

1 **Abstract**

2 Humans are increasingly viewed as active agents of environmental and land cover change in the
3 moist Neotropics. While the scale and extent of pre-Columbian anthropic impacts are actively
4 debated, the effects of post-Contact patterns of land use are rarely examined over the long term,
5 defined here as centennial timescales. This article examines a putative area of historical low human
6 impact located in the western Guiana Shield, the upper Cuao River, using an exploratory agent-
7 based modelling approach. Based on an extensive ethnographic literature on the Piaroa, who have
8 inhabited the region for at least four centuries, the model investigates the legacy effects of
9 ethnographic patterns of land use in the interval between European Contact and the present.
10 Model outcomes indicate that the potential range of anthropic changes to the environment of the
11 study area are significantly greater in scale than previously assumed. Interpretative discrepancies
12 between present vegetation conditions and the model are likely the product of sparse
13 palaeoecological and archaeological research in the upper Cuao. More broadly, the results imply
14 that small-scale agriculture and agroforestry can lead to extensive and persistent structural changes
15 to ecosystems in relatively short timescales. The experiment bolsters existing cautions against
16 assuming the “natural” baseline of Neotropic forests based on present appearance. As a form of
17 middle-range theory, the model demonstrates how computational approaches can promote closer
18 integrations between ecological, archaeological, and ethnohistorical data, as well as frame the
19 expectations of future research.

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21 *Keywords: Agent-based modelling, Orinoco, Amazonia, simulation, Anthropocene, swidden agriculture*

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27 **1. Introduction**

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29 The impact and legacy of indigenous land use practices continues to be a contentious area of
30 research in the Neotropics (McMichael et al., 2012; Ter Steege et al., 2013; Stahl, 2015; Piperno et
31 al., 2015; Watling et al., 2015; Watling et al., 2017; Piperno et al., 2017). Current debates in
32 archaeology and historical ecology have moved on from dichotomous “cultural parkland” versus
33 “pristine forest” scenarios (Denevan, 1992a; Heckenberger et al., 2003; Meggers, 2003), and focus
34 now on defining the scale and extent of environmental modifications effected by the human
35 inhabitants of Amazonia (Mayle and Iriarte, 2014; Clement et al., 2015). Critical readings of the
36 ethnohistorical record frequently inform discussion on these topics, either to contrast with
37 archaeological data, or to serve as a point of departure for useful comparisons between observed
38 and archaeologically-inferred behaviours (Whitehead, 1998; Stahl, 2015). Crucially, the
39 “wilderness” described in early European records of the Neotropics is now recognized to be, in
40 many cases, an artefact of a post-Contact transition to an anthropogenic, yet largely human-free,
41 environment that bears little resemblance to the pre-Columbian period (Denevan, 2016; cf.
42 McMichael et al., 2012). Although colonization had disastrous consequences, which left
43 measurable signals in the palaeoenvironmental and archaeological records (Faust et al., 2006; Bush
44 et al., 2008), total demographic collapse and transition to historical patterns was not the only
45 possible outcome (Ramenofsky, 1987; Thornton, 1987). The following examines the impact of
46 historical indigenous land use in the interval between Contact and the present, in order to refine
47 hypotheses and theoretical expectations for further field research

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49 This study identifies the *long-term* (centennial) consequences of historical indigenous land use as an
50 area in need of scholarly attention, particularly where ethnohistorical, palaeoecological, or
51 archaeological data are sparse. Short-term research on swidden agriculture (defined here as sub-
52 generational) has for decades formed a programmatic part of Tropical forest research in ecology,

53 geography, and archaeology (Conklin, 1961; Harris, 1971; Cairns, 2015). Within the context of the
54 European colonization of the Americas, suggested as one potential “tipping point” marking the
55 onset of the Anthropocene (Erlandson and Braje, 2013), swidden agriculture emerged as the
56 dominant agrarian subsistence strategy among post-Contact indigenous groups. This occurred
57 under a particular set of internal and external societal pressures, as one strategy of many along a
58 spectrum of possible responses by indigenous groups (Arroyo-Kalin, 2012; Balée, 2013: 185;
59 Neves, 2013; Mayle and Iriarte, 2014). Swidden farming was also likely precocious among the first
60 tropical forest cultivators, certainly making it a part of the complex legacy of human inhabitation
61 in the Neotropics (Piperno and Pearsall, 1999; Oliver, 2001; Kirch, 2005; Roosevelt, 2013; Kawa,
62 2016). If the (circum-) Amazonian world is to be represented in the Anthropocene debate, it
63 becomes necessary to be able to define the scale of historical indigenous environmental impacts
64 in precise terms.

