within each city.

#### 465 Production of raw material and tableware

466 The food container, spoon, plastic bag, cutlery wrapper, and PE film of the paper are made of petroleum-467 based polymers. Chinese average data of PS and LDPE granule production from CLCD have been applied 468 <sup>56</sup>. The production of PP and silicone came from the rest of the world (RoW) of Ecoinvent, which was 469 aggregated data for all processes from raw material extraction until delivery at plant <sup>57</sup>. The polymers 470 were extruded and thermoformed to final products of tableware and packaging, while conversion 471 processes, including injection moulding, foaming, blow moulding and stretch blow moulding came from 472 the RoW dataset, Ecoinvent<sup>57</sup>, and the losses and auxiliaries in the production process were included. 473 The nonwoven bag is made of nonwoven textiles from PP granules. The consumptions of nonwoven 474 fabrics, electricity, and cotton yarn were from the local manufacturer, while LCI of electricity production 475 was sourced from market for electricity, medium voltage (CN) dataset in Ecoinvent<sup>57</sup>, and others came 476 from RoW dataset.

477 Paper container, paper bag, napkin, toothpick wrapper, and corrugated board box belong to paper 478 products.  $CO_2$  emission inventories of production of packaging paper, corrugated board, and tissue paper 479 in China were sourced from Chen, et al. <sup>59</sup>. Chinese CO<sub>2</sub>, SO<sub>2</sub>, NO<sub>x</sub>, PM<sub>2.5</sub>, COD emissions and water inventories of writing paper were collected from Ren<sup>60</sup> to model the production of the toothpick 480 481 wrapper. The life cycle inventories of kraft paper (bleached, unbleached) were used to model the 482 production of the paper container and paper bag <sup>57</sup>. The single-wall corrugated board box was sourced from the corrugated board box production (RoW) dataset <sup>57</sup>. The production of napkin and 100% 483 484 recycled printing paper respectively sourced from the production of tissue paper production (virgin,

GLO) and graphic paper production of Ecoinvent <sup>57</sup>. The electricity consumed in cutting and folding into
small sized portable napkin was collected from local manufacturer. The electricity and ethylene-vinyl
acetate copolymer consumed in cutting and gluing during toothpick wrapper production were collected
from local manufacturer.

489 Single-use chopstick and toothpick are made from birch with 0.45 g/cm<sup>3</sup> of air-dried density, and
 490 reusable ones are made of beech wood with 0.79 g/cm<sup>3</sup> of air-dried density. The chopstick

491 manufacturing process involves logging, milling, shaping, bleaching, natural drying, and polishing, while

492 inputs of electricity, water, sulfur dioxide and paraffin wax came from local manufacturer. The wood

493 effective utilization rate during disposable chopsticks manufacturing was 60% <sup>61</sup>. See Supplementary

494 Table 3 for unit process and data source of production of each tableware and packaging.

### 495 <u>Transportation</u>

496 Transportation includes the transportation of secondary materials for tableware production, tableware 497 production for suppliers and takeaway delivery. The tableware manufacturer distributions at city level 498 came from Alibaba (www.1688.com), one of the largest online wholesale platforms in China. More than 499 7,000 manufacturers of tableware and packaging are located in Zhejiang, Guangdong, Jiangsu, Fujian and 500 Shandong Provinces of China (see Extended Data Figure 4). The raw materials were assumed to travel 501 150 km from raw material production plants to the tableware and packaging manufacturers by a heavy-502 duty diesel truck <sup>31</sup>. After being packaged in the above provinces, the tableware and packaging was 503 transported to the distribution centre across the country, while the transport route was determined 504 based on the shortest path principle, and distances are collected from Baidu map (map.baidu.com) listed 505 in Supplementary Table 4. Tableware and packaging were then distributed to the retailer, assuming a 506 distance of 150 km<sup>26</sup>. Life-cycle inventories for heavy diesel truck (18 tonnes) were collected from CLCD 507 <sup>56</sup>. The transportation of post-consumer tableware from waste collection plants to the final disposal sites 508 was included in the final disposal phase.

