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**Migration in the Anthropocene: how collective navigation,
environmental system, and taxonomy shape the
vulnerability of migratory species**

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22 **Abstract**

23 Recent increases in human disturbance pose significant threats to migratory species using
24 collective movement strategies. Key threats to migrants may differ depending on behavioural
25 traits (e.g. collective navigation), taxonomy, and the environmental system (i.e. freshwater,
26 marine, or terrestrial) associated with migration. We quantitatively assess how collective
27 navigation, taxonomic membership, and environmental system impact species' vulnerability by
28 1) evaluating population change in migratory and non-migratory bird, mammal, and fish species
29 using the Living Planet Database (LPD), 2) analysing the role of collective navigation and
30 environmental system on migrant extinction risk using International Union for Conservation of
31 Nature (IUCN) classifications, and 3) compiling literature on geographic range change of
32 migratory species. Likelihood of population decrease differed by taxonomic group: migratory
33 birds were more likely to experience annual declines than non-migrants, while mammals
34 displayed the opposite pattern. Within migratory species in IUCN, we observed that collective
35 navigation and environmental system were important predictors of extinction risk for fishes and
36 birds, but not for mammals, which had overall higher extinction risk than other taxa. We found
37 high phylogenetic relatedness among collectively navigating species, which could have obscured
38 its importance in determining extinction risk. Overall, outputs from these analyses can help guide
39 strategic interventions to conserve the most vulnerable migrations.

41 **Keywords:** migration, birds, mammals, fishes, Living Planet Database, collective navigation

43 **Introduction**

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3 44 Migration is one of life's most complex and ecologically consequential behaviours. In
4
5 45 recent years, while technological advances have enabled scientists to describe for the first time
6
7 46 the complex social mechanisms (e.g. collective navigation) that facilitate many migrations [1,2],
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9 47 scientists have also suggested that many such migrations may be in peril [3,4]. However, more
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11 48 spatially and taxonomically ambitious analyses are needed to help identify what specific factors
12
13 49 may control the extinction risk and population status of collectively navigating and other
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15 50 migratory species.
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20 51 Loss of migration is non-trivial: migratory species have been shown to provide unique
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22 52 functions to ecosystems, such as vectoring nutrients and seeds long distances, maintaining
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24 53 grassland ecosystems through cyclical grazing, and providing pulses of prey to resident predators
25
26 54 (reviewed in [5]). Economies that are reliant on tourism or harvesting migratory species (e.g.
27
28 55 wildebeest migrations in the Serengeti) could suffer from their decline [6,7]. Collective
29
30 56 navigation (i.e. group-level pooling of information) aids migratory species' ability to travel long
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32 57 distances, and conserving this behaviour requires maintaining a minimum population size [8],
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34 58 which helps drive important ecological processes [5].
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39 59 Several studies have reported declines of migratory species, although this research has
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41 60 primarily focused on specific taxonomic groups [4], or is qualitative [9]. Sanderson et al. [10],
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43 61 for example, found European long-distance migratory birds to have declined relative to resident
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45 62 European birds. Similarly, Harris et al. [11] reported that of the twenty-four large-bodied
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47 63 migratory ungulates they reviewed, six migrations have been lost. However, studies on
48
49 64 extinction risk found that migration was not an important predictor of extinction risk [12]. One
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51 65 study on birds even suggested that migration decreased risk of extinction [13].
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3 66 In light of these contradicting predictions, exploring how collective navigation,
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6 67 environmental system (i.e. the freshwater, marine, or terrestrial system associated with the
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8 68 migratory population), and taxonomy shape vulnerability to population decline and extinction
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10 69 risk could help focus efforts to preserve migrations. Species that travel long distances and rely on
11
12 70 multiple habitat types are likely to be disproportionately impacted by human actions [10,14].
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15 71 Road building, agricultural development, and dam construction create barriers for long-distance
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17 72 migratory species [15,16], and such impacts can be further compounded by habitat loss,
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20 73 overharvest, and climate change [3,17]. Threats may be even more serious for collectively
21
22 74 navigating species. If these populations fall below a threshold size, Allee effect-caused collapses
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25 75 might occur that are difficult or impossible to reverse [8].
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27 76 The environmental system through which a species migrates can also alter which threats
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29 77 they face; for example, landscape fragmentation and urban development are terrestrial
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32 78 disturbances that will likely have a greater impact on species that migrate on land than on marine
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34 79 or freshwater migrants. Similarly, the comparative lack of human infrastructure and associated
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36 80 fragmentation in marine systems could potentially leave migrating marine species at less risk of
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39 81 extinction [18]. The discrete and comparatively easy-to-modify nature of many freshwater
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41 82 ecosystems (e.g. dams in rivers) may increase risk of extinction for species migrating through
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44 83 these systems [19].
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46 84 However, some characteristics of migratory species could increase their ability to
47
48 85 circumvent potential threats. For example, the synthesized integration of environmental cues can
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51 86 help collective migrants traverse fragmented and complex landscapes [20]. Furthermore,
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53 87 migratory species with large geographic ranges and flexible movement patterns may reduce risk
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56 88 by moving to a new habitat when part of their range is fragmented or degraded [21].
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3 89 In this study, we follow previous definitions of migration, defining it as cyclical and
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5 90 directed round-trip movement of animals between discrete areas [4]. We restrict this definition to
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8 91 include only movements greater than 100 km in one direction [22]. We leverage the power of
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10 92 two large datasets (International Union for Conservation of Nature Red List; IUCN and Living
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12 93 Planet Database; LPD) to examine the status of migrations from three perspectives: 1) by
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14 94 comparing population change over time; 2) by analysing the role of collective navigation and
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16 95 environmental system (freshwater, marine, terrestrial, and their combination) on migrant
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18 96 extinction risk; and 3) by examining contemporary shifts in the geographic range of migratory
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20 97 species. The results from these multiple perspectives contribute to our general understanding of
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22 98 how well migrations will fare in the Anthropocene and better position us to strategically respond
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24 99 to these risks.
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30 **Methods**

31 **Population change over time**

32 *Data selection and coding*

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36 102 To evaluate the relative vulnerability of migratory birds, mammals, and fishes, we
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38 103 created indices using population change data from the Living Planet Database (LPD) [23].
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40 104 Reptiles, amphibians, and invertebrates were excluded from the analyses because of a lack of
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42 105 information in the LPD. The LPD is one of the largest repositories of time-series data relevant to
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44 106 recent population change (see [24] for further details).
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50 108 For this analysis, we included data on all LPD populations of birds, mammals, and fishes
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52 109 (Actinopterygii only, as adequate trend data were available) with sufficient information to
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54 110 determine migratory status. Each population record was coded by its ecological affiliation to
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56 111 three different environmental systems: freshwater, marine, and terrestrial. Environmental
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3 112 systems were assigned based on a number of considerations, including where the population
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5 113 spends the majority of its time, which environmental system(s) help sustain the population, and
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8 114 where the population is most threatened (electronic supplementary material, table S1). Seven
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10 115 candidate biogeographic realms were assigned to terrestrial and freshwater populations and five
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13 116 realms were assigned to marine populations (electronic supplementary material, table S1).

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15 117 We used the Global Register of Migratory Species (GROMS) for migratory coding [22].
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17 118 As above, GROMS also defines migration as predictable and cyclical movements of more than
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20 119 100 km. For details on migratory coding, see electronic supplementary material, tables S2 and
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22 120 S6.

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24 121 After removing replicates (populations monitored in the same location and at the same
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26 122 time), the final data set created for analysis included 11,717 populations of 2,978 species
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29 123 (electronic supplementary material, table S3).

30 31 32 124 *Data analysis and statistical methods*

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35 125 To estimate patterns of abundance change, we followed the Living Planet Index (LPI)
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37 126 method and calculated the geometric mean of trends for each species whereby each population
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39 127 time-series with six or more data points was modelled using a Generalised Additive Modelling
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42 128 (GAM) framework [24]. Population time-series with fewer than six data points or those that
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44 129 resulted in a poor GAM fit were modelled using the chain method [25]. In cases with more than
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47 130 one population time-series per species, the modelled annual trends for each population were
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49 131 averaged within each realm, class, and realm/class combination to provide a single set of annual
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51 132 trends for each species. These were then averaged across species and converted to index values
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54 133 with a baseline year of 1970. For each index, we generated 95% confidence intervals using a
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56 134 bootstrap resampling technique for 10,000 iterations [24]. These confidence intervals
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3 135 demonstrate the uncertainty in the index values inherited from the baseline in 1970 and
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6 136 propagated through the time-series. An R package, *rlpi*, for calculating these index values is
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8 137 available at <https://github.com/Zoological-Society-of-London/rlpi>.

9
10 138 Using these population trends, we explored a range of models to examine the relationship
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13 139 between migratory behaviour and population trends. For total change in abundance across all
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15 140 years we constructed linear mixed effects models in R using the lme4 package [26,27]. For
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17 141 likelihood of annual increases, beta-binomial models were used to model the ratio of increasing
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20 142 and decreasing years for each population using the glmmadmb package in R [28]. In these
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22 143 comparisons, we also excluded populations with no available GROMS status. Across all models,
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24 144 species and realm were included as random effects. Models were compared using the Akaike
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27 145 Information Criterion (AIC).

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29 146 To examine how these differences in migration vulnerability varied by taxonomic
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32 147 grouping all analyses were repeated for birds, mammals, and fishes (for taxonomic breakdown
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34 148 see electronic supplementary material, table S3).