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66 In parallel to historical ecology, yet largely isolated from it, is the growing adoption of simulation
67 and model-based inferential frameworks in archaeological science (Kohler and van der Leeuw,
68 2007; Lake, 2014; cf. Johnson et al., 2005), including studies of coupled human-environmental
69 systems and land cover/land use change (LUCC) in the past (Wilkinson et al., 2007; Barton et al.,
70 2010; Barton, 2014). Driven by questions about human environmental impact, ecologists and
71 geographers increasingly deploy computational approaches for exploring complex interactions
72 between agents and ecosystem processes (Parker et al., 2003; Perry and Millington, 2008).
73 Exploratory agent-based modelling (ABM), using known behaviours as a point of departure, are
74 ideal for testing intuitions or hypothesis about a target system (Davies, 2016: 83). Computationally-
75 minded archaeologists have already noted the synergy between experimental simulation on one
76 hand (Godfrey-Smith, 2006; Potochnik, 2012) and middle-range theory in archaeology on the
77 other (Premo, 2007; Davies, 2016). Simulation is perhaps best conceptualized as a “virtual
78 laboratory” (Magliocca and Ellis, 2016) that serves to formalize assumptions about the operation

79 of a system of interest, quantify variability in model outcomes and, ultimately, test these against
80 knowledge claims. This is the approach adopted here.

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82 The study aims to refine current understandings of the scale and temporality of impacts caused by
83 interfluvial slash-and-burn (swidden) agriculture in the moist Neotropics. Clearer insights into the
84 legacy and effects of post-Contact patterns of swidden agriculture is crucial to the agenda of
85 historical ecological research, and Anthropocene studies more broadly. The following synthesizes
86 a significant body of ethnographic work from the Orinoco, focusing on the Piaroa indigenous
87 group (see Zent, 1992; Overing and Kaplan, 2011), to investigate the variation in ecosystem
88 disturbances under swidden farming systems. This research was carried out in the context of a
89 larger project investigating indigenous occupations along the Middle Orinoco (Lozada Mendieta
90 et al., 2016). The model generalizes nearly half a century of anthropological and historical research
91 on Piaroa subsistence strategies, however, it does so usefully and explicitly (Epstein, 2008). I argue
92 that the lack of archaeological and environmental field research in the study area presents a gap in
93 knowledge which ethnographically-informed computational models may be particularly suited to
94 tackle.

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96 2. Background to model: Piaroa interfluvial subsistence

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98 The Piaroa occupy the western edge of the Guiana Shield, an upland interfluvial zone in what is
99 today southern Venezuela. This area has historically received scant attention from archaeologists
100 and palaeoecologists, yet is rich in ethnographic data. The Piaroa are traditionally highly mobile
101 farmers whose subsistence is heavily reliant on manioc cultivated in short-lived plots of cleared
102 and burned forest, which is supplemented with fishing, hunting, and agroforestry. It is highly likely
103 that, due to disease, conflict, slaving, and new technologies following European contact, the
104 ethnographically-observed settlement pattern is contingent on recent historical processes (such as

105 the gradual spread of metal tools) rather than being the end state of an uninterrupted cultural
106 evolution (Denevan, 1992b; Huber and Zent, 1995: 48; Zent, 1998). Since the rural expansion of
107 the Venezuelan market economy and healthcare programmes in the 1980s, the Piaroa have
108 increasingly settled permanently along the main channel of the Orinoco and its navigable
109 tributaries to access to these services (Mansutti Rodriguez, 1988; Freire, 2007; Freire and Zent,
110 2007). Modes of subsistence recorded in the highland interfluvial zones, now considered their
111 “traditional” homeland, differ considerably from modern practices. In comparison to the Orinoco
112 floodplain, game is more dispersed, the rivers support fewer fish, and cropping cycles are
113 shortened while fallows are lengthened due to the shallow soils (Huber and Zent, 1995: 50). Still,
114 Zent (1995: 87), citing Gilij (1987), considers the Piaroa to have maintained stable settlement in
115 the uplands for 400 years at a minimum. As contact before the eighteenth century was sporadic
116 and archaeological data are non-existent in the interfluvial zone, this is an estimate for the duration
117 of occupation (Mansutti Rodriguez, 1988: 9; Zent, 1992: 48; Overing and Kaplan, 2011: 506).
118 Sáliva-speaking ancestors of the Piaroa were present in the upper-Middle Orinoco before this, but
119 the circumstances and duration of pre-Columbian settlement here is unknown at present.