509 There are 2.7 million Meituan riders in 2018, 45% of the riders receiving more than 20 orders per day, 510 and 40% of the riders travel more than 50 kilometres per day<sup>62</sup>, and annual total travel distances and 511 total delivery orders were determined based on these distributions. By dividing the total number of 512 takeaway orders by annual travel distance, the delivery distance of each order was 2.0 km, identical with 513 survey results in Wen, et al. <sup>28</sup>. Electricity consumption per 100 km of electric bikes is estimated by the voltage, current and endurance mileage <sup>63</sup>. Due to the large market share, we take two-wheeled food
delivery electric bike produced in Zhuhai Weifan Lithium battery technology co. LTD (48V, 48AH, 155km)
for example, the charge-discharge efficiency of lithium battery is 95% and its electricity consumption is
1.56 kWh per 100 km. Electricity consumed per order during takeaway delivery is 0.032 kWh. The life
cycle emission factor of the provincial electricity grid mix in China from Ecoinvent is adopted to reflect
the regional environmental differences of electricity production <sup>57</sup>.

#### 520 Utilization

521 Single-use tableware and packaging produce no additional environmental impact in this process. For the 522 reusable items, impacts of takeback logistics and tableware washing were considered. The energy and 523 water consumed in manual and machine dishwashing were from a research report, indicating to clean 74 524 dishes and achieve the same acceptable level of cleaning performance, manual dishwashing consumed 525 45.9 litres of water and 1.39 kWh of electricity (mainly from hot water), and machine dishwashing only 526 consumed 11.5 litres of water and 0.92 kWh of electricity<sup>64</sup>. They found that electric dishwashers have a 527 substantialsignificant water-saving effect, which is consistent with the finding of Europe study <sup>65</sup> and 528 Chinese test reports <sup>66,67</sup>. The detergent consumed in machine and manual dishwashing was respectively from the local manufacturer and Gallego-Schmid, et al.<sup>26</sup>. The life cycle inventory in production of water 529 530 and detergent come from tap water production (RoW) and non-ionic surfactant production (RoW) of 531 Ecoinvent <sup>57</sup>. Takeback logistics for centralized collection by courier was included in the tableware 532 delivery phase. The tableware in collection points is delivered to central cleaning centre and sent back to 533 restaurants after cleaning and disinfecting (heavy diesel truck, 18 tonnes), assuming a distance of 100 534 km.

#### 535 End-of-life

We assumed that the takeaway tableware and packaging within each province were disposed of in the same way. The proportion of incineration and landfill of MSW for each province were collected from the China statistical yearbook <sup>68</sup>. The treatment of waste paper, wood, and various waste plastic in municipal incineration and sanitary landfill were sourced from RoW dataset, Ecoinvent <sup>57</sup>. The dioxin emission factor of Chinese MSW incineration was collected from Ni, et al. <sup>69</sup>. The inventories of sorting and recycling of waste plastic, paper and wood were from Ecoinvent <sup>57</sup>. Due to a lack of data on the treatment of waste silicone, treatment of waste PE for recycling was used to estimate end-of-life impactsof silicone tableware and spoon.

## 544 Sensitivity analysis

The LCI datasets from different geographical regions and the weight of tableware and packaging may
affect the emission factor and activity data (quantities of raw material and production resources
required). The effects of LCI datasets from Europe (RER) and RoW, Ecoinvent (V3.5) on environmental
impacts were investigated under three scenarios. Since food container and bag was responsible for more
than three-fifth of entire environmental impacts, the sensitivity analysis of weights of container and bag
was then performed. Baseline, paper substitution and tableware sharing scenarios were considered as
the benchmarks and the weights of container and bag are designed 5% heavier than the benchmark.