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36 149 We then further analysed differences for taxonomic groups in select regions for which
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39 150 there were sufficient data for robust comparison between taxa and systems. Although system was
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41 151 not an important factor for birds and mammals overall, regional differences between systems
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43 152 were explored. Many regions did not have sufficient data to run these finer-scale analyses or to
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46 153 compare to other systems or taxa (see electronic supplementary material tables S4 and S5 for
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48 154 breakdown by realm). To illustrate the nature of these regional level differences, we focused on
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50 155 the Palearctic and northern temperate Pacific realms because these regions had sufficient data for
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53 156 taxonomic and system comparisons.

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3 158 **Species extinction risk**

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6 159 *Data selection and coding*

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8 160 Our second analysis investigated patterns of migrant extinction risk using data from the
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10 161 IUCN Red List [29]. The IUCN database collects information from global assessments based on
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12 162 standardized protocols, which are linked to population size and structure, population trends, and
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14 163 geographic range. We extracted data on the threat status and environmental system (terrestrial,
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16 164 marine, freshwater, or their combination) for all bird, mammal, and fish species listed in both
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18 165 GROMS and IUCN (n = 3,447) [29]. The migratory status of each species was coded using
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20 166 GROMS. IUCN species were classified as terrestrial, freshwater, marine, or a combination of the
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22 167 three environmental systems based on IUCN's classification scheme (electronic supplementary
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24 168 material, table S6). IUCN threat status was converted to a 1 to 6 ordinal numerical index
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26 169 (following [30]).

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31 170 We used the resultant database to test the effects of environmental system (referred to as
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33 171 "system") and tendency to collectively navigate on migratory species extinction risk. Due to
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35 172 considerable gaps in information on whether species directly exhibit collective movement or
36
37 173 collective navigation, we coded each species based on proxy measures. Based on findings from
38
39 174 the literature exploring the emergent dynamics of collective navigation [8,31], we coded
40
41 175 migratory species as putative "collective navigators" if they were determined to travel in social
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43 176 groups (e.g. pods, herds, flocks, etc., as distinguished from loose aggregations of non-interacting
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45 177 individuals). Principle databases used for these determinations were FishBase and IUCN Red
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47 178 List (fishes); Avibase, NatureServe, USFWS Migratory Bird Data Center, and BirdLife
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49 179 International (birds); and IUCN and NOAA Fisheries Office of Protected Resources (mammals);
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51 180 which were supplemented with literature searches. Mammals were coded at the species level.
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3 181 Birds and fishes were coded at higher taxonomic levels, to overcome strong geographic biases in
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5 182 data resolution, lack of data accessibility, and extreme intraspecific variation in migratory
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8 183 behaviour [32]. Birds were coded at the family level. Fishes were initially coded at the order
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10 184 level for those with strong phylogenetic association of schooling, and groups with mixed
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12 185 schooling behaviour were more thoroughly examined and coded at the family or genus level
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15 186 (electronic supplementary material, table S6).
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20 188 *Data analysis*

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22 189 We used cumulative link models (clms) to test the effect of system and its interaction
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24 190 with collective movement on IUCN risk category (EW, EX, CR, EN, VU, NT, LC) for birds,
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26 191 mammals, and fishes. Within each taxonomic class, species were typically found in only a subset
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28 192 of the systems available. For birds, very low sample sizes in freshwater, freshwater/marine, and
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30 193 marine systems required that we pool those groups into terrestrial/freshwater,
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32 194 terrestrial/freshwater/marine, and terrestrial/marine respectively. For fishes, only the systems
33
34 195 freshwater, freshwater/marine, and marine were applicable, and low sample sizes required we
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36 196 pool the highest threat categories (CR, EX, and EN) into one highest risk category. For
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38 197 mammals, species that were classified as inhabiting any system that included freshwater
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40 198 (freshwater, freshwater/marine, freshwater/terrestrial, and freshwater/marine/terrestrial) were
41
42 199 grouped as freshwater to retain an adequate sample size for analyses. When comparing across
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44 200 taxonomic groups, we did not have the proportional odds to compare all threat categories, so the
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46 201 categories were grouped into Lower Risk (LC, NT, VU) and Threatened (EN, CR, EW, EX).
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52 202 We modelled extinction risk with clms using a flexible threshold and logit link function
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54 203 [33] using the ordinal package in R [34]. We then compared all model subsets using AIC to
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3 204 select the best model of extinction risk. Analysis of deviance tables (ANODE) were then used to
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5 205 evaluate the importance of parameters in the best model [34]. Finally, we conducted Tukey-
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8 206 adjusted pairwise comparisons of the nominal effects of the best model to identify significantly
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10 207 different groupings using the lsmeans package [35].
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15 209 **Geographic range shifts**

17 210 *Data collection and coding*

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20 211 To examine changes in the spatial distribution of migratory species, we conducted a
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22 212 literature search investigating geographic range shifts of any type for migratory species of birds,
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24 213 mammals, fishes, and insects. Insects were included in these analyses (and not above) because
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26 214 data were sufficient in the literature to include them in this compilation. All searches were
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28 215 conducted in Web of Science using the search terms “migra*” and “range”, six terms for
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30 216 taxonomic groupings (mammal*, marine mammal*, bird*, fish*, insect*), and four terms
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32 217 denoting geographic change (i.e. shift*, contract*, expand*, change*). We also searched within
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34 218 the citations of the relevant papers for additional papers. This resulted in a total of 5163 papers,
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36 219 of which 66 had quantitatively investigated geographic range shifts in migratory species and
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38 220 produced results concluding some change or no change had occurred (not predictive). All species
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40 221 treated in this literature search were confirmed to match our definition of migration. For each
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42 222 paper, we recorded the study species, taxonomic class, and the system and realm as defined by
43
44 223 IUCN and matched to the LPD. We then categorized the type of geographic range change for
45
46 224 each species as one of four possible categories: contraction, shift, no change, or expansion. The
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48 225 data were split taxonomically and by type of geographic range change. See electronic
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50 226 supplementary material, table S7 for coding and literature.
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6 228 **Results**7
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9 229 **Population change over time**10
11 230 *Global patterns*

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13 231 When data derived from the LPD from all species, systems, and realms were examined
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16 232 together there was no significant difference between migratory and non-migratory species ($p =$
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18 233 0.138) (Fig. 1A).

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20 234 Effect of migration on likelihood of annual increase differed by taxonomic group, with
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23 235 migratory birds significantly more likely to have faced declines than non-migratory birds ($p <$
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25 236 0.001; Table 1) and migratory mammals significantly less likely to experience annual declines
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27 237 than non-migratory mammals ($p = 0.032$; Table 1). From AIC, the best model for predicting the
28
29 238 likelihood of annual increase for birds and mammals was determined to be the simplest model,
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31 239 with Migratory as the only explanatory variable (no interaction between migration and system).
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34 240 However, the best model for fishes was determined to be Migratory*System, which includes the
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36 241 interaction between migration status and system (freshwater and marine). There was no
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38 242 significant difference found between migratory and non-migratory fishes in either system (Table
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41 243 1).

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44 244 *Regional patterns*

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46 245 Data from the LPD were sufficiently robust in nine cases to permit further subdivision
47
48 246 and analyses of pattern at finer regional levels. Of the nine regional trends, six show migratory
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50 247 populations tending to perform worse than non-migratory populations (Fig. 2 and electronic
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52 248 supplementary material, figures S1-S4).
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3 249 In the Palearctic and Pacific regions, migratory populations performed worse than non-
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6 250 migratory populations for each group (Fig. 2), with the exception of Pacific marine birds, for
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8 251 which migratory and non-migratory species appear to have been on similar trajectories
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10 252 (electronic supplementary material, fig. S2). Terrestrial and freshwater Palearctic migratory birds
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13 253 are both faring worse than non-migrants (Fig. 2A and 2B), which matches the taxonomic level
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15 254 results. The trend for terrestrial Palearctic mammal change over time is different from the
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17 255 overall results for mammals; here, Palearctic terrestrial migrants are doing worse than non-
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20 256 migrants (Fig. 2C). Similarly, migratory Pacific marine fishes appear to be doing worse than
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22 257 non-migrants (Fig. 2D), whereas there was no significance in the taxonomic analysis. For Living
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25 258 Planet Indices used to create the figures see electronic supplementary material, table S8.
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27 259

260 **Species extinction risk**

261 Within all migratory species, analysis of IUCN data indicates that extinction risk differed
262 significantly among taxonomic groups. Migratory mammals were more likely to be placed in a
263 higher extinction risk category than either migratory fishes or birds (Table 2). Analyses within
264 taxonomic groups yielded additional insights: for birds and fishes, the best model included model
265 terms System, Collective behaviour, and their interaction. However, no model outperformed the
266 null for mammals (electronic supplementary material, table S9). Extinction risk was higher for
267 non-collective migratory birds in some systems, but not all (Table 2). Migratory birds in
268 terrestrial/marine systems (seabirds) had a higher risk of extinction than avian migrants in other
269 systems; non-collective migratory seabirds were at greater risk of extinction than collective
270 migratory seabirds.