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121 Upland Piaroa settlements in the catchment of the Cuao River, a black water tributary of the
122 Orinoco (Figure 1), are typically organized around a single large house (Piaroa: *isode*). This valley
123 forms the focus of the present study. Isodes are co-inhabited by an extended family or household
124 group of several extended families, consisting of 16-50 individuals (1-10 nuclear families), although
125 actual settlement population fluctuates somewhat due to high individual mobility. Camping trips,
126 visits, and temporary relocation can take place in a given year, with the home isode and its extended
127 family group acting as an anchor (Zent, 1998: 255). Overall population density is very low, with at
128 most 10 isodes per 100 km² “neighbourhood” (Monod, 1970; Overing and Kaplan, 2011). Isodes
129 can reportedly be occupied for as little as one year before relocation, or for 10 years or more (Zent,
130 1992; Heckler, 2004: 204).

131

132 [Insert Figure 1 here]

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134 Although fishing, hunting, and wild resource gathering feature in the diversified Piaroa subsistence
135 base, significant time and effort is invested in the swidden agricultural system, characterized by a
136 short cropping phase (six years or less) followed by long fallows of 20 years or more (Zent, 1992:
137 184). Manioc (*Manihot esculenta*) and maize (*Zea mays*) together comprise nearly 80% of a harvest
138 and supply the bulk of Piaroa caloric needs, with other wild and cultivated foods providing mainly
139 supplementary nutrition (Zent, 1992; 1998; Overing and Kaplan, 2011: 514). Swidden plots, always
140 cleared by slashing and burning climax forest, range in size from 0.25 to 6 ha, a figure that varies
141 based on the number of individuals a plot will support. A nuclear family will generally self-sustain
142 with a single swidden plot. A survey of 178 swidden plots in the upper Cuao corroborates this
143 (Figure 1, bottom right), and shows a notable skew towards such plots (<2 ha in area, Perera, 1986;
144 Zent, 1992: 190). When the productivity of a coeval group of ageing fields drops off due to
145 vegetation succession and weed growth, members of an isode will relocate to the site of a former
146 plot and establish a new settlement. Altering the forest cover exposes the thin tropical soils to
147 depletion or erosion, in particular if secondary regrowth is cleared again after an insufficiently long
148 fallow (<20 years) (Fölster, 1995: 70). The viability of subsequent cropping cycles and fallows will
149 be severely curtailed if this occurs. Similarly, maintenance of swidden plots for longer than a few
150 years risks limiting the regeneration potential of subsequent regrowth to bush fallow (Eden, 1974:
151 48). These pathways are typically avoided if at all possible, since fallowed plots and secondary
152 forest both supply ecosystem services through improved hunting and gathering grounds (Harris,
153 1971; Zent, 1995). To summarize, Piaroa swidden farming in the upper Cuao is an interplay
154 between meeting subsistence requirements, monitoring existing plots, and managing wild
155 resources effectively, while avoiding intensive localised pressure on forest cover and overtaxing a
156 limited pool of labour (Mansutti Rodriguez, 1988; Zent, 1992).

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Modelling Piaroa subsistence practices and forest succession in the upper Cuaio allows some important aspects of post-Contact land cover and land use to be discussed. The implementation of a small number of conservative and straightforward decision-making routines and parameters is based on reasonable assumptions and derivations from the ethnographic literature. For instance, the model uses a static or declining proxy for population, and swidden plots provide consistent returns over their period of cultivation. The world initializes as a homogenous primary forest, in which site selection is mostly unconstrained and guided only by economic logic insofar as it impinges on farming. It slightly underestimates starting population density as roughly 1 isode per 118 km², below that estimated by Monod (1970), and does not consider isode migration or fission-fusion dynamics. These serve to usefully constrain the number of factors considered in the parameter space, and focus the discussion of the output. Rather than precisely mirroring every aspect of Piaroa settlement patterns and subsistence behaviour, the model provides a reasonable approximation, which allows for inferences to be made on long-term human-environmental interaction.

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173 3. Model overview and design

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The above account of Piaroa socioecology is implemented computationally through a coupled agent-based and cellular automata (CA) model based on spatial grid cells in NetLogo 6.0.1 (see Wilensky, 1999; Moreno et al., 2007; Mena et al., 2011; Luus et al., 2013), to create an abstraction of observed behaviour capable of investigating its long-term operation. The model parameters are summarized in Table 1. For the study, individual isodes are treated as agents, with population per isode generalized to the nuclear family level. At the start of a given run of the simulation, the number of families is randomly assigned a value of 1 – 10, which determines the number of swidden plots an isode will produce after each relocation to meet requirements. Isodes are