552 The reuse time is one of the significant-important parameters for evaluating the environmental benefits of shared tableware and packaging <sup>25,54</sup>. Each environmental indicator was calculated to explore how 553 554 many times reusable packaging should be used to balance out the impacts of one use for single-use 555 alternatives in the baseline and paper substitution scenarios. Since the impact of food delivery is the 556 same, it was excluded from the estimation. The production, transport, and end-of-life of corrugated 557 carton for packaging tableware is excluded. The return rate of sharing packaging is another parameter 558 with high uncertainty, which mainly relies on the takeback behaviour of consumer. Based on the average 559 return data of a Chinese takeaway restaurant named Yi Kou Liang Shi, we assumed 90% of shared-560 tableware can be centralizedly collected in real operation. There is no decentralized collection example 561 in the Chinese takeaway industry but the express delivery industry. Based on the return rate of sharing 562 express packaging in pilots of Zhejiang's universities, 75% of shared tableware is assumed to be 563 decentralized collected in practical application. It means that to replace one unit of single-use 564 alternative, it is respectively required 1.1 unit and 1.3 unit of shared tableware set for centralized and 565 decentralized collection. The effects of return rate on the environmental differences-in for each indicator 566 were explored.

## 567 Data availability

The weight of tableware and packaging and cities' takeaway order data are respectively provided in
Supplementary Table 1 and Supplementary Table 5. The life-cycle inventories are sourced from

- 570 manufacturers' data, China life cycle database <sup>56</sup>, Ecoinvent <sup>57</sup> and literature sources<sup>59,60,69</sup>. All data used
- 571 in the study are available from the corresponding author upon reasonable request. Source data are
- 572 provided.

# 573 Code availability

574 All programming codes are available from the corresponding author upon reasonable request.

# 575 **References**

- Hirschberg, C., Rajko, A., Schumacher, T., & Wrulich, M. The changing market for food delivery.
   *McKinsey & Company* (2016).
- 578 2 Hotrec. Shedding light on the 'meal-sharing' platform economy. (2018).
- Maimaiti, M., Zhao, X.Y., Jia, M.H., Ru, Y. & Zhu, S.K. How we eat determines what we become:
  opportunities and challenges brought by food delivery industry in a changing world in China. *European Journal of Clinical Nutritionvolume* **72**, 1282–1286 (2018).
- 582 4 China's Catering Industry Development Report 2018 (in Chinese). (China Cuisine Association,
  583 Beijing, 2018).
- 5845Research report on China online food delivery industry 2018-2019 (in Chinese). *iiMedia*585*Research*. <u>https://www.iimedia.cn/c400/64223.html</u> (2019).
- 5866Research report on China takeaway development in 2017 (in Chinese). Meituan Research587Institute (2018).
- Jambeck, J. R. *et al.* Marine pollution. Plastic waste inputs from land into the ocean. *Science* 347, 768-771 (2015).
- 590 8 China's food-delivery business is booming. So is waste. *The Economist* (2017).
- 591 9 Zheng, J. & Suh, S. Strategies to reduce the global carbon footprint of plastics. *Nature Climate* 592 *Change* 9, 374–378 (2019).
- 593 10 Wang, A. N., Ma, X.C., & Zhang. C. Xinhuanet. <u>http://www.xinhuanet.com/mrdx/2017-</u>
   594 09/21/c\_136626055.htm (2017).
- 595 11 NBS. National Data (National Statistical Data Repository), National Bureau of Statistics of China.
   596 <u>http://data.stats.gov.cn/easyquery.htm?cn=A01</u> (2019).
- Annual report on prevention and control of environmental pollution by solid waste in Chinese
   large and medium cities in 2018 (in Chinese). *Ministry of ecology and environment of the People's Republic of China* (2018).
- Implementation plan of waste sorting system (in Chinese). *General office of State Council of the People's Republic of China*. <u>http://www.gov.cn/zhengce/content/2017-</u>
   <u>03/30/content\_5182124.htm</u> (2017).
- 603 14 Work Plan for the Pilot Program of "Zero Waste Cities" Construction (in Chinese). General office