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3 271 For fishes, the best model included the full model that included the interaction of system
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5 272 and collective behaviour. The interaction was significant ($p < 0.001$; Table 2); for collective
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7 273 migrants, marine fishes were placed in lower risk categories than their freshwater counterparts.
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10 274 Non-collective freshwater and anadromous migrants were substantially more at risk of extinction
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12 275 than any marine migrants (Table 2).
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17 277 **Geographic range shift**

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20 278 The literature review of geographic range change of migratory species produced different
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22 279 patterns of change for birds ($n = 451$), mammals ($n = 30$), fishes ($n = 138$), and insects ($n = 107$)
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24 280 (Fig. 3). Eighty percent of the bird records indicated shifts in geographic range, with the
25
26 281 remaining records indicating contractions or expansions. Fish literature produced the second-
27
28 282 largest source of records on geographic range change of all the taxonomic groups with the
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30 283 majority (73%) of these records experiencing range contractions. Although there were relatively
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32 284 few records of mammal geographic range change, 47% of mammal records indicated range
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34 285 contractions. There was extremely low taxonomic diversity in data available on insect
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36 286 geographic range, with all records identified coming from Lepidoptera. Seventy percent of these
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38 287 records showed range expansions.
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45 289 **Discussion**

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48 290 Our global analysis of population change across all taxa did not reveal discernible
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50 291 differences between migratory and non-migratory species; however, analyses of spatially and
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52 292 taxonomically disaggregated data revealed that migration vulnerability and extinction risk are
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54 293 highly context-dependent.
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3 294 *Overall taxonomic vulnerability*

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5 295 Across all taxonomic groups, migration did not consistent predict vulnerability to
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8 296 population decline. This suggests characteristics of taxonomic groups, such as evolutionary
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10 297 history, determine how migration interacts with other variables (i.e. environmental system and
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12 298 collective navigation).

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15 299 *Birds*

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17 300 In agreement with previous studies that used different data sources and analytical
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19 301 approaches [10,36], we observed elevated vulnerability for migratory birds relative to non-
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21 302 migratory birds (Table 1). One possible explanation for the elevated vulnerability of migratory
22
23 303 birds is that they are subject to “multiple jeopardy” because they require many different sites
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25 304 throughout their annual cycles: stopover sites, overwintering habitat, and breeding grounds [14].
26
27 305 Recent field studies have found disproportionate rates of bird mortality during migratory periods
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29 306 (e.g. raptor mortality was six times higher during migration than stationary periods; [37]), and
30
31 307 habitat loss in breeding or overwintering grounds are a significant threat to migratory birds [38].
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33 308 While migratory birds are disproportionately vulnerable to population decline compared to non-
34
35 309 migratory birds, overall extinction risk for birds in IUCN is lower than mammal extinction risk
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37 310 (Table 2). This indicates that while migratory bird populations are more likely to decrease than
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39 311 non-migratory birds, many migratory bird populations are still relatively robust and there is high
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41 312 potential for conservation.

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44 313 *Mammals*

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46 314 Contrary to the well-documented losses of high profile mammal migrations [11,39], our
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48 315 population analysis suggests that migratory mammals were more likely to increase annually than
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50 316 non-migrants. The vagility of migratory mammals could aid their ability to escape anthropogenic
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3 317 threats in areas where non-migrants would have more difficulty moving territories. Additionally,
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5 318 migratory mammals appear to receive disproportionate conservation attention due to their
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8 319 visibility (e.g. ungulate herds and great whales) and influence on tourist-based economies (e.g.
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10 320 safaris and whale-watching), which have helped motivate the direct investment of conservation
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12 321 funds and attention to such species (e.g. rinderpest eradication; [40]).
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15 322 Despite the increases of migratory mammals relative to non-migrants, IUCN lists
16
17 323 migratory mammals in higher threat categories than birds or fishes (Table 2). This discrepancy
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19 324 likely derives from different time periods and variables considered (e.g. IUCN considers
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21 325 variables such as population size and connectivity that the LPD does not). This is likely true for
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23 326 birds and fishes as well.
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27 327 It is also important to note that increasing the abundance of migratory mammals is a
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29 328 necessary but not sufficient condition for the persistence of the migrations themselves. South
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31 329 African wildlife, for example, is often managed in carefully controlled fenced plots. This has
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33 330 facilitated population increases for a number of species in recent years, but this management
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35 331 strategy effectively eradicates migration potential [41,42].
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39 332 *Fishes*

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41 333 There was no significant difference in the likelihood of annual population increase
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43 334 between migratory and non-migratory fishes. In our environmental system-level analyses of
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45 335 IUCN data, however, migratory freshwater fishes were at more risk of extinction than migratory
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47 336 marine fishes (Table 2). Marine fishes face immense threats from overharvest; but, as with other
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49 337 taxa, the mobility of marine migrants could allow migrants to more successfully avoid
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51 338 anthropogenic threats, whereas migratory freshwater fishes are less able to escape habitat
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53 339 modification, pollution, and water extractions [16,43].
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3 340 *Regional vulnerability*

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5 341 Data limitations made regional analyses challenging, but in the northern temperate
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7
8 342 Palearctic and Pacific realms, migrants were more vulnerable to extinction than non-migrants
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10 343 across all environmental systems (Fig. 2). Terrestrial and freshwater migratory birds were again
11
12 344 more vulnerable than non-migrants (Fig. 2A and 2B); although, the heightened vulnerability of
13
14 345 terrestrial mammal and marine fish migrants relative to non-migrants differs from the overall
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16 346 results for mammals and fishes (Fig. 2C and 2D). This suggests declines in migratory
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18 347 populations can be obscured in the aggregated data and emphasizes the need to examine migrants
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20 348 in their local context.
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24 349 *Collective navigation*

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26 350 As previously discussed, collective navigation could be either an aid or a liability for
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28 351 migratory species. While the best model for bird and fish extinction risk included collective
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30 352 navigation, extinction risk did not differ significantly between collective and non-collective
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32 353 species for most groups. However, the lower risk of collectively-moving migratory seabirds
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34 354 relative to non-collectively moving migratory seabirds suggests there might be some benefit to
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36 355 seabirds moving in flocks, in that they are more able to avoid threats than their non-flocking
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38 356 counterparts (e.g. [44]). It is interesting to note that migratory seabirds have the highest risk of
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40 357 extinction as compared to the migratory birds in other environmental systems (Table 2). The
41
42 358 reliance of seabirds on both terrestrial spaces for breeding and marine spaces for feeding could
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44 359 elevate their extinction risk [45], while collective navigation could aid their immense movements
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46 360 and unique use of different resources and habitats. Although collective navigation seems to aid
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48 361 migratory seabirds, the overall population declines in migratory birds are concerning, given the
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3 362 observed relationship between collective navigation and population abundance [8]. With
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5 363 continued decline, collectively moving migrants could face the threat of population collapse.

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8 364 We did not find evidence in migratory mammals for an effect of collective movement.

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10 365 However, the small sample size of migratory mammals in several environmental systems may
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12 366 have played a role in our inability to detect differences between collective and non-collective
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14 367 migrants. It is interesting to note, however, that the most threatened migratory mammal (i.e.
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16 368 Critically Endangered and Extinct in the Wild) species were collective migrants. While largely a
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18 369 benefit, the reliance of collectively navigating species on inter-individual cues can also result in
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20 370 cascading consequences when one individual makes a mistake (e.g. failed wildebeest river
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22 371 crossings). Increasing anthropogenic change could result in greater potential for mistakes and
23
24 372 greater cost to collective migrants.

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26 373 Although fishes similarly showed no significant interactions between extinction risk and
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28 374 collective movement, non-collectively moving freshwater and diadromous fishes were more
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30 375 likely to be placed in higher threat categories than the rest of the migratory fishes, once again at
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32 376 least partially suggesting some benefit to moving collectively.

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34 377 Finally, while coding for collective movement we found that species which tend to move
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36 378 in social groups were typically phylogenetically related (electronic supplementary material Table
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38 379 S6). The interrelatedness of collectively moving species could be a confounding factor in this
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40 380 analysis, potentially obscuring the importance of collective navigation.

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42 381 *Geographic range*

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44 382 Restricting movement, particularly access to key resources, also endangers migrations
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46 383 [11]. As with previous results, the literature review on migratory species geographic range
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48 384 suggests that the type of range change differs between taxonomic groups. Insects (i.e. butterflies)