183 otherwise identical and do not interact, except for avoiding “claiming” active swidden plots owned
184 by another isode. As isode relocation is directly linked to the agricultural strategy (new house plots
185 are always established in old fields), a single parameter handles both swidden logistical range and
186 mobility magnitude (parameter 2). During a run, isodes monitor returns from surrounding grid
187 cells in range each time step and relocate to establish new swidden plots (*conucos*) when
188 requirements are not being met by existing fields. Grid cells targeted for swidden clearing always
189 consist of fallow secondary or primary forest. The model assumes that historically observed
190 clearance practices, employing efficient metal tools, are present at the start of a run. Stone axes,
191 for example in a pre-Contact scenario, would require different model assumption as to the time
192 and energy cost of creating clearings (Denevan 1992b). The choice of new house gardens when
193 relocating (parameter 3) can be varied to be a random cell out of the pre-existing swidden plots
194 owned by an isode, or optimized to always be the plot with the least regrowth. If, in any case, there
195 is an insufficient quantity of grid cells to meet requirements in logistical range, an isode may enter
196 a stressed state when relocating and lose a family (to death or outmigration). The likelihood of
197 stress events occurring is controlled by the magnitude of the deficit of available land.

198 [insert Table 1 here]
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200 Vegetation succession and forest transition is handled as an integer by the cellular automata model,
201 mapped onto a simplified emic schema of forest regrowth stages (see Zent, 1992: 331; Zent, 1995:
202 91; Table 2), which functions on a per-grid cell basis. Grid cells regenerate from initial clearance
203 to initial secondary regrowth, with an equal likelihood of adding no growth (zero), little growth
204 (1), or significant growth (2) per time step. Each time step in the model approximates a season,
205 with two seasons per year. This introduces a degree of stochasticity as to when isode relocation
206 will take place, as plots may regenerate unevenly. As grid cells of secondary regrowth provide some
207 returns (a tenth of a swidden plot) in the form of mature tree stands and improved hunting, a

208 sufficient quantity of old gardens in logistical range could theoretically offset the gradual losses of
209 some fields becoming unproductive.

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211 Forest succession (from bush to secondary forest, and from secondary to climax vegetation) are
212 modelled probabilistically (parameter 4). Every time step, grid cells with incipient or mature
213 secondary status (Table 2) count the number of surrounding primary forest cells. Every grid cell
214 of primary forest in the Moore neighbourhood (up to eight neighbouring cells) increases the
215 likelihood that stage succession will occur in increments of 0.025 per cell. The contribution of each
216 cell is varied as a model parameter (see Table 1), giving a range of values from 0.025 (a single
217 primary forest neighbour at the lowest parameter setting) to 1 (eight neighbours with primary
218 forest at the highest parameter setting). This implementation of forest succession approximates
219 the gradual colonization of fallowed plots by pioneer species and the maturation of economic tree
220 species planted by Piaroa at the start of its life cycle. It also reflects the long secondary regrowth
221 period (on average a century) typical of the nutrient-poor and fragile soils of the study area.

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223 [insert Table 2 here]

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225 The landscape employed in the simulations is a 609 by 426 cell digital terrain model of the upper
226 Cua watershed (3568 km²), resampled to half of its original resolution of 90 m to improve
227 processing time. Use of the terrain model (parameter 5) does not affect the behaviour of either
228 isodes or cells, but allows certain results of the parameter sweeps to be visualized more easily. The
229 initial state of the landscape assumes 100% primary forest cover. As noted above, there were
230 undoubtedly legacy effects of settlement in the landscape prior to the Piaroa uptake of swidden
231 agriculture. The present state of knowledge in the study area only allows these to be guessed at
232 and for present purposes, legacy effects have been left out. Each run of the model lasts for 800
233 time steps, equivalent to 400 years, or stops when all isodes have disbanded or expired. This

234 timeframe represents a hypothetical maximum for the presence of ethnographic Piaroa subsistence
235 behaviour in the study area: the seventeenth century to the present. Two sets of runs were carried
236 out separately due to prohibitively long processing times for large numbers of continuously
237 measured runs. The first set measured the end state of the landscape and isodes over 100
238 repetitions of each parameter combination (4500 runs), while the second set measured model
239 outputs diachronically over just 30 repetitions of each parameter combination (1350 runs).