604 605		of State Council of the People's Republic of China. <u>http://www.gov.cn/zhengce/content/2019-</u> <u>01/21/content_5359620.htm</u> (2019).
606 607 608 609	15	Opinion on Further Strengthening the Control of Plastic Pollution (in Chinese). <i>National Development and Reform Commission, Ministry of Ecology and Environment of the People's Republic of China,</i> <u>http://www.mee.gov.cn/xxqk2018/xxqk/xxqk10/202001/t20200120_760495.html</u> (2020).
610 611 612	16	T/31SAFCM006-2018. Management Standards for food service (Internet) Take-out of paper bowl and food delivered bag (in Chinese). <i>Shanghai Municipal Quality and Technical Supervision Bureau</i> (2018).
613 614	17	T/31SAFCM004-2018. General technical requirements for food service (Internet) Take-out of paper bowl (in Chinese). Shanghai Municipal Quality and Technical Supervision Bureau (2018).
615 616 617	18	T/31SAFCM005-2018. General technical requirements for food service (Internet) Take-out of food delivered bag (in Chinese). <i>Shanghai Municipal Quality and Technical Supervision Bureau.</i> (2018).
618 619 620 621	19	First group standards of takeaway containers were launched and plastic container would be replaced by paper one (in Chinese). Shanghai Municipal People's Government <u>http://www.shanghai.gov.cn/nw2/nw2314/nw2315/nw17239/nw17240/u21aw1306741.html</u> (2018).
622 623 624	20	Shanghai Municipal Solid Waste Management Regulation (in Chinese) Shanghai Municipal People's Congress. <u>http://www.spcsc.sh.cn/n1939/n1944/n1946/n2029/u1ai185433.html</u> . (2019).
625 626 627	21	Madival, S., Auras, R., Singh, S. P. & Narayan, R. Assessment of the environmental profile of PLA, PET and PS clamshell containers using LCA methodology. <i>Journal of Cleaner Production</i> <b>17</b> , 1183-1194 (2009).
628 629	22	Dormer, A., , Finn, D.P., Ward, P., & Cullen, J. Carbon footprint analysis in plastics manufacturing. Journal of Cleaner Production <b>51</b> , 133-141 (2013).
630 631 632	23	Accorsi, R., Cascini, A., Cholette, S., Manzini, R. & Mora, C. Economic and environmental assessment of reusable plastic containers: A food catering supply chain case study. <i>International Journal of Production Economics</i> <b>152</b> , 88-101 (2014).
633 634 635	24	Leejarkpai, T., Mungcharoen, T. & Suwanmanee, U. Comparative assessment of global warming impact and eco-efficiency of PS (polystyrene), PET (polyethylene terephthalate) and PLA (polylactic acid) boxes. <i>Journal of Cleaner Production</i> <b>125</b> , 95-107 (2016).
636 637 638	25	Gallego-Schmid, A., Mendoza, J.M.F. & Azapagic, A. Improving the environmental sustainability of reusable food containers in Europe. <i>Science of The Total Environment</i> <b>628–629</b> , 979-989 (2018).
639 640	26	Gallego-Schmid, A., Mendoza, J. M. F. & Azapagic, A. Environmental impacts of takeaway food containers. <i>Journal of Cleaner Production</i> <b>211</b> , 417-427 (2019).
641 642	27	Cheroennet, N., Pongpinyopap, S., Leejarkpai, T. & Suwanmanee, U. A trade-off between carbon and water impacts in bio-based box production chains in Thailand: A case study of PS, PLAS,