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3 385 mostly exhibited expansions in range (Fig. 3), which have largely been attributed to increasing
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5 386 temperatures in temperate areas such as Europe, where many such studies are concentrated [46].
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7
8 387 The majority of studies on bird ranges only documented overall shifts, likely because many bird
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10 388 studies were not able to track bird populations across their entire range, thus limiting their ability
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12 389 to conclude whether an absence in one area represented a contraction or simply a change in
13
14 390 geographic location. Shifts in migratory birds could be due to a variety of factors, including
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16 391 habitat loss and climate change [47]. The contractions in migratory mammals seem to support
17
18 392 the notion that while migratory mammals are faring better than non-migrants in respect to
19
20 393 population change, threats to the functional integrity of mammal migrants may still exist. Both
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22 394 freshwater and marine fishes exhibited contractions, with the contractions in freshwater fishes
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24 395 largely due to damming of rivers [48] and marine fish contractions likely caused by intense
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26 396 fishing pressure extirpating even highly mobile species from parts of their range [49].
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31
32 397 It is interesting to note that aerially mobile migratory groups (i.e. insects and birds),
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34 398 experienced many expansions and shifts while the groups restricted to land and water movement
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36 399 (i.e. mammals and fishes) appeared to experience more contractions in range. The energetic cost
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38 400 of moving on land and the potential for barriers in aquatic and terrestrial systems might decrease
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40 401 the ability of mammals and fish to respond to anthropogenic changes relative to aerially mobile
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42 402 species [15]. We posit that these patterns support the idea that locomotion on land or in water
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44 403 might expose migratory species to increased threat from anthropogenic habitat fragmentation
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46 404 (e.g. roads, dams, deforestation), but other correlates need to be considered to make such links
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48 405 definitive.
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53 406 *Caveats*
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3 407 We draw attention to four key limitations inherent to these analyses. First, we highlight
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5
6 408 the recent nature of the abundance comparisons (population trends were examined relative to a
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8 409 baseline of 1970), which would not detect population declines for species that were negatively
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10 410 impacted prior to 1970 (e.g. marine mammals). Second, the Living Planet Index relies on
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12 411 aggregated publicly available data and is therefore prone to biases associated with over-
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15 412 representation of certain taxa and regions [50], specifically toward temperate regions (electronic
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17 413 supplementary material, tables S4 and S5). We analysed the LPD data in both aggregate and
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19 414 within disaggregated regions, but were unable to use the diversity-weighted approach normally
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21 415 employed to account for such biases [50] due to availability of data and species numbers. While
22
23 416 we fully agree there are biases and pitfalls to this approach, previous analyses of these biases
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25 417 indicate that even an unweighted index is likely indicative of overall patterns [24]. While the
26
27 418 LPD does contain trends on commercially important species, the impact of these is reduced
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29 419 through the inclusion of harvest or catch data only if they include a measure of effort. Third, our
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31 420 study purposefully does not thoroughly examine correlates for extinction risk, as many studies
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33 421 have already usefully done [13,30]. We focused on patterns and trends in migratory species,
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35 422 necessitating further study to conclusively identify mechanisms behind observed patterns of
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37 423 vulnerability and extinction risk for migrants. Fourth, GROMS does not comprehensively cover
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39 424 all migratory species. GROMS includes a total of 4,472 records, of which approximately 3,400
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41 425 migratory species are also in IUCN.

426 *Conclusions*

427 In the instances where migration is indeed under threat, we risk losing important
428 ecological functions and charismatic biological phenomena. From these analyses, it is clear that
429 migration vulnerability is dependent on complex interactions between behavioural traits,

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3 430 taxonomy, and the environmental system through which the species navigates. Species with a
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5 431 high potential for collective navigation might be able to more efficiently avoid threats in certain
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8 432 environmental systems (i.e. marine birds, freshwater and diadromous fishes), but collective
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10 433 navigation could amplify the vulnerability of declining species, such as migratory terrestrial
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12 434 mammals in the Palearctic. Migratory bird populations are disproportionately vulnerable
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15 435 compared to non-migrants, whereas migratory mammals are vulnerable than non-migrants.
16
17 436 However, the overall pattern of high extinction risk in mammals, coupled with the high
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20 437 proportion of range contractions experienced by migratory mammals, is concerning. Spatial
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22 438 distribution of migratory mammals, rather than population declines, might be more relevant to
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24 439 preserving their migrations. While there is certainly cause for concern, some of our results
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26
27 440 suggest possibilities for recovery, namely: recent increases in certain migratory mammals,
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29 441 potential benefits of collective navigation, and remaining high abundances of birds despite
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32 442 declines.

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34 443 Of note, if we aim to recover the functional significance of migration, we cannot simply
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36 444 focus on increasing the population sizes of migratory species, but need to ensure that these
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38 445 important increases are matched with recovered ecological potential for these larger groups to
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40 446 meaningful execute migration. In our increasingly fragmented world, the latter goal perhaps is
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43 447 more challenging than the former. The broader patterns reported here suggest both an
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46 448 opportunity and a roadmap for mounting strategic interventions to protect this ecologically,
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48 449 socially, and economically important part of the behavioural portfolio of life.

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53 451 **Additional Information**

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56 452 **Acknowledgements**

1
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3 453 We thank Francis Joyce and Michael Kapuscik for assistance with data synthesis and
4
5
6 454 computation. We thank the two anonymous reviewers for their constructive comments on an
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8 455 earlier version of the manuscript.
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11 12 13 457 **Data Accessibility**

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15 458 All IUCN, geographic range data and LPD model outputs and synthesized information used in
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17 459 this project are included as supplementary material. LPD raw data can be found at
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20 460 <http://www.livingplanetindex.org/>.

21
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23 24 462 **Author Contributions**

25
26
27 463 MHM, DJM, GCT, ED, KS, HSY, MK, AB, AMK, DO, LFH designed the concept of the study.
28
29 464 MHM, GCT, EMD, KS, MK, AB, AMK, DO, LFH found, synthesized, and interpreted data. RF
30
31 465 and SD provided, analysed, and helped interpret all LPD data. GCT analysed IUCN data. EMD
32
33 466 and KS analysed geographic range data. MHM wrote the manuscript with input from DJM, and
34
35
36 467 all authors contributed to revisions.
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39 468

40 41 469 **Competing Interests**

42
43 470 We have no competing interests.
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46 471

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27 **Tables**

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30 **Table 1.**

Taxonomic Group	Environmental System	Odds Ratio	Standard Error	Z score	P-value
Birds	All	1.317	0.11 (1.12 – 1.55)	-4.10	< 0.001
Mammals	All	0.668	0.10 (0.496 – 0.901)	2.15	0.032
Fishes	Freshwater	1.173	0.15 (0.91 – 1.51)	-1.37	0.171
	Marine	0.756	0.15 (0.51 – 1.13)	1.256	0.210

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612 **Table 2.**

Analysis of Deviance for best model					Pairwise comparisons of groups		
Taxa	Predictor	Df	Chisq	P	Collective Migrant	Environmental System	Group
 Birds	System	3	448.48	< 0.001	No	Terrestrial	a
					No	Terrestrial/Freshwater	abc
					Yes	Terrestrial/Freshwater	ab
					Yes	Terrestrial	ab
					Yes	Terrestrial/Freshwater/Marine	b
					No	Terrestrial/Freshwater/Marine	abcd
					Yes	Terrestrial/Marine	c
No	Terrestrial/Marine	d					
 Fishes	System	2	79.402	< 0.001	No	Marine	a
					Yes	Marine	ab
					Yes	Freshwater/Marine	abc
					Yes	Freshwater	bc
					No	Freshwater	c
No	Freshwater/Marine	c					
 Mammals	Class	N/A			All levels equivalent		
Overall extinction risk	Class	2	1023.6	<0.001	Birds		a
					Fishes		a
					Mammals		b

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616 **Figure and table captions**

617 **Fig. 1.** Abundance indices measured over time relative to 1970 baseline values for non-migratory
618 (nm; orange) and migratory (mig; green) species derived from the Living Planet Database. The
619 data in this figure represent a compilation of population time series from monitored sites around
620 the world. Data are (A) aggregated across all birds, mammals, and fishes, and subdivided by
621 taxon (B-D). Number of species included in analyses are provided on each figure.