240

241 4. Results

242

243 This study modelled Piaroa land use through the interplay between mobility, population, and forest
244 regeneration potential to investigate legacy effects on the landscape of the upper Cuao. Figure 2a
245 summarizes the effects of isode mobility on the composition of primary forest in the study area.
246 This compares the number of grid cells that are never cultivated (old growth) against the total
247 amount of primary forest (old growth + reverted primary forest grid cells) at the end of 800 model
248 time steps, equivalent to four centuries. Grid cells of reverted primary forest are former swidden
249 plots that have succeeded back to climax vegetation after a century or more of fallow, making
250 them functionally indistinguishable from old growth primary forest to both Piaroa and non-Piaroa.
251 The pattern implies that as logistical radius (parameter 2) increases, the total amount of primary
252 forest remains stationary and very close to 98%, while unmodified primary forest sees a steady and
253 proportional decrease. Isodes in very low mobility regimes (logistical radius ≤ 3) experience
254 frequent stress events due to the local depletion of arable grid cells, with coeval losses of
255 population. Loss reduces subsistence requirements, with the consequence of fewer cultivated grid
256 cells overall, leading to a lower ratio of reverted to old growth grid cells. Probability of regeneration
257 (parameter 3) has a negligible effect on the above outputs.

258

259 The total area roamed by a co-resident group of families is what Zent (1992) would term their
260 "neighbourhood". Forest succession in the upper Cuao is observably affected by the average
261 distance travelled by isodes from their origin, which is a function of logistical radius (Figure 2b
262 and c). Using this as a proxy for isode neighbourhood, changes in the proportion of grid cells with
263 scrubland ("incipient") regrowth and secondary forest both correlate with neighbourhood size.
264 Plotting the proportion of secondary regrowth against the proxy for neighbourhood, the variable
265 describes an S-shaped curve which fans out considerably as mobility increases, reaching a plateau
266 at approximately 20 grid cells from the origin (see Figure 2b). A plausible mechanism behind this
267 pattern is that high-mobility isodes are both: a) unlikely to suffer stress events and thus maintain
268 starting populations (due to a comparatively large logistical radius for establishing new plots), and
269 b) more likely to move further away from their starting position and disproportionately target old
270 growth primary forest over fallowed secondary forest for new swidden plots.

271

272 The highest proportion of incipient regrowth in the study area (up to 1% of all cells) is caused by
273 isodes that maintain a neighbourhood of 8-15 grid cells (Figure 2c). This can be suggested to be
274 the tipping point at which isodes can move a sufficient distance from their previous house garden
275 to avoid stress events (thus maintaining population levels) while creating new swidden plots close
276 enough to old plots to adversely affect their odds of regeneration to secondary forest. This results
277 in isodes creating large tracts of scrubland that have difficulty succeeding beyond this stage of
278 regrowth, due to a lack of nearby old growth forest from which to draw pioneer species. The
279 prevalence of incipient regrowth does not appear to impact on the relative proportion of secondary
280 forest, however.

281

282 [Insert Figure 2 here]

283

284 The average time grid cells spend as scrubland (a cumulative metric over the duration each run) is
285 also correlated with mobility magnitude. The peak in the latter between values of 2-4 (Figure 2d)
286 is likely due to the same mechanism identified above: short average distances between successive
287 house gardens. Due to the proximity of a given isode to its point of origin, and hence fallowed
288 plots, cleared grid cells are more likely to border grid cells from previous cropping cycle. In one
289 exceptional run, the *mean* time grid cells spent as scrubland was 200 time steps, or a quarter of the
290 total length of the run. Runs with a logistical range of eight or higher display a sharp contraction
291 in the average time spent as scrubland. This is because an increase in the logistical radius of an
292 isode causes an exponential growth in the quantity of grid cells available for cultivation, reducing
293 the likelihood that fallowed swidden plots will affect the regeneration potential of new plots. This
294 is not true when mobility is lowered, which the results imply leads to outcomes with numerous
295 small yet persistent grid cells of scrubland.

296

297 Spatial visualization of some examples lends additional support to these trends, and provides some
298 useful contrasts between the influence of mobility magnitude on forest composition (Figure 3a).
299 Larger distances between new settlements results in more spatially diffuse tracts of anthropogenic
300 forest, although the absolute quantity of secondary forest is greater than in low mobility regimes
301 (see Figure 2b). Simultaneously, the amount of incipient regrowth is minimized by the distances
302 between swidden plots. **Large swidden logistical radii** leads to the appearance of tracts of modified
303 forest following long-term abandonment (>200 time steps), and the formation of diverse mosaics
304 of active swidden plots at different stages, secondary regrowth, and reverted primary forest (see
305 Figure 3a, inset). **Smaller radii (low mobility)** epitomizes the opposite, displaying spatially
306 circumscribed zones of secondary regrowth that frequently centre on groups of grid cells with
307 scrub vegetation (Figure 3b). These were observed to form through the mechanism hypothesized
308 from Figure 2c and Figure 2d, where isodes are unable to move far from old grounds, prolonging

309 the regeneration of incipient vegetation further. The temporality of anthropogenic land cover
310 change is therefore an important factor to consider.