PLAS/starch, and PBS. Journal of Cleaner Production 167, 987-1001 (2017). Wen, Z., Zhang, Y. & Fu, D. The environmental impact assessment of a takeaway food delivery order based on of industry chain evaluation in China (in Chinese). China Environmental Science , 4017-4024 (2019). Razza, F., Fieschi, M., Innocenti, F. D. & Bastioli, C. Compostable cutlery and waste management: an LCA approach. Waste management 29, 1424-1433 (2009). Fieschi, M. & Pretato, U. Role of compostable tableware in food service and waste management. A life cycle assessment study. Waste management 73, 14-25 (2018). Siracusa, V., Ingrao, C., Lo Giudice, A., Mbohwa, C. & Dalla Rosa, M. Environmental assessment of a multilayer polymer bag for food packaging and preservation: An LCA approach. Food Research International 62, 151-161 (2014). Life cycle assessment of grocery carrier bag. The Danish Environmental Protection Agency. Copenhagen, Denmark (2018). Marsh, K. & Bugusu, B. Food Packaging and Its Environmental Impact. Food Technology Magazine, 46-50 (2007). Rossi, V. et al. Life cycle assessment of end-of-life options for two biodegradable packaging materials: sound application of the European waste hierarchy. Journal of Cleaner Production 86, 132-145 (2015). Tecchio, P., Freni, P., De Benedetti, B. & Fenouillot, F. Ex-ante Life Cycle Assessment approach developed for a case study on bio-based polybutylene succinate. Journal of Cleaner Production , 316-325 (2016). Van Doorsselaer, K. & Lox, F. Estimation of the energy needs in life cycle analysis of one-way and returnable glass packaging. Packaging Technology and Science 12, 235-239 (1999). Wood, G. & Sturges, M. Final report: Reusable Packaging - Factors to Consider. Single Trip or Reusable Packaging - Considering the Right Choice for the Environment. ISBN: 1-84405-84437-84432 (Wrap (Waste and Resources Action Programme), Banbury, UK., 2010). The new plastics economy-rethinking the future of plastics. World Economic Forum, Ellen MacArthur Foundation, McKinsey & Company (2016). New plastics economy: reuse-rethinking packaging. Ellen MacArthur Foundation (2019). Heinrichs, H. Sharing economy: a potential new pathway to sustainability. GAIA 22, 228–231 (2013). Research report on the market development of online takeaway service (in Chinese). Data Center of China Internet (2019). Towards the Circular Economy: an economic and business rationale for an accelerated transition. Ellen MacArthur Foundation, McKinsey & Company (2013). EIA. International Energy Outlook 2016. (U.S. Energy Information Adimistration, 2016). Cadman, J., Evans, S., Holland, M. & Boyd, R. Proposed Plastic Bag Levy - Extended Impact Assessment. Scottish Executive (2005).