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3 623 **Fig. 2.** Regional abundance indices measured relative to 1970 for data rich northern temperate
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5 624 terrestrial and aquatic species in either the Palearctic or Pacific realm. Data is derived for non-
6
7 625 migratory (nm; orange) and migratory (mig; green) species from the Living Planet Database. The
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9 626 data in this figure represent a compilation of population time series from monitored sites around
10
11 627 the world. Graphs depict change in abundance indices for A) Palearctic terrestrial birds, B)
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13 628 Palearctic freshwater birds, C) Palearctic terrestrial mammals, and D) Pacific marine fishes.
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15 629 Number of species included in analyses are provided on each figure.
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22 631 **Fig. 3.** Proportion of types of migratory range changes reported for migratory insects (i.e.
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24 632 butterflies), birds, mammals, and fishes. Range changes can be contractions, shifts (i.e. change
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26 633 in geographic location), expansions, or no change. Width of x-axis represents relative sample
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28 634 size of taxonomic groups. The number of records of change for each taxonomic group is
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30 635 reported.
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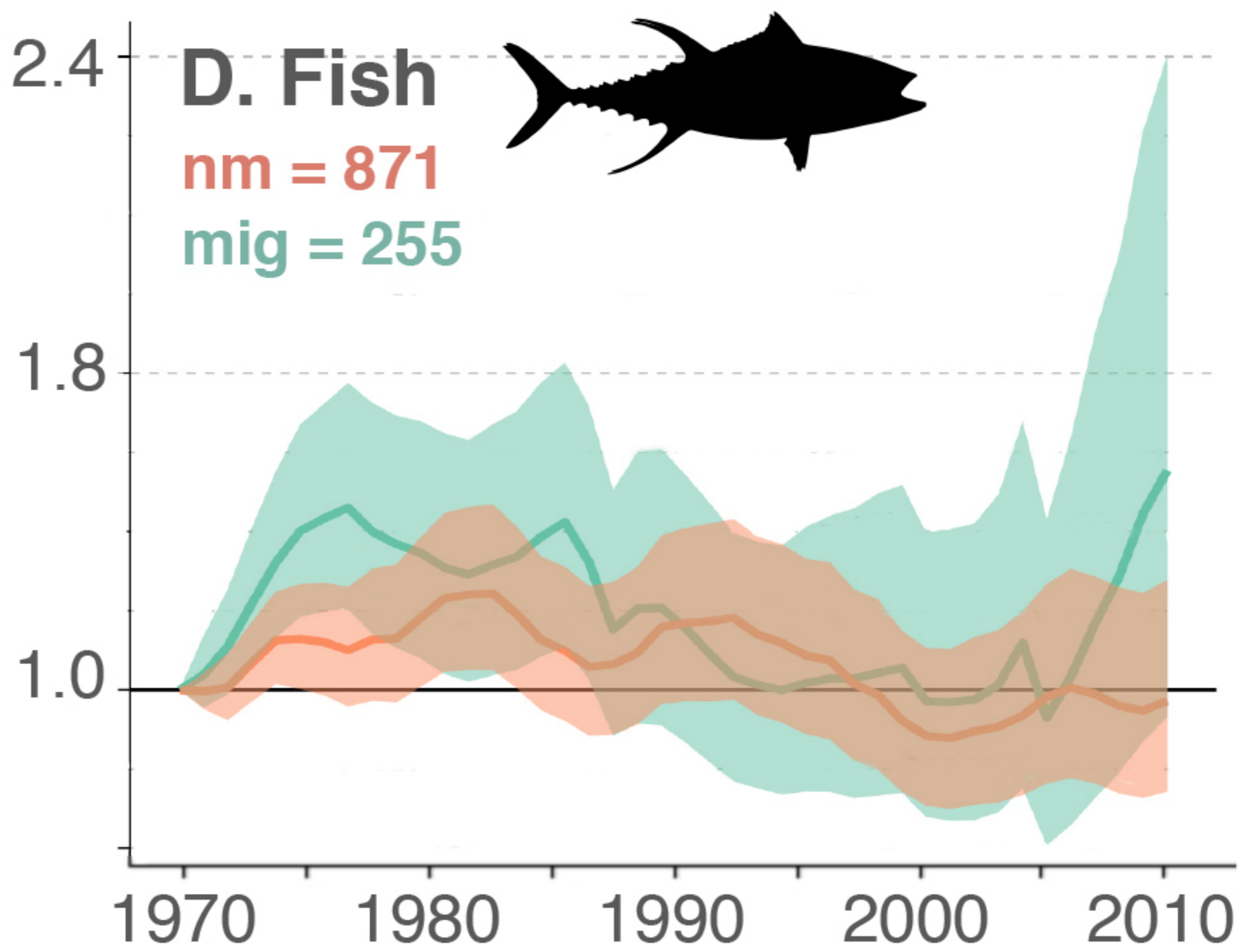
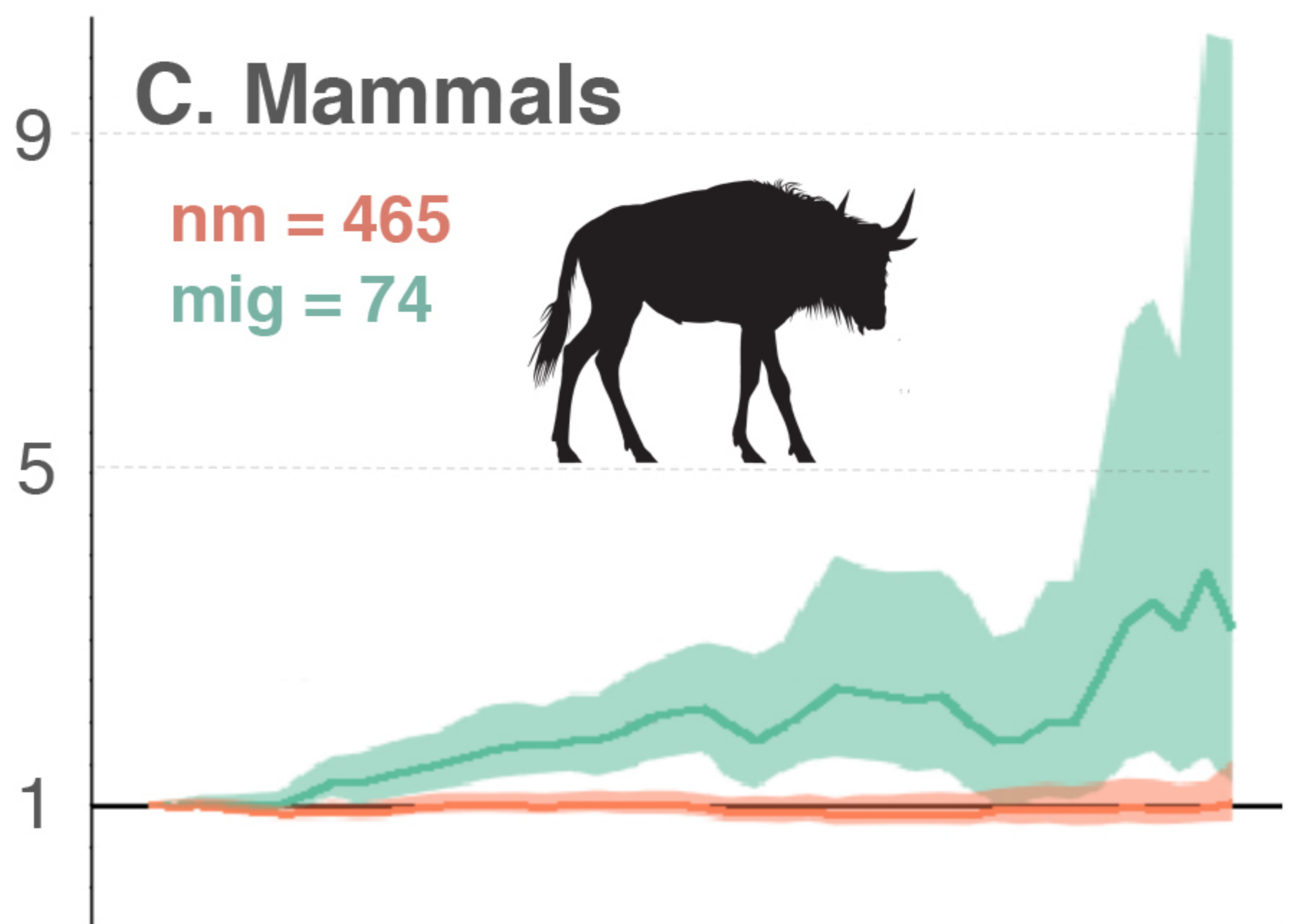
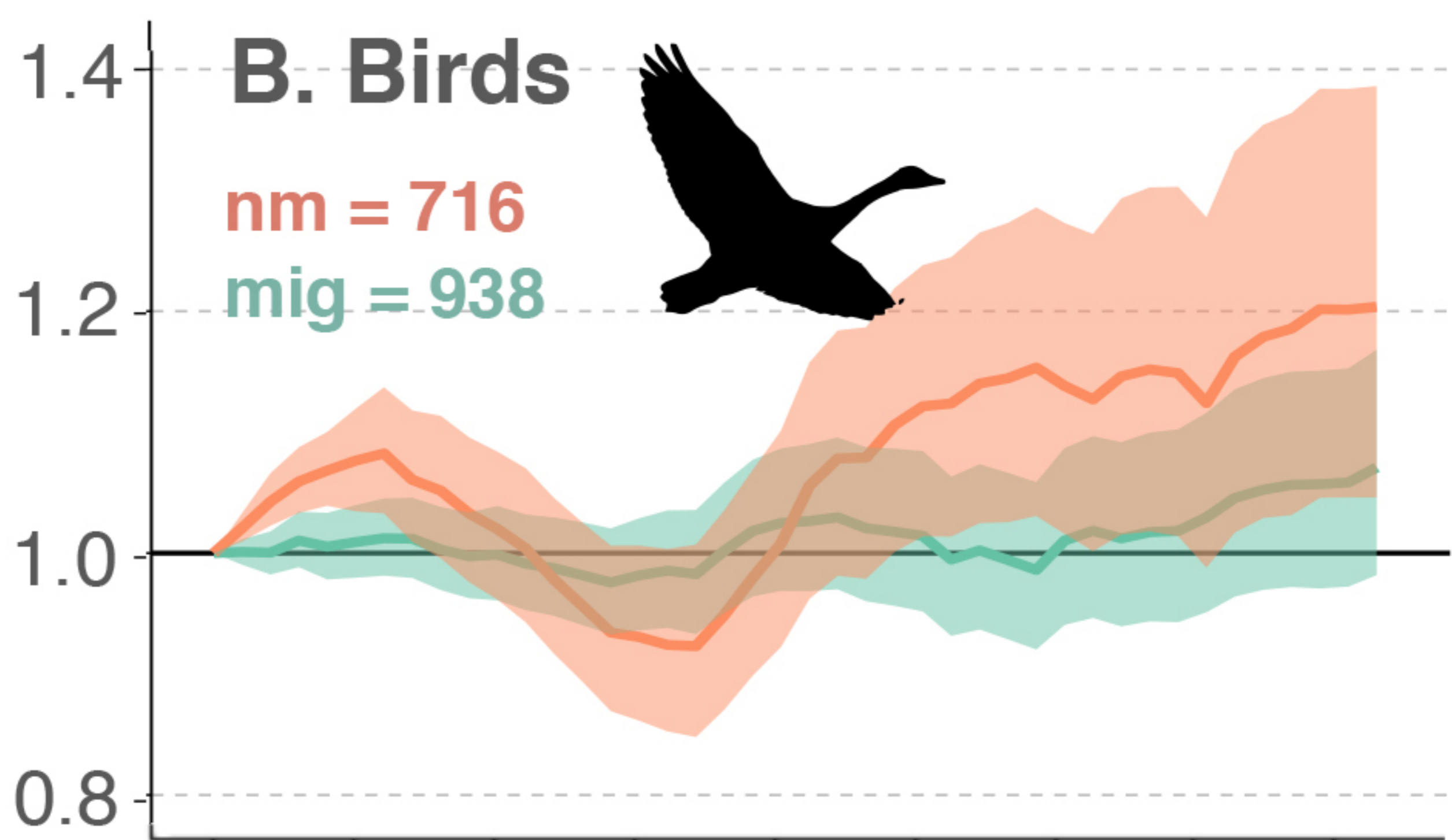
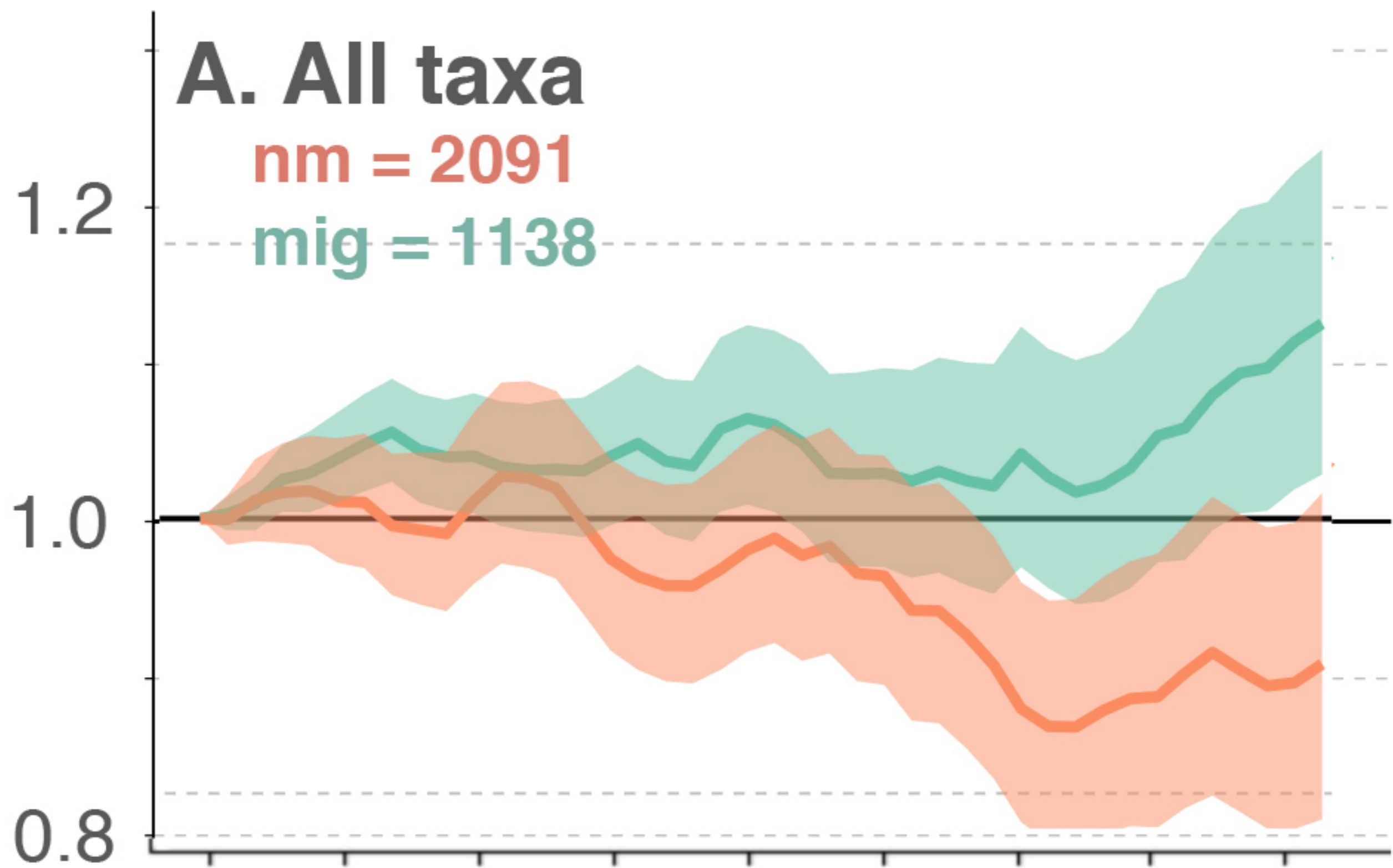
34 636
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36 637 **Table 1.** Model results from Living Planet Database likelihood of annual increase analysis for
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38 638 birds, mammals, and fishes. Odds ratio, standard error with confidence intervals, Z score and P-
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40 639 value are reported for each taxonomic group, with fishes split into freshwater and marine
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42 640 systems.
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48 642 **Table 2.** Analysis of Deviance (ANODE) table and Tukey-adjusted pairwise comparisons of
49
50 643 model coefficients for all IUCN extinction risk analyses. ANODE results shown at left
51
52 644 correspond to the best model of extinction risk for each analysis. Pairwise comparisons for group
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54 645 differences in each best model are shown on the right. Species were coded as collective migrants
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3 646 based upon database or literature indicating they travel in social groups. Environmental system
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6 647 (freshwater, marine, terrestrial, or their combination) was coded from IUCN classifications. The
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8 648 mammal model showed no difference from the null and thus comparisons were not possible.
9
10 649 Group levels denote significant differences from pairwise comparisons within each dataset for
11
12 650 birds, fishes, mammals, and all taxa. Lower lettering (i.e. “a”) indicates a lower risk of extinction
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14
15 651 than higher lettering “b-d”. Least-square means and asymptotes of their 95% confidence limits
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17 652 were used to differentiate groups, but their numeric values are arbitrary; these are listed in Table
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20 653 S9.
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LPI Index (1970=1)