311

312 [Insert Figure 3 here]

313

314 Examining model output diachronically permits three broad trends for incipient vegetation
315 formation to be defined (Figure 4). The first (A) indicates that low mobility causes an early and
316 rapid spike in the quantity of incipient forest plots, irrespective of regeneration potential. It was
317 hypothesized that this would cause frequent stress events among isodes, leading to a reduction in
318 total population and hence anthropogenic impact. This is not, however, borne out by the
319 experimental results; the legacy effects of early large-scale clearances persist until the end of runs.
320 Regeneration probability appears to exacerbate the magnitude of this outcome. The second trend
321 (B) displays a slower rise in incipient forest cover followed by plateaus or continued upward slopes
322 towards the end of runs. Finally, and in the majority, are the set of parameter combinations which
323 result in the creation of few, if any, persistent grid cells of incipient regrowth (C). The mean trend
324 is essentially flat. It is worth noting that, except for a single parameter combination (mobility = 6,
325 probability = 0.025), mobility is the key determinant of incipient vegetation trends, while
326 probability of regeneration only affects the magnitude of the trend.

327

328 [Insert Figure 4 here]

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330 An additional experiment was conducted over a limited number of parameter combinations to
331 investigate changes in the returns of active grid cells over time (Figure 5). The model was allowed
332 to run for three times the number of time steps as in the previous simulations (2400) to detect any
333 long-term diachronic trends in its output that shorter runs might not capture. The most significant
334 qualitative differences in the shapes of these curves is again in low mobility regimes. Except for

335 where population starts to collapse within a few hundred time steps, the mean returns from active
336 fields (a) become stationary almost immediately. Conversely, the returns from secondary forest (b)
337 fluctuate significantly in the first 1000 time steps and only flatten out after approximately 1500
338 time steps, far beyond the duration investigated above. Kolmogorov-Smirnov tests indicate
339 significant differences between all four curves in both time series at $p < 0.001$, in addition to the
340 qualitative variability between parameter combinations. Extrapolating from Figure 5b, the ratio of
341 secondary forest to primary forest appears to reach a dynamic equilibrium, where on average only
342 approximately 2000 grid cells of secondary forest exist any point in time, or $<2\%$ of the modelling
343 domain. Secondary forest which has succeeded to climax vegetation (reverted forest) can be
344 predicted to continue to expand over time (see Figure 2a). Finally, while active plots and secondary
345 fields may appear comparable in terms of the sums of their returns, it is worth noting that swidden
346 plots in this model are by definition located within the logistical radius of isodes. Secondary forest,
347 on the other hand, can be widely distributed in the landscape of the upper Cuao (see Figure 3a)
348 and are not necessarily exploited synchronically.

349

350 [Insert Figure 5 here]

351

352 5. Discussion

353

354 The aim of this experiment was to constrain the range of plausible scenarios for the scale of
355 impacts of Piaroa land use. Returning to this, assertions that the upper Cuao is a pristine
356 environment largely unaffected by significant land cover/land use changes over its history of
357 human inhabitation (Harris, 1971; Perera, 1986; Zent, 1992; Huber and Zent, 1995; Zent, 1995;
358 Freire, 2007) must be questioned in light of the modelling results. **Instead, the modelling results**
359 **provide quantitative backing to the idea that groups of small-scale agriculturalists can cause**
360 **significant qualitative changes to an environment.** In the context of Piaroa land use patterns, the

361 notion that shifting cultivation is necessarily a “high mobility, low impact” adaptation ought to be
362 modified. The results imply that the quantity of old growth forest affected by human action in the
363 interval since Contact may be up to four times greater than previously assumed, given the
364 parameter space explored in this case. Counterintuitively, high mobility *exacerbates* this pattern
365 rather than diminishes it. The deleterious effects of shifting cultivation (deforestation, transition
366 to savannah, and attendant losses of resilience and biodiversity) emerge only under the most
367 extreme combinations of low mobility and environmental vulnerability. Under conditions where
368 forest succession is impacted over the long term, this manifests in a spatially-circumscribed
369 manner, centred on individual Piaroa isodes and their ranges (see Figure 3b).