681 682 683	45	Measures for the supervision and administration of food Safety in online catering services (in Chinese). <i>State Council Bulletin of the People's Republic of China.</i> <u>http://www.gov.cn/gongbao/content/2018/content_5268787.htm</u> (2018).
684 685 686	46	Food safety operation specification for the catering service (in Chinese). <i>State Administration for Market Regulation of the People's Republic of China.</i> <u>http://www.samr.gov.cn/spjys/tzgg/201902/t20190226_291361.html</u> (2018).
687 688	47	ISO. ISO14040: 2006 Environmental Management-Life Cycle Assessemnt-Principles and Framework. ISO Standards, Geneva (Switzerland) (2006).
689 690	48	ISO. ISO14044: 2006 Environmental Management-Life cycle Assessment-Requirements and Guidelines. International Organization for Standardization (2006).
691 692	49	China's Sharing Economy Development Report 2018 (in Chinese). <i>State Information Center of China</i> (2018).
693 694	50	China real-time distribution industry development report in 2018 (in Chinese). <i>China Federation of Logistics and Purchasing</i> (2018).
695 696	51	China General Chamber of Commerce. Survey report on the use of disposable tableware (in Chinese). <i>China Journal of Commerce</i> <b>7</b> , 34 (2010).
697 698	52	Wei, Z. The market situation and suggestions of disposable food tableware in China (in Chinese). <i>China packaging</i> <b>10</b> , 59-61 (2011).
699 700 701	53	The decision to amend the Beijing Municipal Solid Waste Management Regulation (in Chinese). The People's Government of Beijing Municipality. <u>http://www.beijing.gov.cn/zhengce/zhengcefagui/201912/t20191204_834225.html</u> (2019).
702	54	Ellie Moss, R. G. The dirty truth about disposable foodware. The Overbrook Fundation (2020).
703	55	Chinese catering report 2019 (in Chinese). Meituan-Dianping, Beijing (2019).
704 705	56	Liu, X. L. <i>et al.</i> Method and basic model for development of Chinese reference life cycle database (in Chinese). <i>Acta Scientiae Circumstantiae</i> <b>30</b> , 2136-2144 (2010).
706	57	The Swiss centre for life cycle inventories. Ecoinvent database v3.5. (2018).
707 708	58	Wernet, G. <i>et al</i> . The ecoinvent database version 3 (part I): overview and methodology <i>The International Journal of Life Cycle Assessment</i> <b>21</b> , 1218-1230 (2016).
709 710	59	Chen, S., Yang, X. G., Li, Y. P., Cao, L. & Yue, W. C. Life-cycle GHG emissions of paper in China (in Chinese). <i>Journal of Beijing University of Technology</i> <b>40</b> , 944-949 (2014).
711 712	60	Ren, L. Methodology research and typical paper products of life cycle assessment [Master thesis] (in Chinese), Beijing University of Technology, (2011).
713 714	61	Tian, M. & Yin, Z. Study on the export of disposable wooden chopsticks from China to Japen in great quantities (in Chinese). <i>Journal of Beijing Forestry University (Social Sciences)</i> <b>3</b> , 1-5 (2006).
715	62	Takeaway rider employment report in 2018 (in Chinese). Meituan Research Institute (2019).
716 717	63	GB17761-1999. General technical conditions of electric bicycle (in Chinese). <i>Standardization Administration of the People's Republic of China</i> .

- Research report of the performance of machine washing and manual washing (in Chinese). All China Environment Federation, University of Bonn, Beijing University of Technology, Shanghai
   Jiao Tong University (2014).
- Stamminger, R., Elschenbroich, A., Rummler, B. & Broil, G. Washing-up behaviour and techniques
  in Europe. *HuW* 55, 31–37 (2007).
- 66 Comparative test report on household dishwashers (in Chinese). *China Consumer Association*,
  724 *Consumer Protection Committee of Zhejiang Province, Consumer Protection Committee of*725 *Qingdao City, Consumer Association of Jinan City.* (2019).
- 67 Comparative test report on household dishwashers in 2018 (in Chinese). *Consumer Council of Guangdong province, Consumer Council of Foshan City* (2018).
- 728 68 NBS. *China Statistical Yearbook 2018*. (China Statistical Press, 2019).
- Ni, Y. W. *et al.* Emissions of PCDD/Fs from municipal solid waste incinerators in China. *Chemosphere* **75**, 1153-1158 (2009).
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# 737 Author contributions

- 738 Y.Z. and D.G. designed the study. Y.Z., W.X. and J.L. prepared data. Y.Z. conducted calculations and drafted
- the manuscript. D.G., Y.Z. and Y.S. led the analysis. Y.Z and Y.S drew the figures. All authors (Y.Z, Y.S., D.G,
- 740 X.L., Y.C., J.L., W.X., J.X., Z.M. and Z.Y.) participated in discussing the results and contributed to writing the
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# 742 Competing interests

743 The authors declare no competing interest.

# 744 Additional information

- 745 **Extended data** is available for this paper.
- 746 **Supplementary information** is available for this paper.

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