1970 1980 1990 2000 2010

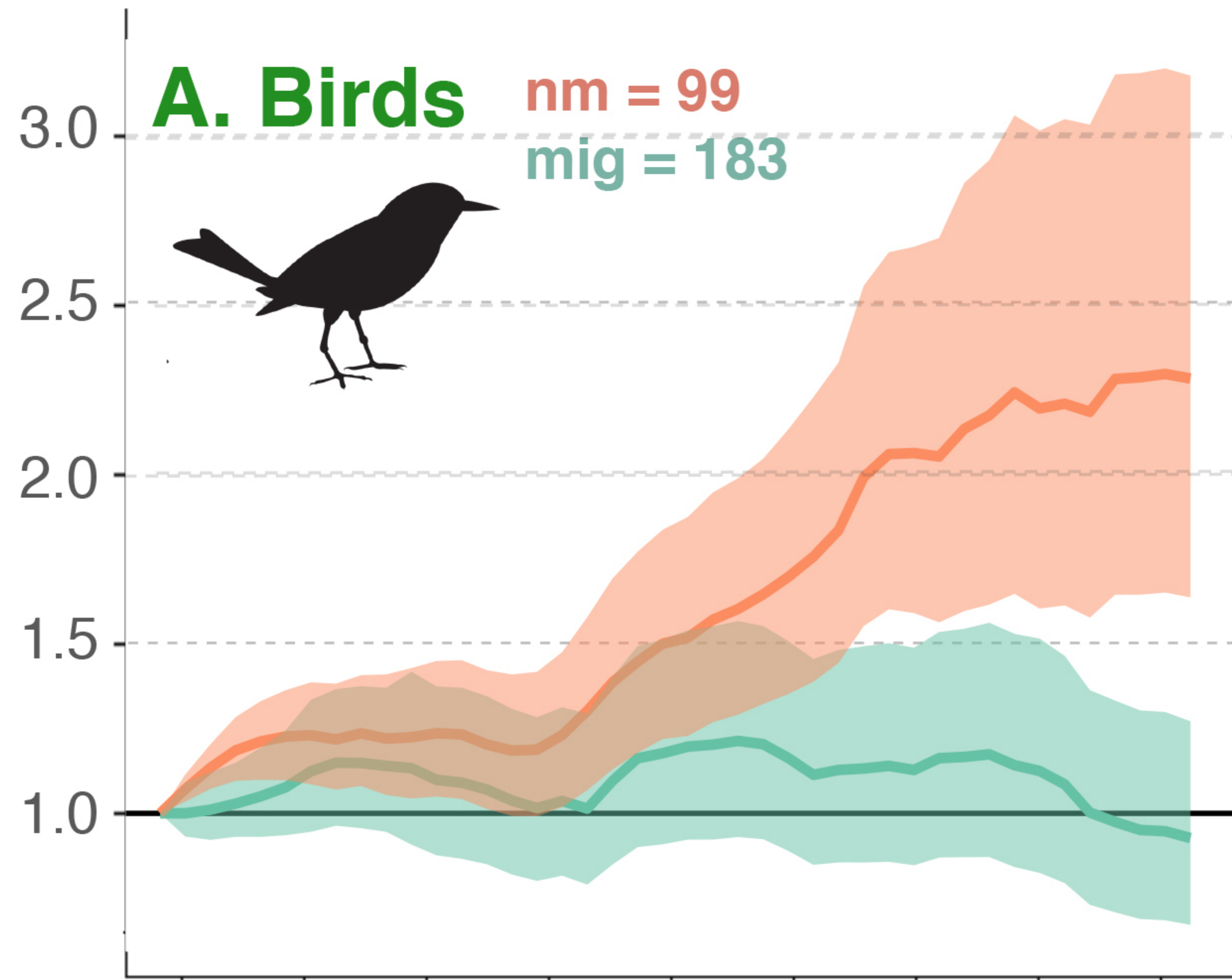
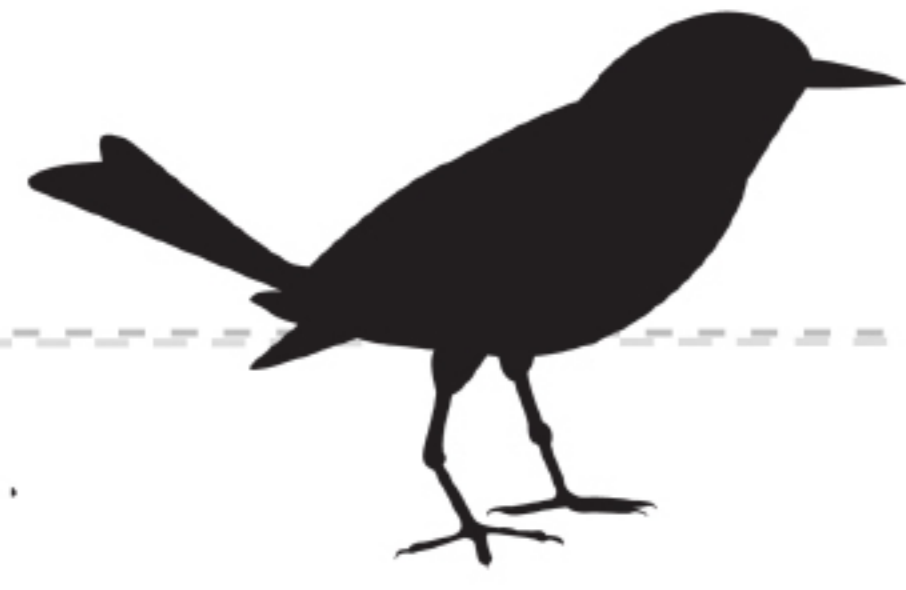
Year