370

371 Shifting swidden cultivation has been compared unfavourably with “the cumulative impact of
372 passive, small- to medium-scale and localized disturbance”, meaning agroforestry (Stahl, 2015:
373 1600). While slash and burn farming is arguably an active form of landscape intervention, its impact
374 is demonstrably cumulative while almost by definition local in scale. In this respect, it is very similar
375 to the various types of terra firme agroforestry documented elsewhere in Amazonia that the above
376 quote refers to. Its persistent reputation as the *bête noire* of Neotropical subsistence, viewed as
377 limiting to cultural development, wasteful, and ultimately unsustainable is only just being shed
378 (Arroyo-Kalin, 2012; papers in Cairns, 2015). The depth of indigenous knowledge on managing
379 swidden plot nutrient cycles, including biodiverse secondary forest, provides useful templates for
380 exploring this topic computationally (Harris, 1971; Denevan, 1992b). It is clear that the intensive
381 and sedentary pre-Columbian agricultural systems on the margins of rivers, such as those
382 responsible for the creation of anthropogenic dark earths in the central Amazon basin between
383 2500 and 500 BP, were far from being the only response to subsistence requirements. Indigenous
384 people developed and deployed a variety of different systems depending on the setting and
385 historical circumstances they faced (Arroyo-Kalin, 2010; Neves, 2013; McMichael et al., 2014),
386 each of which left imprints on the landscape. Against this backdrop, the Piaroa modelling

387 experiment must be tempered with palaeoenvironmental data implying low-impact and temporary
388 human occupation in the Amazonian interfluves since the initial colonization of the basin (see
389 McMichael et al., 2012; Levis et al., 2012; Watling et al., 2015; Stahl, 2015).

390

391 Although subordinate to manioc production, the Piaroa obtain a great variety of non-domesticated
392 resources from tracts of secondary forest (Zent, 1992; Overing and Kaplan, 2011). In the context
393 of this experiment, it is the emergence of diffuse yet extensive secondary forest tracts which
394 promote high logistical mobility as a comparatively sustainable pattern of land use. Conversely,
395 low mobility regime runs evidence how the odds of isode survival and the rate of forest succession
396 are adversely affected by the expansion of large scrubland tracts. The fact that the upper Cuao
397 today remains dominated by closed canopy forest suggests that pre-twentieth century land use
398 patterns tended towards the equivalent of “high mobility” runs. Due to the diffuse nature of Piaroa
399 swidden farming, extant tracts of secondary forest (and succeeding climax communities) may
400 become nearly undetectable to the untrained eye. Heckenberger et al., (2007: 197) note that satellite
401 imagery shows Amazonia to be largely flat and green, yet a look under the canopy reveals high
402 biocultural diversity. Johnson et al. (2005: 104) observe:

403

404 “[A]rchaeologists are not very good at estimating in their heads the effects over long periods of time
405 of slow processes [...], particularly when those have to be weighed against a countervailing force such as forest
406 regrowth. It is important to address such processes in a quantitative framework to allow them to be
407 understood.”

408

409 Taken together, the above observations by Heckenberger et al., (2007) and Johnson et al. (2005)
410 highlight a tension between “deep time” on one hand, and modern observational data on tropical
411 forests and their inhabitants on the other (e.g. Politis, 2007; Rival, 2016). While prior accounts of
412 tropical swidden farming societies (e.g. Conklin, 1961; Harris, 1971; Zent, 1992) are not incorrect

413 or guilty of errors of omission, the shallow time depth of ethnographic fieldwork relative to long-
414 term ecological processes may be a contributing factor to this apparent discrepancy. Historical
415 ecology aims to understand landscapes as historically contingent and emergent artefacts of human
416 activity, the products of centuries (if not millennia) of small-scale actions enacted across the entire
417 biosphere. Consequently, understanding how short-term mechanisms create enduring anthropic
418 patterns of landscape heterogeneity is high on the agenda (Balée 2006). Due to the nature of the
419 archaeological and palaeoenvironmental records, however, the focus tends to be on the *longue durée*
420 of human-environmental history through the aggregation of numerous unresolvable individual
421 actions (Isendahl, 2016).

422

423 Computational modelling grounded in ethnographic data to represent such actions has
424 demonstrated one way in which this gap may be bridged, using simple generative rules to explore
425 a series of “what if” scenarios (Lake, 2014; Cegielski and Rogers, 2016; Davies, 2016; Magliocca
426 and Ellis, 2016). Building on the premise that the Piaroa operated a system of swidden farming in
427 the upper Cuao for approximately four centuries lends further weight to the idea legacy effects in
428 tropical forests are more profound and pervasive than previously hypothesized. Going further, can
429 Piaroa secondary forest tracts be identified as post-Contact equivalents of pre-Columbian
430 “anthropic islands in a sea of forest” (Neves, 2013: 379)? If that is the case, the results suggest that
431 they are not only extensive but follow people around the landscape too. When viewed together
432 with the ephemerality and frequent movement of isodes in this environment, an archaeology of
433 the upper Cuao (like other interfluvial zones of the Neotropics) becomes a challenging prospect.
434 Extant ecological indicators of disturbance and yet-to-be-collected palaeoecological records, with
435 high-resolution radiometric age determinations, may be the only recourse to reconstructing the
436 pathways taken by the post-Contact Piaroa and their predecessors.