Terrestrial

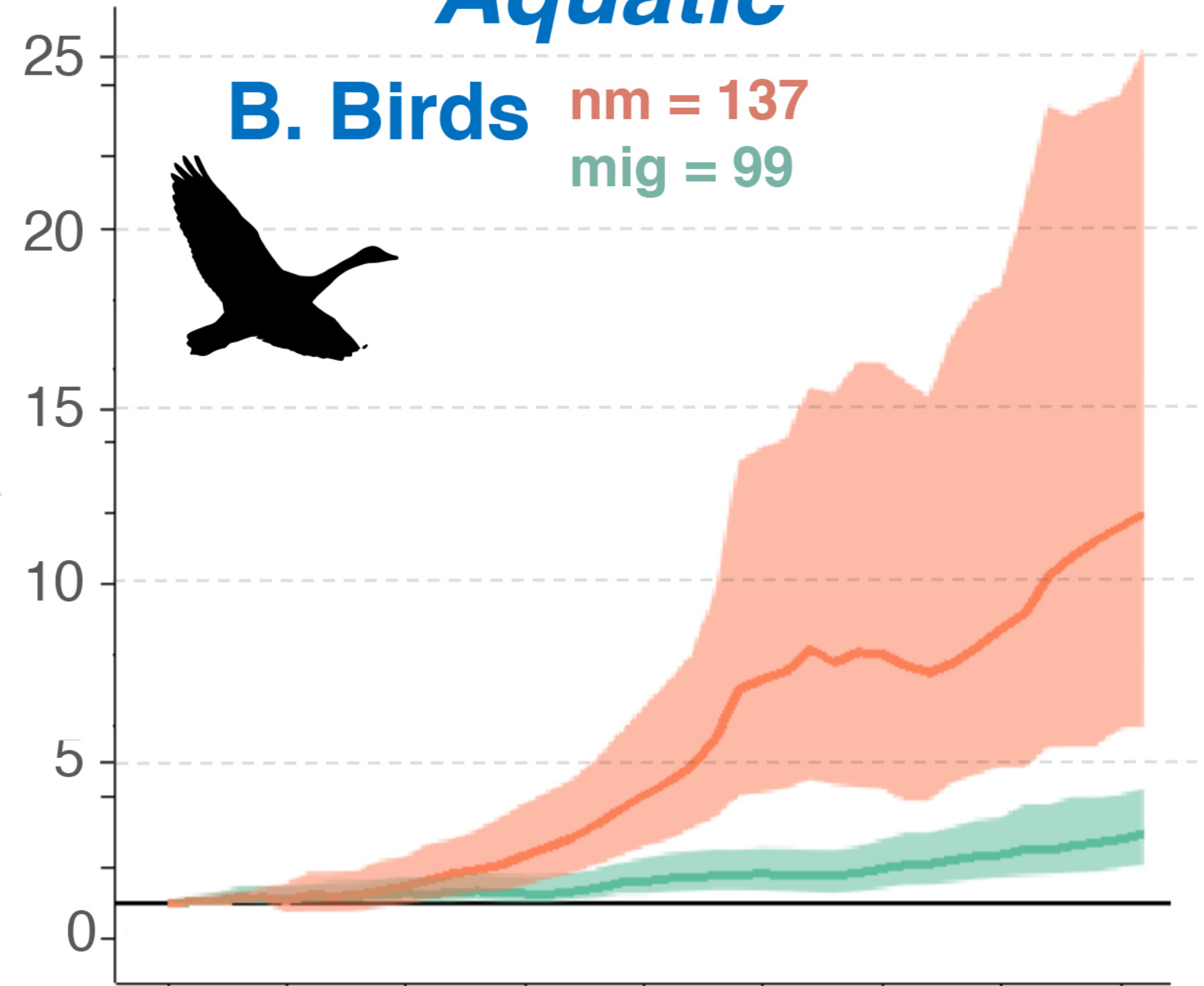
Aquatic

LPI Index (1970=1)

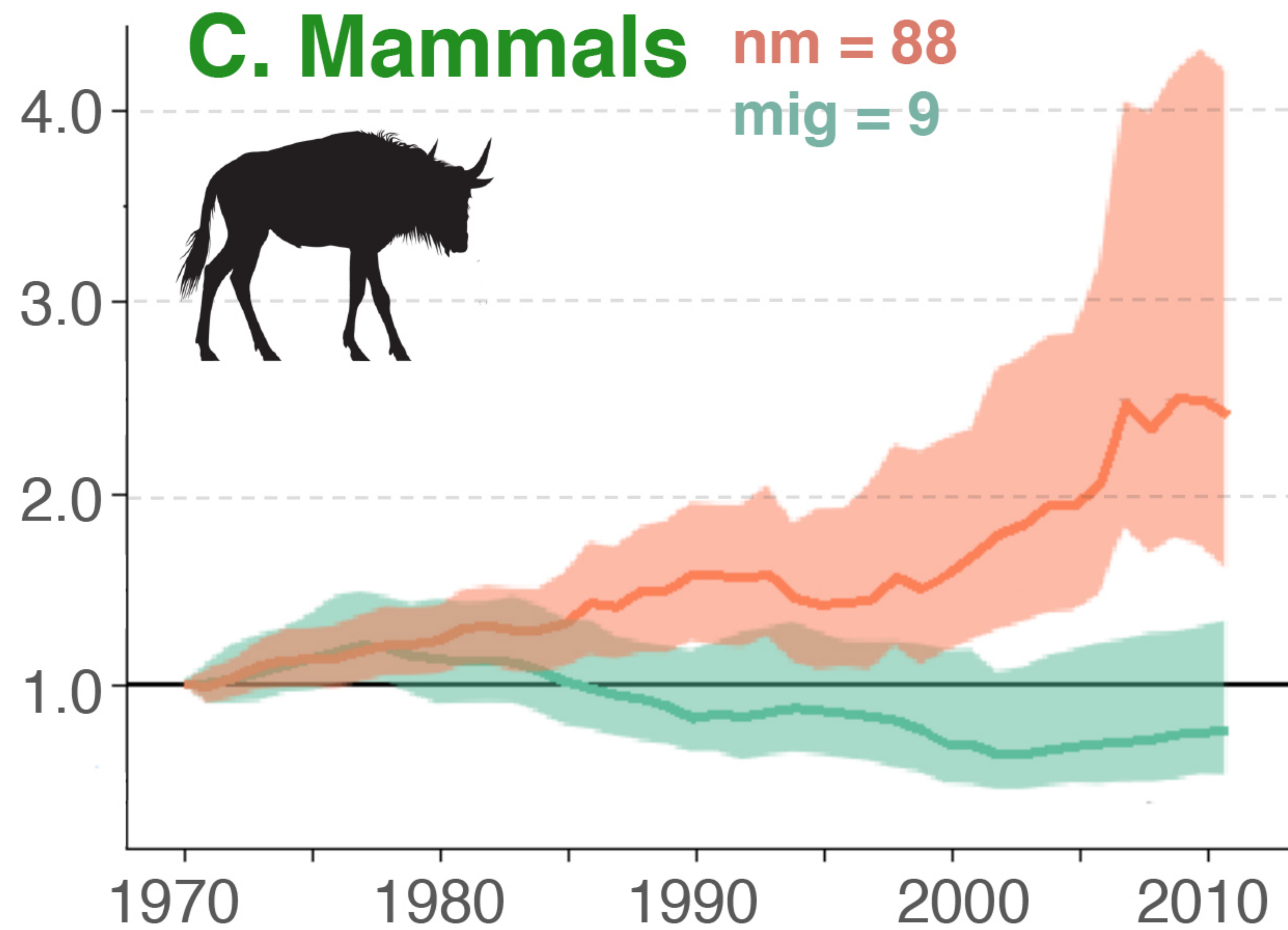
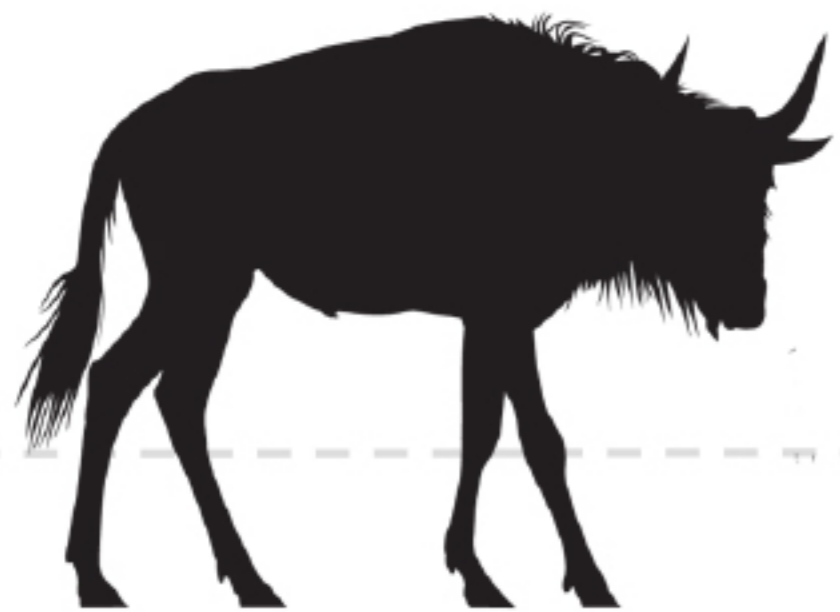
A. Birds nm = 99
mig = 183



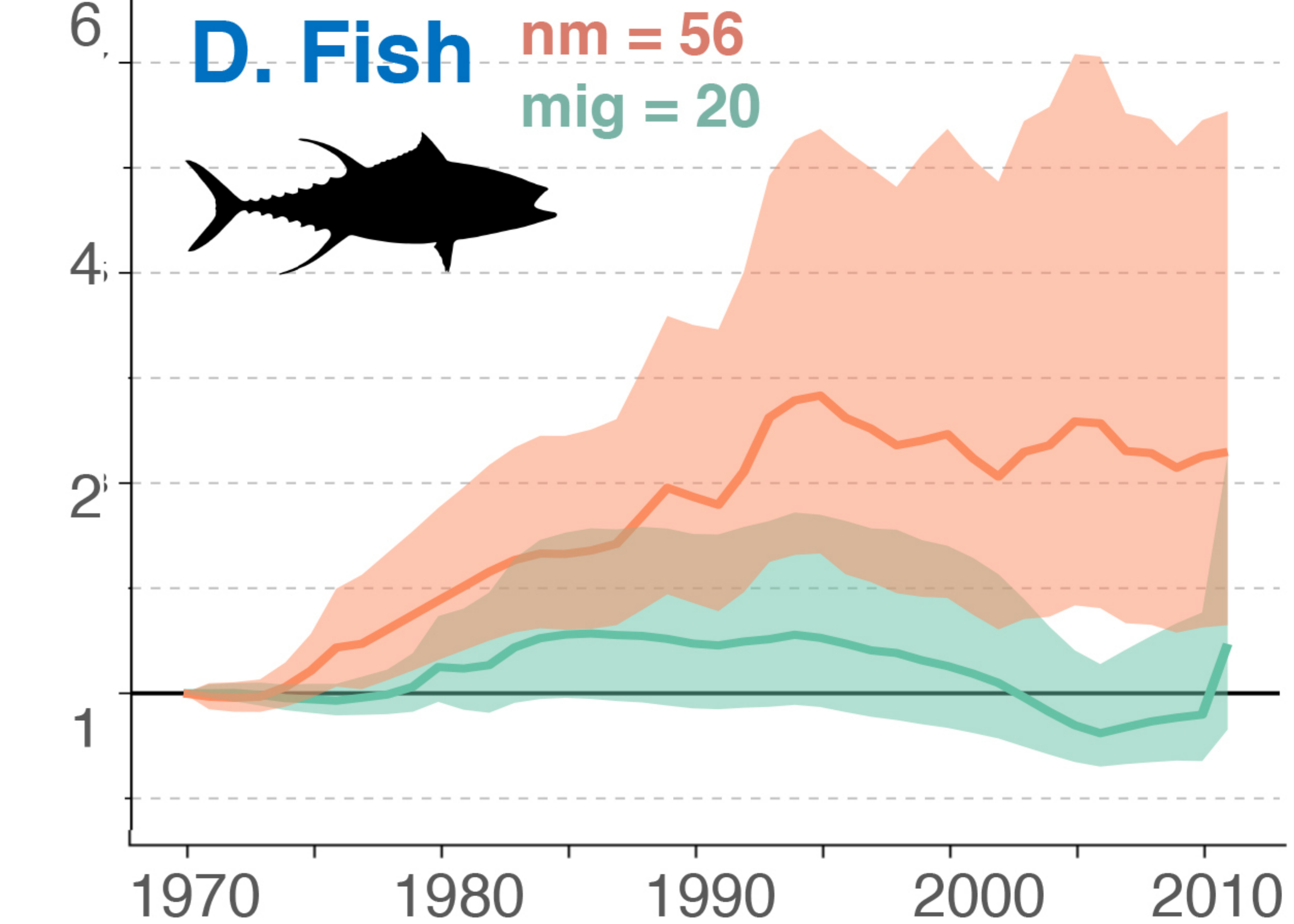
B. Birds nm = 137
mig = 99



C. Mammals nm = 88
mig = 9

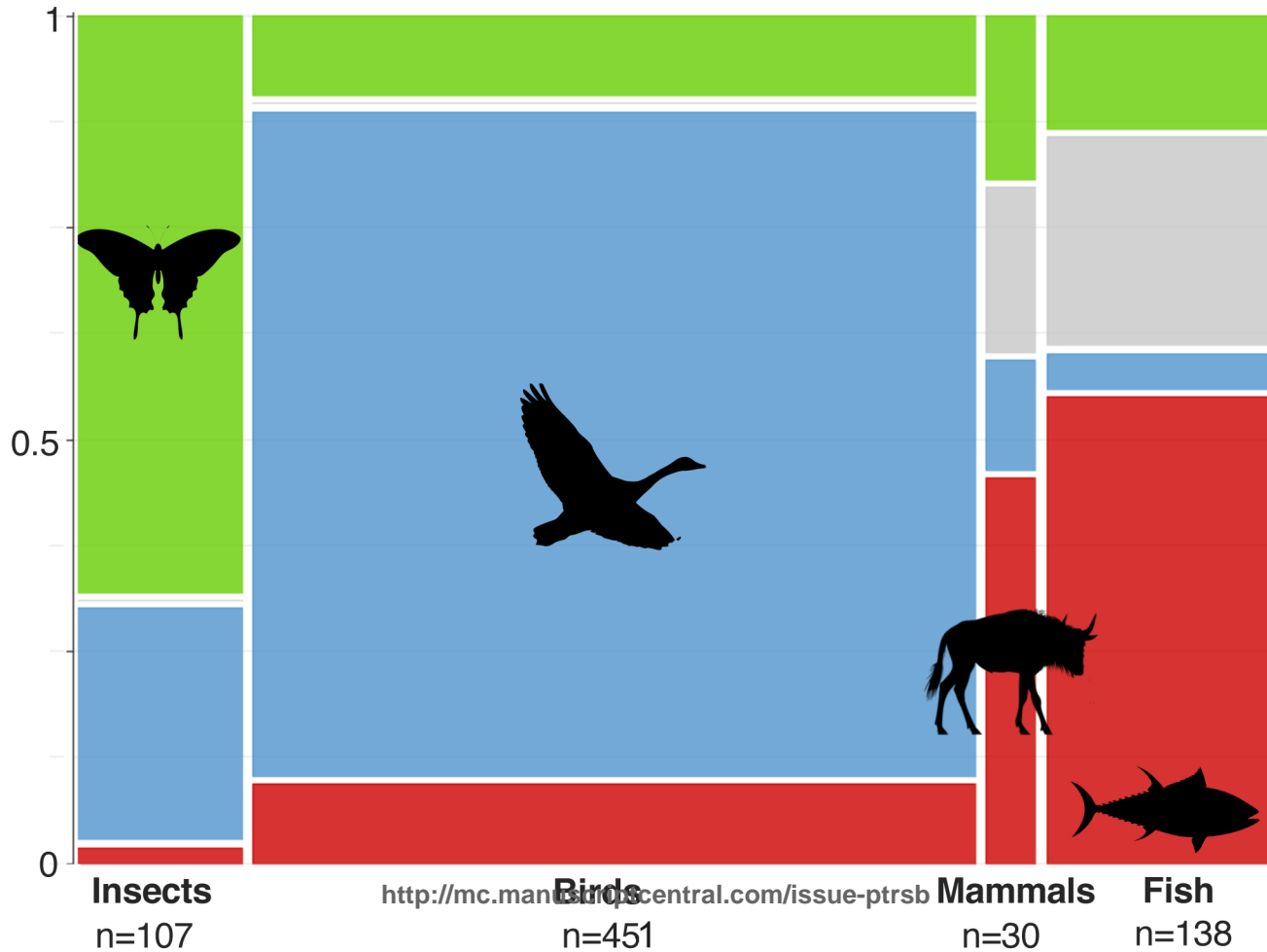


D. Fish nm = 56
mig = 20



Year

Submitted to Phil. Trans. R. Soc. B - Issue
Contraction Shift No change Range Expansion



Insects
n=107

Birds
n=451

Mammals
n=30

Fish
n=138

<http://mc.manuscriptcentral.com/issue-ptrsb>

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