437

438 The present study area centres on only one of the secondary rivers occupied by Piaroa groups (and
439 other indigenous people) in the upper Orinoco valley (see Freire and Zent, 2007; Overing and
440 Kaplan, 2011), which was selected as the modelling domain specifically for its relative isolation
441 (see Zent, 1992). It is important to underline that it is not reasonable to expect these results to be
442 mirrored exactly in other parts of the Neotropics where swidden agriculture is practiced; the model
443 design was deliberately particularistic (cf. Barton 2013). Significant factors documented elsewhere,
444 such as the re-use of pre-Columbian landscape alterations (Heckenberger et al., 2007) do not play
445 a major role here. On a general level, however, the results do caution further against assuming a
446 that the baseline model for Neotropical forests should necessarily be a natural one, and here it is
447 relevant to note that a consensus on the definition of a “natural” Amazonian environment is itself
448 elusive (McKey et al., 2010: 7827; Clement et al., 2015; Watling et al., 2017). Placing this in a pan-
449 Amazonian context, understanding indigenous environmental impacts and their legacy may be a
450 question of identifying and translating between patterns observable at different spatio-temporal
451 scales (Arroyo-Kalin, 2010; Balée, 2013; Boivin et al., 2016; Levis et al. 2017). A key element of
452 this will be correctly identifying the longevity, magnitude, and cumulative effects of pre- versus
453 post-Contact indigenous impacts in different locations (McMichael et al. 2017).

454

455 This study of the post-Contact Piaroa shows that simulation may prove a valuable tool in achieving
456 this goal. Returning to the introduction, a fruitful path forward would be to develop comparative
457 frameworks for the degree and types of coupled human-environmental systems rather than re-
458 tread old debates contrasting spatially extensive, low impact patterns with those which are spatially
459 intensive and locally high impact. Without overstating the extendibility of this specific experiment
460 beyond the upper Orinoco, simulation and computational methods can help bridge ethnohistory,
461 archaeology, palaeoecology, and the subsistence regimes inferred or hypothesized to exist from
462 these sources of information on the past.

463

464 6. Conclusions

465

466 The kinds of data that historical ecologists, anthropologists, and archaeologists are accustomed to
467 handling demonstrably benefit from ever-closer integration. Advances in this area will be especially
468 crucial to properly situating Amazonia in relation to human-induced changes in the biosphere
469 observed globally (Kirch, 2005; Erlandson and Braje, 2013; Kawa, 2016). Computational modelling
470 was adopted to characterize and develop a study of Piaroa ecodynamics, and proposed how
471 discrepancies between the model and the reported environmental conditions of the upper Cuao
472 may be resolved. By taking ethnographies at face value, the agent-based simulations suggest that
473 the environments inhabited by these groups cannot be considered pristine by any metric, even in
474 the relatively short time between European Contact to present. Furthermore, persistent concerns
475 that conceptual models of subsistence in Amazonian risk overemphasising some parts of a system
476 to the detriment our understanding of the whole (Whitehead, 1998: 38) is not a concern of formal,
477 explicit models. A deliberate epistemic focus on “imagined concrete” properties (Godfrey-Smith,
478 2006: 738) of a real-world system, rather than precise isomorphism, enabled a robust consideration
479 of the long-term effects of Piaroa swidden farming to be built from the ground up using aspects
480 of known behaviours as a point of departure. The results suggest that the anthropic impacts of
481 swidden farming in the modelling domain can manifest on very short timescales. This has
482 implications for future investigations of the effect of millennia of agroforestry and intensification
483 across the Neotropical world. As noted in the introduction, the ties between model-based science
484 and middle-range theory (Premo, 2007; Davies, 2016) are worth exploiting for historical ecologists.
485 Conceptual models, as semantic entities open to interpretation, will always have barriers between
486 the intent of their original authors and their audiences. Although computational models are not a
487 perfect replacement, they are also not vulnerable to this critique. With effort and informed theory-
488 building anyone can extend, modify, or contest the model discussed here. Indeed, this is welcomed,

489 and may be instrumental to framing expectations for future archaeological research into tropical
490 forest cultures as agents of biodiversity and change.

491

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501

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